Table 4-20—Average surfactant (ABS) concentration as a function of depth at the Rio Hondo Test Basin

Date	Average concentration in milligrams per liter at					
	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
anuary 1963	1.30	0.60	0.50	0.50	0.60	
ebruary	1.38	0.77	0.65	0.38	0.35	
farch	0.88	0.32	0.20	0.15	0.12	
pril	0.59	0.16	0.08	0.08	0.08	
ay	1.35	0.20	0.08	0.10	0.10	
une	1.46	0.12	0.00	0.00		
ily	2.02	0.23	0.10	0.07		
ugust	2.12	0.28	0.10	0.08		
ptember	1.26	0.13	0.10	0.10	0.10	
ctober	0.37	0.10	0.10	0.10	-	
ovember	0.46	0.10	0.00	0.03	0.07	
ecember	0.28	0.10	0.10	0.10	0.10	
nuary 1964	0.37	0.12	0.12	0.10	0.10	
bruary	0.28	0.20	0.05	0.10	0.10	
arch	0.55	0.35	0.05	0.20	0.10	
oril						
ay	1.90	1.50	0.40	0.10		
ne	0.40	0.20	0.10		3	
ly	0.30	0.10	0.10	0.10		
gust	2.00	1.00	0.30	0.10	a	
ptember	1.90	1.10	0.45	0.15		
tober	1.70	0.80	0.30	0.10	a	
vember	0.40	0.30	0.10	0.10		
cember	0.40	0.25	0.05	0.05		
nuary 1965	0.30	0.10	0.05	0.05		
bruary	0.40	0.15	0.10	0.10		
arch	0.28	0.10	0.10	0.10	a	

[·] No samples available for analysis.

Table 4-21—Average chloride concentration as a function of depth at the Rio Hondo Test Basin

Date	Average concentration in milligrams per liter at					
	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963	63	62	58	92	80	
February	54	52	51	55	84	
March	84	70	79	78	76	
April	64	73	71	88	72	
May	90	100	93	94	86	
June	94	103	98	110	00	
July	100	107	114	121		
August	119	122	126	128		
September	62	72	88	134	94	
October	51	39	31	68	3 1	
November	43	59	94	87	73	
December	86	89	88	73	86	
anuary 1964	69	82	80	88	73	
ebruary	87	88	88	81	10	
farch	68	50	87	101	86	
April	. a			101	00	
May	111	106	114	111		

a No samples available for analysis.

Table 4-22—Average nitrate concentration as a function of depth at the Rio Hondo Test Basin

	Average concentration in milligrams N per liter at					
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963	1.4 1.4 2.3 5.0 5.2 4.0 1.8 0.6 0.9 1.1 1.1 1.1 2.9 2.0 5.2 2.0 2.0 5.2	1.8 2.7 3.9 2.2 2.0 >10.0 >10.0 >6.5 2.5 3.7 2.1 1.7 6.0 6.0 3.1 0.3 1.7 6.1	1.1 2.7 4.4 2.0 0.2 6.3 7.9 6.5 2.9 3.1 2.0 1.4 1.2 1.5 5 4.7 0.4 2.9	1.1 >10.6 4.5 2.5 2.5 2.2 1.5 5.9 7.6 8.7 3.8 3.8 3.8 3.8 3.4 3.8 6.2 6.2 6.8 5.5	2.14.53 3.35.55 1.22 0.00 0.99 1.22 3.1.75 b b b b b b b b b b b b b b b b b b b	
October	0.4 2.0 0.7 0.7 0.8 1.1	0.6 1.0 1.3 3.2 5.8 1.6	0.7 1.3 1.2 3.3 b	3.6 4.4 2.2 3.6 3.4 3.8	b b b	

No samples taken for analysis.
 No samples available for analysis.

Table 4-23—Average nitrite concentration as a function of depth at the Rio Hondo Test Basin

	Average concentration in milligrams N per liter at					
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963	3	0.17	0.02	0.00	0.08	
February		0.01	0.00	0.02	0.03	
farch	3	0.00	0.00	0.00	0.03	
April	0.54	0.00	0.00	0.01	0.01	
fay	0.84	0.41	0.18	0.08	0.01	
une	0.56	0.04	0.00	0.02	1	
uly	>1.98	0.28	0.02	>0.42	E	
ugust	2.13	0.35	0.08	0.20	b	
eptember	1.19	0.03	0.01	0.03	b	
ctober	0.07	0.01	0.01	0.04	ъ	
ovember	0.08	0.01	0.06	0.02	0.00	
December	0.32	0.00	0.00	0.01	0.06	
anuary 1964	0.18	0.09	0.03	0.00	0.00	
ebruary	0.29	0.00	0.00	0.00	0.00	
farch	0.06	0.03	0.00	0.00	0.00	
pril	ь	b	ь	ь	b	
fay	1.10	0.87	0.58	0.00	ь	
une	1.74	1.65	0.01	0.00	ь	
uly	0.00	0.00	0.01	0.00	b	
ugust	0.67	0.05	0.01	0.01	b	
eptember	1.04	0.02	0.03	0.02	b	
ctober	0.01	0.01	0.01	0.01	ь	
ovember	0.10	0.02	0.01	0.01	b	
ecember	0.20	0.02	0.04	0.01	b	
anuary 1965	0.17	0.18	0.01	0.01	b	
ebruary	0.14	0.04	ь	0.01	Ь	
farch	0.07	0.01	0.01	0.01	b	

No samples taken for analysis.
 No samples available for analysis.

Table 4-24—Average organic nitrogen and ammonia as a function of depth at the Rio Hondo Test Basin

Date	Average concentration in milligrams N per liter at					
	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963 February March April May June July August September October November December January 1964 February March Angil May June July Angust September October November December January 1964 February March April May June July August September Decober Decober January 1965 February March	7.66 >11.2 8.1 9.00 2.5 4.9 3.3 2.7 2.3 3.00 1.1 1.0 1.4 2.1 3.3 1.4 2.1 3.3	4.6 1.3 1.0 0.9 1.2 2.0 4.1 1.9 2.8 1.8 1.0 0.9 8.2 0.4 0.4 0.4 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	1.7 1.6 1.1 1.2 0.8 1.0 2.1 4.0 2.9 2.3 0.4 0.3 1.0 0.6 0.4 0.7 1.8 0.6 0.6 0.6	1.2 1.5 1.3 0.8 0.8 0.4 1.0 2.9 4.7 2.4 2.1 0.3 0.4 0.7 0.4 0.6 1.1 0.4 0.6 1.1 0.4 0.6	1.2 1.3 1.4 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8 0.8	

No samples taken for analysis.
 No samples available for analysis.

Table 4-25—Average total nitrogen as a function of depth at the Rio Hondo Test Basin

	Average concentration in milligrams N per liter at					
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963	a	6.6	2.8	2.3	3.4	
February	a	3.0	4.3	>12.1	5.8	
March	n.	4.9	5.5	5.8	4.7	
April	9.3	3.3	3.1	3.3	3.3	
May	>13.4	8.9	3.4	3.1	2.0	
une	11.0	>11.8	1.0	2.0	1	
fuly	>16.0	>11.4	7.3	>7.3	1	
August	9.8	>10.2	10.1	10.7	į.	
September	10.1	>10.6	10.5	13.4	t	
October	5.2	4.4	6.8	6.3	l l	
ovember	7.0	6.5	6.1	6.6	1.9	
December	3.6	3.9	4.3	5.1	2.7	
anuary 1964	3.3	2.8	2.2	2.6	2.1	
ebruary	3.3 2.3	6.9	1.6	3.7	6.4	
March	2.3 b	6.9	1.8	4.2	2.2	
prilday	15.0	12.2	6.3	6.9	1	
une	5.0	2.4	0.8	0.9 b	, t	
uly	3.4	2.0	3.5	3.6	b	
ugust	8.9	7.2	5.6	7.2	b	
eptember	5.0	5.0	4.4	6.1	b	
ctober	2.8	3.8	3.1	4.7	b	
ovember	4.2	1.4	1.4	4.8	b	
ecember	4.2	2.1	1.8	2.8	b	
anuary 1965	2.3	3.9	3.9	4.1	b	
ebruary	3.9	7.0	b	4.4	b	
March	2.6	2.2	4.1	4.2	b	

No samples taken for analysis.
 No samples available for analysis.

Table 4-26—Average turbidity as a function of depth at the Rio Hondo Test Basin

	Average concentration in units of turbidity a					
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
January 1963 February March April May June July August September October November December January 1964 February March	4.8 2.8 2.7 1.3	0.0 0.8 7.2 1.2 0.7 0.5 0.7 0.9 1.4 1.8 1.4	0.6 0.6 0.8 0.7 0.9 1.2 1.6 2.3 1.4 1.8	0.6 0.1 3.6 1.1 0.6 0.7 0.8 1.0 1.1 1.3 1.7 1.6 1.4	1.6 1.2 0.8 1.2 0.6 2.1 1.3 1.7 1.0 7.1	
April May June July	2.8 8.3 2.4	2.2 2.4 1.6	1.7 2.1 1.0	1.7 a 2.1		
August	2.1	1.3	1.0	1.0		

No samples taken for analysis.
 No samples available for analysis.

Table 4-27—Summary of average chemical data at the Whittier Narrows Test Basin

Period I. January-April 1963

	Average concentration in milligrams per liter						
	Surface	2 ft.	4 ft.	6 ft.	8 ft.		
Total solids Fixed solids Volatile solids Chemical oxygen demand Alkylbenzenesulfonate Nitrate as N Total nitrogen as N	688 442 246 47.6 3.23 0.2 21.3	959 587 372 41.1 2.38 0.3 >9.4	1,059 654 405 44.0 1.72 >6.8 >8.7	1,312 812 500 52.2 0.84 >16.3 >18.4	2,206 1,634 572 30.3 0.76 >13.4 >14.2		

Period II. May 1963-January 1964

Total solids. Fixed solids. Volatile solids. Chemical oxygen demand. Alkylbenzenesulfonate. Nitrate as N. Total nitrogen as N.	667 462 205 33.7 1.96 6.2 >22.2	937 625 312 24.5 1.08 7.6 >10.3	974 620 354 21.4 0.59 >15.7 >17.6	965 609 356 79.4 0.57 12.2 >14.2	1,000 618 382 52.9 0.45
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Period III. February-June 1964

Total solids. Fixed solids. Volatile solids. Chemical oxygen demand. Alkylbenzenesulfonate. Nitrate as N. Total nitrogen as N.	812	1,057	1,021	1,089	1,174
	532	618	623	652	670
	280	439	398	437	504
	36.7	8.8	9.1	12.6	38.8
	2.07	0.35	0.24	0.18	0.16
	4.7	22.7	24.6	22.7	24.9
	25.3	23.7	25.3	23.5	25.8

Period IV. July 1964-March 1965

Total solids. Fixed solids. Volatile solids Chemical oxygen demand Alkylbenzenesulfonate Nitrate as N Total nitrogen as N	798	1,049	1,025	1,035	1,183
	526	626	602	627	680
	272	423	423	408	503
	39.3	10.4	9.7	17.0	14.6
	2.54	0.34	0.33	0.35	0.29
	4.7	22.5	24.6	20.0	28.7
	27.0	25.6	25.4	21.1	29.8

Entire Test. January 1963-March 1965

Total solids. Fixed solids. Volatile solids Chemical oxygen demand. Alkylbenzenesulfonate. Nitrate as N. Total nitrogen as N.	740	1,000	1,010	1,065	1,335
	494	619	620	655	840
	246	381	390	410	495
	37.8	19.4	17.8	40.8	29.9
	2.36	0.89	0.60	0.46	0.28
	5.0	14.4	>19.0	>17.6	>24.2
	>21.0	>17.8	>20.3	>19.0	24.8

Table 4-28—Summary of average chemical data at the Rio Hondo Test Basin

Period I. January-April 1963

	Average concentration in milligrams per liter at								
	Surface	2 ft.	4 ft.	6 ft.	8 ft.,				
Total solids. Fixed solids. Volatile solids. Chemical oxygen demand. Alkylbenzenesulfonate. Nitrate as N. Total nitrogen as N.	784 490 294 27.4 1.04 1.2 9.3	713 419 294 25.4 0.46 2.7 4.5	739 471 268 18.4 0.36 2.6 3.9	769 542 227 15.8 0.28 >4.7 5.9	785 530 255 17.0 0.29 3.1 4.3				

Period II. May 1963-January 1964

Total solids Fixed solids Volatile solids	660 455 205	665 460 205	737 512 225	850 571 279	727 494 233
Chemical oxygen demand Alkylbenzenesulfonate	34.5 1.08	13.5 0.15	18.1 0.08	12.4	45.5
Nitrate as N Total nitrogen as N	>8.8	>5.8 >7.8	3.6 5.7	>6.3	0.8

Period III. February-December 1964

Total solids. Fixed solids. Volatile solids. Chemical oxygen demand. Alkylbenzenesulfonate. Nitrate as N. Total nitrogen as N.	756	1,018	825	925	878
	485	756	560	610	608
	271	262	265	315	270
	35.0	16.0	8.9	7.0	7.4
	0.98	0.58	0.19	0.11	0.10
	1.8	2.9	2.3	4.3	2.4
	5.4	5.0	3.0	4.9	4.3

Period IV. January-March 1965

Total solids	869	884	974	977	
Fixed solids	641 228	670	666	690 287	
Volatile solids	12.5	214 7.5	308 6.6	7.6	
Chemical oxygen demand Alkylbenzenesulfonate	0.33	0.12	0.08	0.08	
Nitrate as N.	0.9	3.5	3.5	3.6	
Total nitrogen as N	2.9	4.4	4.0	4.2	

Entire Test. January 1963-March 1965

Total solids. Fixed solids. Volatile solids. Chemical oxygen demand Alkylbenzenesulfonate Nitrate as N	740 494 246 31.4 0.99	835 592 243 15.6 0.36 >3.9	800 543 257 13.3 0.16 2.9	880 596 284 10.7 0.12 >4.3	775 529 246 8.8 0.17 2.0
Total nitrogen as N	>6.6	>5.8	4.2	>5.5	3.4

Table 4-29—Water temperature as a function of depth at the Whittier Narrows Test Basin

						Percol	ate at			
	Surfac	e water	2 ft.		4	ft.	6	ft.	8 ft.	
Date	Number of observations	Range of temperature,	Number of observations	Range of temperature, °F	Number of observations	Range of temperature, °F	Number of observations	Range of temperature, °F	Number of observations	Range of temperatu
January 1963 February March April May June July Jugust Eptember Jetober Sovember January 1964 February Jarch Jugust	21 18 15 20 13 5 9 10 5 8 10 11 11 11 9 13 6 4 12 19 22 20 18 19 19 19 19 19 19 19 19 19 19	56-68 62-72 63-70 64-77 63-78 69-78 74-80 76-82 77-81 72-79 70-75 66-70 68-72 68-73 75-79 79-84 81-82 79-81 75-89 71-80 66-69 65-69 65-69	03 41 63 63 44 53 63 63 44 44 44 44 44 45 45 63 53 44 44 63 63 44 44 63	54-62 59-61 59-60 63-69 65-66 68-68 70-73 70-72 70-75 61-68 54-59 52-55 54-61 54-63 64-68 63-72 72-74 76-79 85-64 55-63 55-63 55-63 55-63	24 11 11 44 53 63 00 55 41 41 41 41 41 41 41 45 45 53 41 45 53 41 41 41 53	54-56 59-61 60 58 63-68 65-66 65-66 68-72 70-75 61-70 54-57 54-61 59-68 64-72 64-72 72-75 77-79 82 73-77 60-70 59-65 59-65	33124100034444444533445334443	53-56 58-61 57 62-70 65 68-72 70-73 61-70 54-57 54-61 59-64 64-72 66-73 73-76 78-82 81-82 73-79 62-68 61-67 68-64 68-73 64-73 64-73 73-79 62-68 61-67 63-64 63-64 64-64 65-64 68-64	21100000000000000000000000000000000000	56-58 59 59 59 54-61 59-6-65 72-77-77 79-81 73-77 74-77-74 74-77-74 58-6-58 58-6-58 59-6-58

Table 4-30—Water temperature as a function of depth at the Rio Hondo Test Basin

						Percol	ate at			
	Surface	e water	2	ft.	4	ft.	6	ft.	8	ft.
Date	Number of observations	Range of temperature, °F	Number of observations	Range of temperature,	Number of observations	Range of temperature, °F	Number of observations	Range of temperature,	Number of observations	Range of temperature
anuary 1963 ebruary farch pril farch pril fay une uly uugust evetember eccember anuary 1964 ebruary darch pril day une duy tober anuary 1964 ebruary darch pril day une duy tugust september ebruary farch pril day une duy tugust september Detober November Detober Anuary 1965 ebruary february		54-60 58-69 58-76 57-75 63-77 65-78 74-84 77-86 65-75 62-70 52-64 55-68 64-76 73-75 68 60 57-61 67 73-75 68 60 57-61 60 67 73-75 68 60 67 73-75 68 60 67 73-75 68 60 67 73-75 68 69 60 60 60 60 60 60 60 60 60 60	13 45 45 35 11 3 3 4 2 2 0 2 1 2 1 1 2 2 1 1	53 55-59 54-62 59-64 63-68 65-73 70-74 65 59-65 54-59 50-55 54-58 66-68 69 77-80 68 69 77-80 68 69 59-61 51-58 56 64	1 3 3 5 5 3 1 2 2 5 1 0 0 0 3 4 4 2 1 0 2 1 2 1 1 2 1 0 0 1	53 56-58 53-59-64 63-66 64-67-72 69-74 74-59 49-54 54-55 57-69 74-79 68-72 69-74-79 68-72-69 74-79-76 70-72-68 62-57-60	133554533410013342220220211221122211	54 55-59 59-63 64-66 62-66 66-72 72-73 75 62 55-59 49-54 53-55 57-59 68-72 73-78 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 68-72 70-73 70-73 68-72 70-73 70-	1 1 3 4 2 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	53 56 53 59 63 62 66 62 66 62 66 62 66 62 66 62 66 62 66 62 66 62 66 62 63 62 63 62 63 62 63 62 63 62 63 63 63 63 63 63 63 63 63 63 63 63 63

Table 4-31—Dissolved oxygen as a function of depth at the Whittier Narrows Test Basin

						Perco	late at			
	Surfac	e water	2 ft.		4	ft.	6	ft.	8	ft.
Date	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/
anuary 1963 - ebruary - farch - pril - fay - une - uly - ugust - eptember - etober - eocember - anuary 1964 - beruary - farch - pril - fay - une - uly -	3 0 1 2 2 2 2 0 0 0	1.9-6.7 4.3-8.3 2.6 7.8-8.4 8.1-8.0 5.2-8.0 9.4 6.2-11.1	21 23 1-1 1-4 23 24 23 1-1 24 4 23 23 23 23 23 23 23 23 23 23 23 23 23	1.8-2.2 4.2-6.8 5.4 2.6 0.8-3.1 0 -0.4 0 -2.3 0 -2.5 0.2-3.3 0 -2.5 0.2-1.7 0.2-1.7 0.5-2.2 0.5-2.3 0.5-2.3	1 1 0 0 3 2 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	1.7 0.5 1.4-4.5 0 -2.8	3 2 1 1 1 2 1 1 1 1 1 3 1 3 4 2 4 4 3 33 4	1.8-3.3 0.3 2.4 1.9 1.1-1.4 0 2.8 1.3 2.1 1.9 0.2-2.9 0.3-3.6 1.9-6.8 0.3-6.3 1.9-5.5 0.3-6.3 1.9-7.4 0	4 2 3 0 1 1 1 0 0 0 0 0 0 0 0 2 1 1 1 1 1 1 1	1.0-3.2 0.6-0.8 1.1-1.5 5.1 0
ugust ptember ctober ovember ecember muary 1965 ebruary arch	4 5 4 2 4 4 3 3	8.5- 9.8 9.4-10.1 8.4-10.1 9.0-11.5 4.1-11.3 9.9-11.5 10.5-12.0 10.4-11.8	4 4 5 2 4 4 3 1	0 -0.8 0 -1.6 0 -1.8 0 -0.6 0.2-2.9 1.5-8.6 0.6-1.8	0 0 1 0 0 1	3.9 0.4 8.0	4 4 12 53 44 44 53 53	2.5-4.2 0.4-2.8 0 -7.1 0.2-1.5 0.7-5.0 0.1-0.5 0.1-2.2 0 -0.1	2452453	0 -0.5 0 -0.2 0 -2.0 0.1-0.4 0.2-0.5 0.1-0.2 0.1-0.3

Table 4-32—Dissolved oxygen as a function of depth at the Rio Hondo Test Basin

						Perco	late at			
	Surface water		2 ft.		4	ft.	6	ft.	8	ft.
Date	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/l	Number of observations	Range of D. O., mg/
fanuary 1963 Pebruary March April Asy une uily uugust Peptember Pectober Rovember Roeember Anuary 1964 Pebruary March May Uugust Legember Anuary May Legember Anuary May May May May May May May May May Ma	0 1 1 1 1 2 1 0 0 0 0 0 0 0 0 0 2 2 1 1 2 1 2	4.5 9.7 9.4 9.6 7.6-7.8 8.0 	12113442551103341120112112112110	5.0 - 5.8 10.0 4.5 - 7.7 1.0 - 3.6 1.6 - 2.6 1.7 - 2.5 0.5 - 1.8 0.5 - 1.8 1.9 - 5.9 5.0 - 8.2 9.0 - 10.2 4.1 - 10.5 0 - 3.5 0 - 3.5 0 - 3.5 1.9 - 3.5 1.0 - 3.5	1 2 0 0 2 1 1 1 5 1 0 0 0 3 3 3 1 2 2 0 1 1 2 1 2 1 1 3 1 0 0 1	7.2- 8.7 7.2- 8.2 5.0- 5.5 3.4 1.9 3.4 1.3- 3.6 4.0 6.7- 8.4 9.0 8.5-10.6 7.1 8.4 0 - 5.5 6.8- 7.4 7.5- 9.9 8.0 10.3	1 3 3 5 5 3 4 2 5 5 1 0 0 0 3 4 4 1 1 0 0 2 1 2 2 2 1	8.5 7.5-8.4 7.0-9.2 5.0-9.3 1.1-2.8 1.2-3.6 1.7-2.6 1.7-2.6 1.9-3.3 3.6 1.9-3.3 3.9-4.1 5.0-8.3 9.0 10.5 6.9 4.9-5.7 6.4-6.8 7.2 7.2 6.2-9.1 4.9-6.8 7.8-9.4 8.8	1 3 2 2 1 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	0.6- 9.1 8.8 8.8-10.8

Table 4-33—Water quality of percolate at Whittier Narrows Test Basin * (composite samples, 12-21 Aug. 1964)

	C	Concentration	n in milligra	ms per liter	at
Constituent	Surface	2 ft.	4 ft.	6 ft.	8 ft.
Calcium, Ca	60.8 19.9 152 14.5 40	132 20.9 120 13.0 0	127 19.4 142 15.4	139 17.9 140 12.6 0	158 30.1 138 5.1 0
Bicarbonate, HCO ₂ Sulfate, SO ₄ Chloride, Cl Nitrate, NO ₃ Phosphate, PO ₄	385 164 126 44.0 5.4	369 160 134 44.0 0.60	336 164 131 104 10.0	395 161 130 84.2 0.30	487 168 126 88.0 0.2
Total dissolved solids (sum)	1,011	994	1,050	1,080	1,200
Total hardness (as CaCO ₃)	234	411	398	422	520
оН	8.02	7.69	7.87	7.84	7.78
Electrical conductivity (micrombos at 25°C)	1,230	1,390	1,390	1,410	1,530
			V		

^{*} Analyses by Los Angeles County Flood Control District.

Table 4-34—Water quality of percolate at Rio Hondo Test Basin * (composite samples, 12–21 Aug. 1964)

	C	oncentration	in milligran	ns per liter a	t
Constituent	Surface	2 ft.	4 ft.	6 ft.	8 ft
Calcium, Ca	52.8 18.7 142 15.2 0	52.0 17.0 144 12.5 0	65.6 18.2 142 9.7 0	93.6 20.9 137 9.5	
Bicarbonate, HCO ₃ Sulfate, SO ₄ Chloride, Cl Nitrate, NO ₅ Phosphate, PO ₄	180 170 122 38.3 6.5	174 177 120 30.0 7.0	188 170 126 49.0 7.0	246 193 124 52.2 3.5	:
Total dissolved solids (sum)	746	734	776	880	
Total hardness (as CaCO3)	209	200	239	320	
HH	8.30	8.02	7.84	7.90	
Electrical conductivity (micromhos at 25°C)	1,080	1,060	1,140	1,230	,

^{*} Analyses by Los Angeles County Flood Control District. " No sample available for analysis.

Table 4-35—Phosphate (as PO4) as a function of depth at the test basins Whittier Narrows Test Basin

		Orth	o-phosphate (n	og/1)		Total phosphate (mg/l)				
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	Surface	2 ft.	4 ft.	6 ft.	8 ft.
1964 2-30/12-31	16.8	5.4	15.7	4.8	0.0	29.1	18.6	29.1	16.1	0.8
1965 1-6/1-7. 1-13/1-14 1-20/1-21 1-27/1-28.	16.9 15.4 15.6 21.5	5.5 6.3 7.3 6.4	17.6 18.1 17.7 18.8	6.5 14.3 16.6 15.7	$\begin{array}{c} 0.0 \\ 0.2 \\ 1.5 \\ 0.2 \end{array}$	18.8 17.8 16.2	6.4 7.0 6.5	17.8 19.9 17.0	7.0 16.1 14.3	0.9 0.8 0.0
2-3/2-4. 2-10/2-11. 2-17/2-18. 2-24/2-25.	23.8 18.4 23.1 24.0	6.8 7.6 7.6 8.2	14.5 16.3 15.9 21.1	11.3 13.6 17.2 17.9	$\begin{array}{c} 0.0 \\ 0.0 \\ 0.4 \\ 0.3 \end{array}$	26.4 19.1 34.4 22.4	6.8 9.2 7.9 7.4	14.4 14.9 17.6 22.9	11.3 14.9 17.7 17.2	0.2 0.2 0.6 0.2
3-3/3-4 3-10/3-11. 3-17/3-18. 3-24/3-25.	22.2 22.4 16.8 18.1	10.1 10.6 11.8 10.8	19.9 18.7 21.6 20.7	19.9 16.3 17.3 19.8	0.2 0.0 0.2 0.0	29.1 27.4 16.8 17.4	10.3 12.2 11.1 11.1	16.7 16.4 21.2 22.1	18.1 17.8 17.1 21.7	0.3 0.3 0.3
Totals	255.0 19.6 15.4-24.0	104.4 8.0 5.4-11.8	236.6 18.2 14.5-21.6	191.2 14.7 4.8–19.9	3.0 0.2 0.0-1.5	274.9 22.9 16.2-34.4	114.5 9.5 6.4–18.6	230.0 19.2 14.4-29.1	189.3 15.8 7.0-21.7	4.3 0.4 0.0-0

Rio Hondo Test Basin

		Orth	o-phosphate (r	ng/1)		Total phosphate (mg/1)					
Date	Surface	2 ft.	4 ft.	6 ft.	8 ft.	Surface	2 ft.	4 ft.	6 ft.	8 ft.	
1965 -12	1.8 2.1	3.4 3.6	3.2	2.2 2.4	::	2.4 2.6	4.0 3.7	4.0	2.7 2.6	:	
-8/2-9	3.5 4.8	4.9 2.3	3.5	2.3 4.4	:-	3.6 7.6	4.9 2.2	::	2.4 4.5	::	
-8. -22/3-23.	2.6 2.0	4.2 3.4	3.8 6.1	2.4 2.3		2.4 2.1	4.2 2.1	4.0 3.1	2.6 2.4	::	
Totals	16.8 2.8 1.8-4.8	21.8 3.6 2.3-4.9	16.6 4.2 3.2-6.1	16.0 2.7 2.2-4.4		20.7 3.4 2.1-7.6	21.1 3.5 2.1-4.9	11.1 3.7 3.1-4.0	17.2 2.9 2.4-4.0		

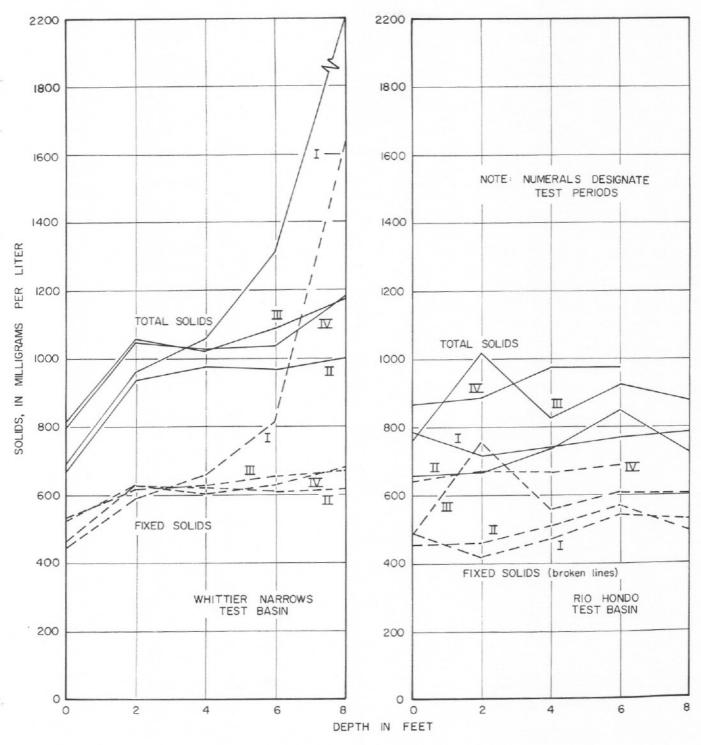


Fig. 4-11—Total Solids and Fixed Solids at Both Test Basins

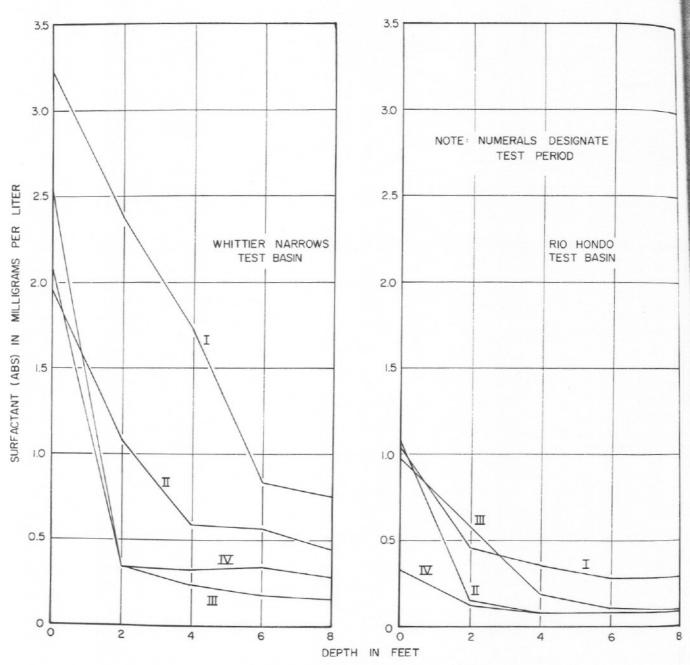


Fig. 4-12—Surfactant (ABS) at Both Test Basins

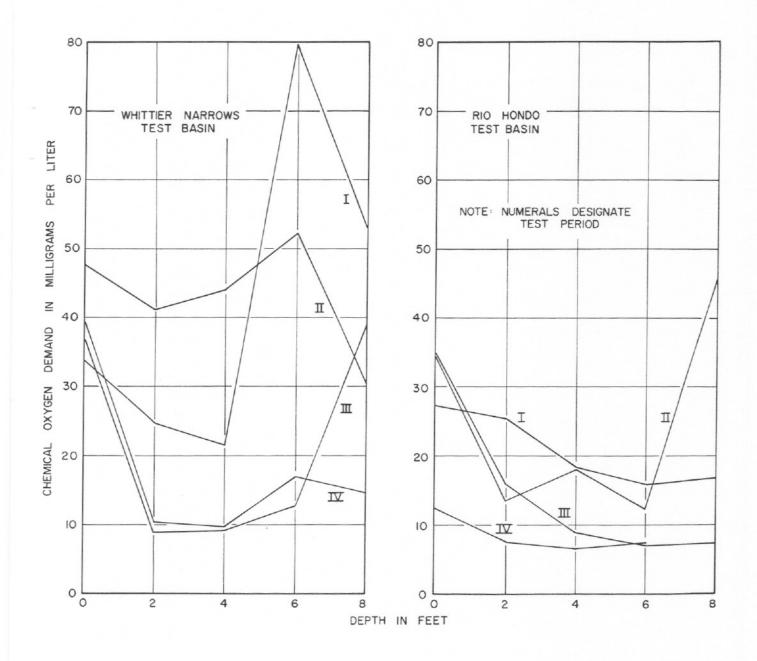


Fig. 4-13—Chemical Oxygen Demand (COD) at Both Test Basins

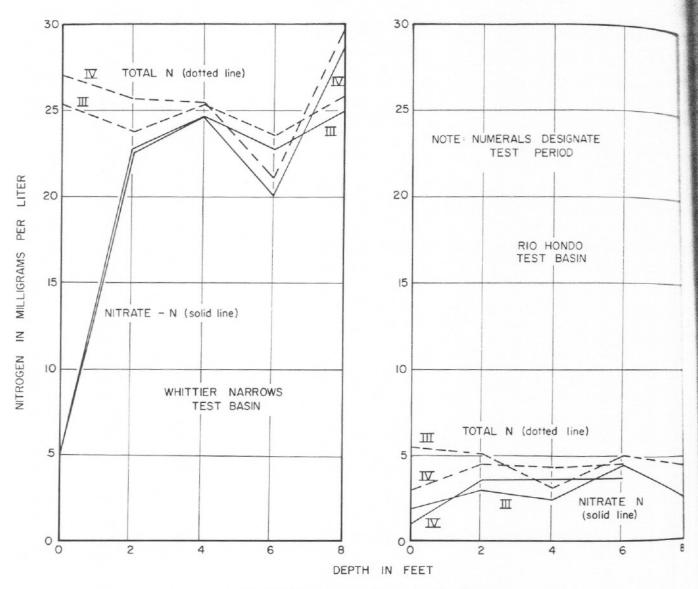


Fig. 4-14—Nitrates and Total Nitrogen at Both Test Basins

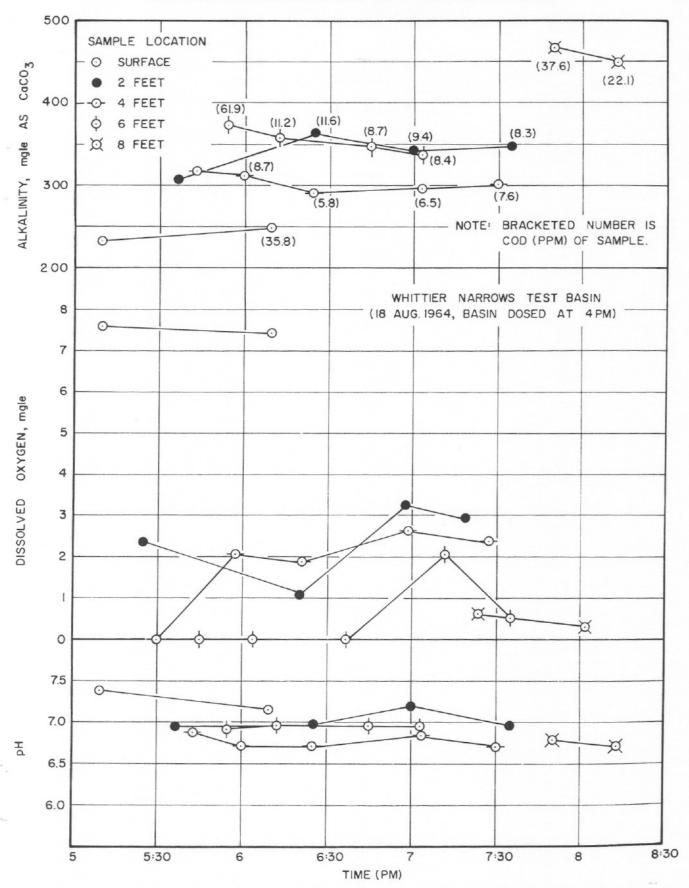


Fig. 4-15—Chemical Analyses of Instantaneous Percolates at Whittier Narrows Test Basin

CHAPTER 5

DATA FROM PROGRAM OF WELL SAMPLING IN THE MONTEBELLO FOREBAY

5.01 Location of the Wells.

In the spring of 1962, seventeen wells located in the Whittier Narrows area were selected by Caltech and the Los Angeles County Flood Control District (LACFCD) to form the network for a shallow-well sampling program. These wells are shown on the vicinity map designated Figure 4-1, Chapter 4. With funds supplied by this project, an additional well identified by the LACFCD as well No. 1573J was drilled at a site just south of the Rio Hondo Spreading Grounds. The wells were selected because they delineate the area directly influenced by the spreading of the effluent from the Whittier Narrows Water Reclamation Plant. Wells were also selected on the basis of good condition of the well hole, access to the well site, knowledge of the condition of the well casing and depth of perforations, and the well having a diameter large enough to accept the portable, selective-depth pumping unit of the LACFCD.

Table 5-1 is a list of the eighteen wells selected for routine sampling. These wells are classified as shallow wells because they penetrate only the zone of aquifers near the ground surface. These near-surface aquifers constitute the Gaspur zone. In the area of the Rio Hondo Spreading Grounds and to the north of the Montebello forebay, the Gaspur zone is fairly continuous between an elevation of about mean sea level and the ground surface which is approximately 200 feet above mean sea level. About one mile south of the Rio Hondo Spreading Grounds an impervious layer caps the Gaspur strata, changing this zone from an unconfined aquifer to a so-called pressure aquifer. The Gaspur zone is of interest because waters introduced into the ground-water basin from spreading grounds in the Montebello forebay first come in contact with ground water in the Gaspur zone.

Because spreading of reclaimed water was practiced on the San Gabriel River side of the Montebello forebay as well as in the Rio Hondo Spreading Grounds, some wells were selected in 1963 for sampling along the San Gabriel River. These wells are shown in Figure 4-1.

5.02 Method of Sampling.

Each of the shallow wells was sampled on a routine basis with a 3-month to 6-month interval between samplings. The collection of water samples was done by the personnel of the Los Angeles County Flood Control District, and the laboratory analyses were performed by Caltech personnel on the campus.

A special selective-depth pumping unit was fabricated by the Los Angeles County Flood Control District for sampling in the shallow wells. This special unit is portable and permits the collection of a water sample from a particular level in a well. Inflatable balloons or packers are positioned above and below

Table 5-1-Shallow wells chosen for multiple depth sampling

Well number*	Well diameter (inches)	Well reference point elevation† (feet above mean sea level)	Casing depth (feet)
2936A	10	216.5	75.9
2936 2937V	8 12	210.5 212.0	89.6 61.6
2917T	8.5	198.2	56
2928N	8.5	186.4	50
2928R	8.5	188.7	50
2918N	12	187.2	118.4
2909 N	12	177.0	105.4
1600L	12	169.9	129.6
1590F	10	168.8	71
1590D	12	167.4	123.9
1590M	12	158.1	89.6
1592A	12	156.9	162.8
1592X	8 8 12	150.0	201
1573J	8	141.5	141
1563H	12	136.5	243.5
1562	12	148.5	260
1561M	8	161.0	168.6

Number based on well grid identification system of the Los Angeles County Flood Control District,
 At or near ground surface elevation.

the pump intake. When inflated, the packers close off the well above and below the pump intake so that water can be pumped from a particular level in a well. This type of operation permits the taking of water samples at multiple depths within a well. Limitations are caused by the spacing of the perforations in a cased well and the location of zones of impervious material.

The Los Angeles County Flood Control District selective-depth pumping unit consists of a Reda pump with a 3-HP motor. It has an overall length of 56 inches and has a 3.77-inch diameter. The pump is 21 inches long and the pump screen is 21 inches long. The packers are made from 40 durometer natural rubber tubing which has a 5½ inch internal diameter and a 1 inch wall thickness. Each packer is about 12 inches long and is held to the pump body with stainless steel clamps. Where the rubber tubing is attached to the pump unit the body has been built up with fiberglass.

After placing the pumping unit at a selected depth, the packers are inflated with nitrogen gas from a portable cylinder to a pressure of 10 psi above the hydrostatic water pressure at that depth. This pressure differential provides a tight seal in the well which prevents the short-circuiting of water around the packers. The pump is operated for a short period of time, generally fifteen minutes, and then a one-gallon sample is taken. The pumping rate is generally 1 to 10 gallons per minute. Measurements of temperature, pH, and dissolved oxygen are made in the field by Flood Control District personnel. Samples are usually taken at 10-foot intervals in a well. However, the sampling pattern varies from well to well, depending on location of the casing perforations and depth of the well.

Table 5-2—Summary of water	analyses	on integrat	ed samples	for shallov	well netw	ork	
Sample No.	25	31	20	34	22	29	36
Well No. Date—1962. Temp., field/lab. (°C.) pH, lab. Dissolved oxygen (mg/l O ₂). Turbidity (units).	2936A 25 June 18/— 7.25 1.4 2.8 788	2936A 16 July 16/23 7.51 1.4 0.2 738	2936 21 June 20/— 7.58 1.0 0.5 347	2936 17 July 23/— 7.63 0.6	2937V 23 June 17/— 7.40 1.5 3.0	2937V 16 July 17.5/22 7.70 0.2	2917T 18 July 15/22 7.3 1.2 0.4 970
Total dissolved solids (mg/l TDS) Total volatile solids (mg/l TVS) Alkalinity (mg/l CaCO ₃)	625 282 258	627 220 279	142	380 174 167	824 185 231	893 272 190	850 255 197
Calcium (mg/1 Ca). Magnesium (mg/1 Mg). Sodium (mg/1 Na). Potassium (mg/1 K). Iron (mg/1 Fe). Magnagnese (mg/1 Mn).	109 36 27.6 1.7 0.55 0.00	112 32 28.1 1.4 0.11 0.00	49 16 12.1 2.1 1.54 0.05	70 12.7 15.7 2.5 0.11 0.00	148 32 52.9 5.0 0.00 0.00	136 30 62.0 4.7 0.01 0.01	112 26 95.5 4.5 0.0 0.0
Chloride (mg/l Cl). Sulfate (mg/l SO ₄). Silfate (mg/l N). Silfate (mg/l N). Silfate (mg/l N). Shosphate (mg/l PO ₄). Soron (mg/l B). Silfac (mg/l SiO ₂).	77.4 210 18.2 0.01 0.2	25.7 157 20.4 0.00 0.11 0.17 19.9	5.4 151 0.16 0.00 0.0	8.5 71.4 6.9 0.10 0.32 0.06 15.4	88.3 288 0.00 0.2	105 247 12.7 0.04 0.05 0.14 15.8	79.6 264 4.1 0.0 0.3 0.0 16.6
Alkyl benzene sulfonate (mg/l ABS)							<0.1
Sample No.	37	39	21	24	32	19	30
Vell No. Date—1962. Femp., field/laboratory (°C.) H, laboratory. Dissolved oxygen (mg/l O ₂) Curbidity (units) Onductivity (µmho/em)	2928N 18 July 13.5/22 7.44 0.06 0.7	2928R 18 July 16/22 7.30 0.05 1.0	2918N 23 June 16.5/— 7.62 0.0 5.3 975	2909N 25 June 18/— 7.40 6.6 1.0 930	2909N 17 July 20/23 7.51 6.6 0.2 1137	1600L 18 June 13/— 7.82 8.6 0.25 875	1600L 16 July 16.5/22 7.4 6.0 0.4 983
Total dissolved solids (mg/l TDS). Total volatile solids (mg/l TVS). Ukalinity (mg/l CaCO ₃).	770 182 158	687 184 144	637 124 145	613 158 116	785 180 112	117	791 211 120
Calcium (mg/l Ca) Aagnesium (mg/l Mg) Odium (mg/l Na) Odium (mg/l Na) Odium (mg/l Fe) Aagnesium (mg/l Fe) Aagnese (mg/l Mn) Odium (mg/l Mn)	99 30 94.8 4.5 0.00 0.00	87 22 82.8 4.5 0.09 0.05	96.4 24 73.3 4.2 0.73 0.00	80.8 28 101 4.0 0.48 0.00	90 27 95.6 4.6 0.19 0.00	97 29 72.6 3.8 0.75	100 23 83.3 4.4 0.5 0.00
Chloride (mg/l Cl) ulfate (mg/l SO ₄) titrate (mg/l N) titrate (mg/l N) titrite (mg/l N) hosphate (mg/l PO ₄) toron (mg/l B). lilica (mg/l SiO ₂)	84 275 1.7 0.00 0.10 0.10 9.8	82.9 243 2.2 0.00 0.14 0.07 13.2	82.2 250 0.00 0.2	81.2 250 3.5 0.00 0.1	89.6 289 2.9 0.00 0.07 0.12 8.6	80 251 0.27 0.00 0.0	89 282 3.8 0.01 0.09 0.09 8.2
alkyl benzene sulfonate (mg/l ABS)	<0.1	<0.1			<0.1		
Sample No.	14	33	13	40	15	35	45
Vell No. Date—1962. emp., field/laboratory (°C.) H. laboratory Dissolved oxygen (mg/l Oz). urbidity (units). onductivity (µmho/cm).	1590F 16 June 14/— 7.60 10.0 0.15	1590F 17 July 23/23 7.53 6.7 0.8	1590D 16 June 15/— 7.74 7.4 0.1 975	1590D 17 July 19.5/22 7.58 5.6 1.0 863	1590M 16 June 13.5/— 7.59 8.7 0.35	1590M 17 July 17/23 7.44 5.3 1.0 1187	1592X 20 July 19/25 7.52 11.3 4.1
otal dissolved solids (mg/I TDS). otal volatile solids (mg/I TVS). lkalinity (mg/I CaCO ₂).	669 147 126	812 216 136	697 191 125	767 164 126	643 158 129	820 212 143	623 140 142
alcium (mg/l Ca)	92 27 85.2 3.5 0.19 0.00	94 26 87.4 4.3 0.22 0.00	105 36 83.9 3.5 0.02	98 26 103 4.4 0.09 0.02	97 19 85.5 3.5 0.03 0.00	103 24 81 4.2 0.12 0.01	104 21 52.6 4.0 1.2 0.02
hloride (mg/1 Cl), ulfate (mg/1 SO ₄), itrate (mg/1 N), itrate (mg/1 N), itrite (mg/1 N) hosphate (mg/1 PO ₄) oron (mg/1 B)	82.8 247 6.9 0.00	87.9 283 5.6 0.00 0.10 0.11	87.0 250 4.1 0.00 0.1	90.5 297 2.3 0.00 0.12 0.10	80.4 237 4.5 0.00	88 285 5.6 0.00 0.07 0.18	68.2 230 6.1 0.00 0.08 0.16
lkyl benzene sulfonate (mg/l ABS)		8.6 <0.1		<0.1		7.5 <0.1	20.0

Table 5-2-Summary of water analyses on integrated samples for shallow well network-continued

Sample No.	23	44	18	17	41	16	42
Well No. Date—1962 Temp., field/laboratory (°C.)	1592A 24 June 18/— 7.07 8.2 0.9 706	1592A 20 July 18/23 7.46 8.3 1.1	1563H 18 June 18.5/— 7.93 7.8 1.0 930	1562 17 June 17/— 7.56 7.4 0.4 794	1562 19 July 18/23 7.50 8.3 1.1 924	1561M 17 June 17.5/— 7.53 9.3 0.45	1561M 19 July 17.5/23 7.41 11.2 1.2 937
Total dissolved solids (mg/l TDS) Total volatile solids (mg/l TVS) Alkalinity (mg/l CaCO ₃)	534 201 121	595 111 137	179	553 183 188	627 197 192	546 180 175	889 225 177
Calcium (mg/l Ca) Magnesium (mg/l Mg) Sodium (mg/l Mg) Potassium (mg/l K) Fon (mg/l Fe) Manganese (mg/l Mn)	76.8 22 61.6 3.3 0.74 0.00	92 19 60.0 3.5 0.07 0.00	111 28 64.4 3.8 1.50 0.02	114 25 39.2 3.0 0.62 0.00	122 14 40.9 2.9 0.00 0.01	108 28 43.2 3.0 0.19 0.04	113 22 45.3 3.5 0.30 0.03
Chloride (mg/l Cl), culfate (mg/l SO ₄) itirate (mg/l N), itirate (mg/l N), itirite (mg/l N), chosphate (mg/l PO ₄), cloron (mg/l B), iliea (mg/l SiO ₂)	50.8 178 8.0 0.00	60.7 227 9.3 0.00 0.07 0.10 16.2	74.4 246 0.41 0.00 0.2	62.4 150 6.16 0.00 0.1	59.5 175 6.6 0.00 0.11 .025 20.8	61.8 140 9.7 0.00	66.2 201 7.8 0.00 0.09 0.15 20.1
lkyl benzene sulfonate (mg/l ABS)							

Considerable difficulty was encountered with the inflatable packers. While operation was very successful in 8- and 10-inch diameter wells, the packers have burst many times in the 12-inch diameter wells. At present the Los Angeles County Flood Control District is critically reviewing the design of the inflatable packers in order to increase the reliability of the unit in the larger wells.

5.03 Analyses of Integrated Samples.

The selective-depth pumping unit was not fabricated until March 1963. Prior to that time, starting in June and July 1962, integrated samples were collected with a portable pumping unit. They represent the water which enters a well from all depths. Table 5-2 is a summary of the chemical analyses performed on these samples. Rather complete mineral analyses were run in order to provide background information on the water quality. In several instances the methyleneblue assay for the alkyl benzene sulfonates (ABS) was performed, but no concentration of ABS was found in any of the wells. A blank space in the table indicates that no analysis was performed. No sample from well No. 1573J was made at this time because the well was not yet constructed.

5.04 Analyses of Multiple Depth Samples.

The purpose of sampling the shallow well network at multiple depths was twofold: first, to discover any changes in water quality attributable to the spreading of the reclaimed waters, and second, to delineate, if possible, any near-surface patterns in the ground-water movement. Because of the complex nature of the ground-water basin and the lack of any clear-cut tracers in the reclaimed water, the second purpose was not attainable.

Table 5-3 is a summary of the chemical analyses performed on the multiple-depth samples. Complete mineral analyses were not run because many of the components are not suitable tracers. Initially, it was thought that ABS was the most satisfactory tracer

for the reclaimed water; however, on this basis alone there has been no apparent change in water quality in the wells within the spreading grounds or the neighboring area.

Tables 5-3 and 5-4 show the results of analyses performed on multiple-depth samples taken from shallow wells at the downstream base of the Whittier Narrows Dam (see Figure 4-1, Chapter 4). These wells are referred to as "toe-drain" wells because they are used to control the phreatic surface through the earth dam. Three of these wells are shown in Figure 4-1, namely 2917T (TD#29), 2928 (TD#43), and 2928R (TD#64). The other toe-drain wells referred to in Table 5-3 by the notation "TD#" are in the same general location along the southern side of the dam. During the whole period of this investigation, these shallow wells were the only ones to show significant concentrations of apparent ABS. The concentration of ABS in these wells tended to be high in fall and winter season and low in the spring and summer. Concentrations of more than 1.0 mg/l were reached in several of the wells, but they decreased markedly on the next sampling to 0.1 mg/l. A concentration of 0.1-0.2 mg/l by the methylene-blue assay is not felt to be significant when done in our laboratory. It should be emphasized also that although the apparent ABS for some of the samples was more than 1.0 mg/l by the methylene-blue assay, these samples did not foam when shaken vigorously.

Samples from wells, such as the San Gabriel Valley Water Co. Plant W2, immediately south of the toedrain wells have had no apparent ABS. None of the wells sampled in this study have shown any tendency to foam. It is important to note that no intentional spreading of water is done upstream of the toe-drain wells, with exception of the small quantities spread at the Whittier Narrows Test Basin. Some reclaimed water will seep into the ground through the unlined transport canals north of the dam, but reclaimed waters in the canal in the vicinity of the toe-drain wells have been diluted most of the time with Colo-

Table 5-3-Summary of multiple depth samples for shallow well network

Sample Number	Well Number	Date 1963	Sample depth (ft)	Sample elevation† (ft)	ABS (mg/l)	Cl (mg/l)	Specific conductance (µmho/cm)	Field temperature (°C)	Field pH
W33 W34 W35 W36 W36 W37 W38 W47 W66 W67 W70 W70 W70 W70 W70 W89 W89 W89 W88	2936 2936 2936 2936 2936 2936 2936 2938 2938 2936 2936 2936 2936 2936 2936 2936 2936	14 Mar. 14 Mar. 14 Mar. 14 Mar. 14 Mar. 14 Mar. 14 Mar. 19 April 19 April 19 April 19 April 19 April 19 April 10 July 10 July 10 July 10 July 10 July 10 July	36 46 56 66 76 81 40 36 46 58 66 76 81 81 81	174.5 164.5 154.5 134.5 134.5 129.5 170.5 174.5 164.5 134.5 129.5 129.5 129.5 174.5 164.5 164.5	0.0 0.0 0.0 0.0 0.0 0.0 0.0 0.0	12 12 12 12 12 12 12 14 14 14 14 14 14 14 19 9.3 10.0 9.5	545 587 620 594 585 606 617 890 607 625 615 620 608 608 437 441 449	21 22 22 22 22 22 22 22 18.5 23 23 22.5 22 21 21.5 22.5 22.5 22.5 22.5 22.	7.2 7.1 7.2 7.1 7.1 7.1 7.1 7.1 7.1
W 49 W 102	2937V 2937V	20 Mar. 12 July	35 35	177.0 177.0	0.1 0.2	61 47.0	1,108 734	18 20.5	7.0 7.2
W21 W20. W22. W23. W60. W61. W62. W94. W91. W90. W127. W121.	2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T 2917T	8 Mar. 8 Mar. 8 Mar. 8 Mar. 18 April 18 April 10 July 10 July 6 Aug. 6 Aug.	40 46 46 52 40 46 52 40 46 52 40 46 52 40 46	158.2 152.2 152.2 146.2 158.2 158.2 146.2 158.2 146.2 146.2 158.2 146.2	0.1 0.3 0.1 0.2 0.1 0.1 0.1 0.2 1.0 1.1 0.4 0.9	85 78 84 82 106 104 106 79 . 7 90 . 8 94 . 5 85 . 1 96 . 9	804 923 809 846 1,215 1,230 1,220 831 779 790 791 756 779	21.5 18 23 22 23.5 27 25.5 22.5 22 24 24.5 24	6.9 6.9
W 12 W 13 W 14 W 15 W 52 W 53 W 54 W 54 W 93 W 97 W 97	2928N 2928N 2928N 2928N 2928N 2928N 2928N 2928N 2928N 2928N 2928N 2928N	11 Mar. 11 Mar. 11 Mar. 11 Mar. 18 April 18 April 18 April 19 July 9 July 9 July	34 40 40 44.5 34 40 40 41.5 34 40 45	152.4 146.4 146.4 141.9 152.4 146.4 141.9 154.4 141.9 154.4	0.1 0.1 0.1 0.3 0.2 0.1 0.3 0.3 0.4 0.3	75 69 73 73 98 96 97 95 81,9 83,7	947 1,000 961 951 1,040 1,100 1,115 1,140 831 831 845	23.5 19 21 21.5 23.5 18.5 24 23 25 25 25.5	7.0
W9	2925R 2925R 2928R 2928R 2928R 2928R 2928R 2928R 2928R 2928R 2928R 2928R	8 Mar. 8 Mar. 8 Mar. 8 Mar. 17 April 17 April 17 April 19 July 9 July 9 July	34 40 40 46 34 40 40 46 34 40	154.7 148.7 148.7 142.7 154.7 148.7 148.7 142.7 154.7 142.7	0.5 0.5 0.5 0.4 0.6 0.6 0.6 0.6	80 81 79 80 94 93 93 94 72.5 73.4	951 951 951 951 1,040 1,040 1,035 1,040 734 734	21 19 24 23.5 	7.3
W48 W106	2918N 2918N	13 Mar. 18 July	60 60	127.2 127.2 77.2	0.0 0.1 0.0	74 80.2 77.1	1,051 756 768	18 22 22.5	7.2
W152 W6 W73 W142 W143 W144	2918N 2909N 2909N 2909N 2909N 2909N	22 Aug. 25 Feb. 22 April 16 Aug. 16 Aug. 16 Aug.	45 45 75 85 95	132.0 132.0 102.0 92.0 82.0	0.2 0.1 0.1 0.1 0.1	33 95 84.4 84.5 83.5	1,620 1,090 831 860 845	20.5 15.5 20 20 19.5	7.5 7.6 6.9
W16	1600L 1600L 1600L 1600L 1600L 1600L	6 Mar. 6 Mar. 6 Mar. 6 Mar. 20 Aug. 20 Aug. 20 Aug.	90 93.5 104.5 114.5 95 105 115	79.9 76.4 65.4 55.4 74.9 64.9 54.9	0.1 0.1 0.1 0.1 0.0 0.1	54 53 53 54 75.8 73.7 74.5	1,300 1,300 1,280 1,330 860 779 818	20 21 20.5 21 22.5 24	6.5
W7 W76	1590F 1590F	21 Feb. 22 April	42 39	126.8 129.8	0.3	60 63	1,050 811	21 22	7.9
W29 W129 W128	1590D 1590D 1590D	7 Mar. 9 Aug. 9 Aug.	99 80 90	68.4 87.4 77.4	0.2 0.1 0.1	69 79.7 83.9	1,025 722 722	19.5 24 22.5	7.0
W1 W2 W3 W4 W5 W5 W149	1590M 1590M 1590M 1590M 1590M 1590M 1590M	15 Feb. 15 Feb. 15 Feb. 15 Feb. 15 Feb. 21 Aug. 21 Aug.	40 50 60 70 80 72 78	118.1 108.1 98.1 88.1 78.1 86.1 80.1	0.2 0.2 0.1 0.1 0.1 0.1	79 78 78 78 82 92.2 91.7	875 880 880 875 900 875 875	19.5 20 20.5 19.5 22.5	7.0
W51	1592A 1592A 1592A 1592A 1592A	20 Mar. 12 July 23 Aug. 23 Aug. 23 Aug.	75 80 85 100 130	81.9 76.9 71.9 56.9 26.9	0.1 0.1 0.0 0.1 0.1	79 84.4 79.3 79.4 80.4	1,155 831 818 860 891	18 17 22 21.5 22	6.9 7.0 6.9

Table 5-3-Summary of multiple depth samples for shallow well network-continued

				Sample			Specific	Field	
Sample Number	Well Number	Date 1963	Sample depth (ft)	elevation†	ABS (mg/l)	Cl (mg/l)	conductance (µmho/cm)	temperature (°C)	Field pH
W31. W30. :W32. W63. W64. W65. W101. W104.	1592 X 1592 X 1592 X 1592 X 1592 X 1592 X 1592 X 1592 X 1592 X 1592 X	12 Mar. 12 Mar. 12 Mar. 18 April 18 April 18 April 11 July 11 July 11 July	169 175 175 169 175 181 169 175	-19.0 -25.0 -25.0 -19.0 -25.0 -31.0 -19.0 -25.0 -31.0	0.1 0.1 0.0 0.0 0.0 0.0 0.0 0.0	67 66 65 93 89 83 90.7 81.3 80.2	1,090 1,090 1,090 993 970 954 756 744 744	21 18.5 21 20 20 20 22.5 21.5 20.5	7.0
W24 W25 W26 W27 W74 W75 W99	1573J 1573J 1573J 1573J 1573J 1573J 1573J 1573J 1573J	11 Mar. 11 Mar. 11 Mar. 11 Mar. 19 April 19 April 11 July 11 July	110 124 130 138 124 130 124 130	31.5 17.5 11.5 3.5 17.5 11.5 17.5	0.0 0.1 0.1 0.1 0.0 0.0 0.1	84 84 84 83 101 101 94.2 94.2	923 904 924 924 1,130 1,130 860 860	18.5 21 20.5 20.5 20.5 20 22 22.5	7.0
•W50	1562	20 Mar.	166	-17.5	0.0	66	963	17.5	7.0
W 40 W 41 W 39 W 42 W 107 W 109 W 108	1561M 1561M 1561M 1561M 1561M 1561M 1561M	15 Mar. 15 Mar. 15 Mar. 15 Mar. 17 July 17 July 17 July	135 145 146 155 138 145 150	26.0 16.0 15.0 6.0 23.0 16.0	0.0 0.0 0.0 0.0 0.1 0.1 0.1	68 68 67 66 76.3 77.5	932 946 943 921 580 702 791	21 21 17.5 21 28 28.5 28.5	7.1 7.1 7.1

† Sample elevation above mean sea level measured relative to reference point in well. ‡ Packer ruptured. * Integrated sample taken without using packers.

rado River water. If the canals are filled with water for an extended period of time, one can expect the soil beneath the canal to become saturated. North of the dam the ground-water table is close to the ground surface, about 10-12 feet down, and it is quite possible that reclaimed water reached the ground-water table without seeping through a soil zone with a highly developed aerobic culture such as exists in the near surface soil of the test basins. Under anaerobic conditions, ABS is readily transmitted with the water and not removed and degraded. Local geologic conditions associated with a high ground-water table could then cause this water to be short-circuited to the shallow toe-drain wells.

There has been no evidence to date that the spreading of reclaimed water has affected the quality of water in wells in the Montebello forebay. The fact that the reclaimed water and Colorado River water have quite similar chemical characteristics has made difficult the tracing of reclaimed water in the forebay. The most sensitive tracer, which occurs in wastewater but not in natural or Colorado River water, is ABS. From the results of this sampling program it is apparent that the ABS from wastewater has not reached water-supply wells in the Montebello forebay, partly because it has been removed in aerobic percolation and partly because of the high dilution of wastewater by natural waters.

Table 5-3—Summary of multiple depth samples for shallow well network—continued

Sample Number	Well Number	Date	Sample Depth (feet)	ABS (mg/l)	Cl (mg/l)	Specific Conductance (µmho/cm)	Field Temp. (°C)	Field pH
W235. W236. W2374. W238. W246. W239. W249. W299. W299. W298. W334. W335. W335. W337.	2936 2936 2936 2936 2936 2936 2936 2936	29 Oct. 63 29 Oct. 63 29 Oct. 63 29 Jan. 64 29 Jan. 64 14 Oct. 64 14 Oct. 64 14 Oct. 64 14 Oct. 64 123 April 65 23 April 65 23 April 65	36 46 56 36 52 60 36 46 56 66 46 56 66	0.0 0.0 0.0 0.0 0.0 0.0 0.1 0.0 0.0 0.0	7.3 6.3 5.6 5.6 4.3 4.5 3.8 3.6 2.5 3.2 3.2	399 378 320 321 318 323 365 359 	24 24 24 18.5 21 22 22 22 20.5 23 23 24 24 25	7.2 7.1 7.1 7.1
W232 W233 W251 W252 W356 W352	1573J 1573J 1573J 1573J 1573J 1573J	30 Oct. 63 30 Oct. 63 31 Jan. 64 31 Jan. 64 30 April 65 30 April 65	120 130 120 130 125 146	0.1 0.1 0.0 0.0 0.0 0.0	89.5 88.1 90.5 85.3 79.5 81.4	1,142 1,124 1,030 1,050	23 24 20.5 20.5 25 23	6.9 6.9
W295 W297 W292 W353 W357 W361	1561M 1561M 1561M 1561M 1561M 1561M	14 Oct. 64 14 Oct. 64 14 Oct. 64 29 April 65 29 April 65 29 April 65	138 145 150 138 145	0.1 0.1 0.1 0.0 0.0 0.0	88.2 80.7 82.3 83.4 82.0 82.7	967 990 990 1,000 1,000 1,020	21 21 20.5 25.5 25.5 24	=======================================
W360	1562	30 April 65	166	0.0	82.3	1,000	17.5	
W172 W293	2917B 2917B	16 Sept. 63 15 Oct. 64	::	0.2	83.4 313	1,015 1,400	::	

Table 5-3-Summary of multiple depth samples for shallow well network-continued

Sample number	Well number	Date	Sample depth (feet)	ABS (mg/l)	Cl (mg/l)	Specific conductance (µmho/cm)	Field temp. (°C)
W302	2917B	9 April 65		0.2	183	1,180	
W327	1590D	28 April 65	99	0.0	71.4	891	15
V332	1590F	28 April 65	42	0.0	86.5	1,020	14
V179	2947V	30 Sept. 63		0.1	14.0	378	
V253V273	2947Z 2947Z	15 June 64 30 Sept. 64	::	0.0	27.5 27.9	576 565	19.5 19.5
V167 V254 V269	2938D 2938D 2938D	16 Sept. 63 15 June 64 30 Sept. 64	==	0.1 0.1 0.1	41.7 45.5 47.2	800 864 834	17 19.5
V173	2948 2948	25 Sept. 63 17 June 64	::	0.0 0.1	33.1 22.4	526 555	18
V166 V256 V274	2939B 2939B 2939B	16 Sept. 63 15 June 64 30 Sept. 64	Ξ	0.2 0.1 0.2	66.0 70.4 72.5	1,015 1,000 1,050	17 21,5
7174 7263 7265	1620GG 1620GG 1620GG	16 Sept. 63 17 June 64 30 Sept. 64		0.1 0.1 0.1	83.2 74.6 74.8	1,050 990 976	17.5 17
7193	1620BB	23 Sept. 63		0.1	77.9	1,010	
7164 257 267	1621SS 1621SS 1621SS	16 Sept. 63 15 June 64 30 Sept. 64	==	0.1 0.1 0.1	65.5 57.0 70.6	920 919 944	17 17
7165 7176 7261 7271	1621S 1613S 1613S 1613S	16 Sept. 63 16 Sept. 63 15 June 64 30 Sept. 64	=	0.1 0.1 0.1 0.1	79.5 16.9 15.3 16.5	1,000 550 569 543	18.5
177	1615R	16 Sept. 63		0.1	66.5	1,020	
175 260 266	1596H 1596H 1596H	16 Sept. 63 15 June 64 30 Sept. 64	=	0.1 0.1 0.1	41.1 38.8 48.0	740 745 794	17 18.5
188	1606	25 Sept. 63		0.1	43.7	797	
191	1606X	25 Sept. 63		0.1	46.3	771	
169 259 264	1597BB 1597BB 1597BB	16 Sept. 63 15 June 64 30 Sept. 64	==	0.1 0.1 0.1	46.7 50.7 53.0	803 846 836	17
170 255 268	1587 Y 1587 Y 1587 Y	16 Sept. 63 15 June 64 30 Sept. 64	==	0.1 0.1 0.1	30.2 31.1 16.9	705 732 557	18 20
333	1590M	28 April 65	60	0.0	79.3	976	14
330	1592A	27 April 65	80	0.0	89.2	1,070	18
171 258 272	1598K 1598K 1598K	16 Sept. 63 15 June 64 30 Sept. 64	=	0.1 0.1 0.1	26.9 29.5 31.6	671 708 698	18 17.5
270	1612Q	30 Sept. 64		0.1	79.6	965	16.5
328	1600L	28 April 65	90	0.0	86.9	1,070	14
331	2909N	27 April 65	50	0.0	27.8	478	15
329	2918N	27 April 65	60	0.6	87.1	976	18.5

Table 5-4—Summary of analyses for Whittier Narrows dam toe drain wells

Sample No.	Well No.	Date	Sample depth (ft.)	ABS (mg/l)	Cl (mg/l)	Specific conduct- ance (µmho/ em)	Sample No.	Well No.	Date	Sample depth (ft.)	ABS (mg/l)	Cl (mg/l)	Specific conduct- ance (µmho/ cm)
W130 W141 W140 W200 W204	TD#21 TD#21 TD#21 TD#21 TD#21 TD#21	9 Aug. 63 9 Aug. 63 9 Aug. 63 24 Oct. 63 24 Oct. 63	34 40 46 34 40	0.1 0.2 0.2 0.4 0.4	56.1 52.5 53.6 73.2 73.6	683 702 692 871 884 889	W219 W223 W245 W243 W247	2917T 2917T 2917T 2917T 2917T	25 Oct. 63 25 Oct. 63 28 Jan. 64 28 Jan. 64 28 Jan. 64	46 52 40 46 52	0.5 0.5 0.0 0.1 0.1	95.5 97.5 85.0 85.4 85.6	981 967 1,110 1,060 1,080
W208 W133	TD#21 TD#26	24 Oct. 63 9 Aug. 63	46 34	0.4	75.8 60.7	692	W316 W320 W324	TD#21 TD#21 TD#21	20 Apr. 65 20 Apr. 65 20 Apr. 65	34 40 46	0.1 0.0 0.0	84.0 82.3 83.4	1,030 1,020 1,030
W132 W131 W220 W216 W212	TD#26 TD#26 TD#26 TD#26 TD#26	9 Aug. 63 9 Aug. 63 23 Oct. 63 23 Oct. 63 23 Oct. 63	40 46 34 40 46	0.2 0.2 0.5 0.5 0.4	60.4 60.2 73.8 75.8 72.5	734 712 816 806 800	W286 W290 W279 W305	TD#26 TD#26 TD#26 TD#26	8 Oct. 64 8 Oct. 64 8 Oct. 64 20 Apr. 65	34 40 46 34	0.5 0.5 0.4 0.0	80.5 84.2 79.9 90.4	945 960 990 1,080
W134 W135 W136	TD#32 TD#32 TD#32	8 Aug. 63 8 Aug. 63 8 Aug. 63	34 40 46	0.3 0.4 0.4	\$6.7 \$8.9 \$9.8	845 860 860	W309 W313	TD#26 TD#26	20 Apr. 65 20 Apr. 65	40 46	0.0	89.9 91.3	1,080 1,080
W214 W218 W222	TD#32 TD#32 TD#32	25 Oct. 63 25 Oct. 63 25 Oct. 63	34 40 46	0.3 0.4 0.3	\$8.8 \$9.8 \$8.1	1,015 990 995	W303 W307 W311	TD#32 TD#32 TD#32	21 Apr. 65 21 Apr. 65 21 Apr. 65	34 40 46	0.4 0.4 0.2	100 99.8 99.0	1,180 1,110 1,100
W118 W117 W202	TD#35 TD#35 TD#35	7 Aug. 63 7 Aug. 63 24 Oct. 63	40 46 34	0.4 0.6 0.4	\$8.5 89.6 88.5	924 845 1,122	W304 W308 W312	TD#35 TD#35 TD#35	21 Apr. 65 21 Apr. 65 21 Apr. 65	34 40 46	0.4 0.3 0.4	102 98.4 101	1,210 1,280 1,140
W206 W210 W120	TD#35 TD#35 TD#39	24 Oct. 63 24 Oct. 63	40 46 34	0.3 0.4 0.3	\$3.0 \$6.6 \$3.1	1,066 1,060 804	W340 W343 W347	TD#39 TD#39 TD#39	22 Apr. 65 22 Apr. 65 22 Apr. 65	34 40 46	0.4 0.6 0.4	102 101 99.4	1,090 1,090 1,090
W123 W126 W201 W205	TD#39 TD#39 TD#39 TD#39	7 Aug. 63 7 Aug. 63 7 Aug. 63 24 Oct. 63 24 Oct. 63	40 46 34 40	0.2 0.5 0.2 0.2	83.7 85.6 72.5 72.0 74.4	818 791 942 941	W341 W342 W346	TD#48 TD#48 TD#48	22 Apr. 65 22 Apr. 65 22 Apr. 65	34 40 46	0.7 0.7 0.6	102 102 102	1,090 1,090 1,090
W209 W116	TD#39 TD#48 TD#48	24 Oct. 63 7 Aug. 63	46 34 40	0.3	74.4 78.4 81.1	949 791 701	W339 W344 W348	TD#54 TD#54 TD#54	23 Apr. 65 23 Apr. 65 23 Apr. 65	34 40 46	0.6 0.5 0.5	97.5 98.0 97.4	1,070 1,080 1,070
W119 W122 W213 W217 W221	TD#48 TD#48 TD#48 TD#48	7 Aug. 63 7 Aug. 63 25 Oct. 63 25 Oct. 63 25 Oct. 63	46 34 40 46	0.3 0.7 0.3 0.3 0.3	90.0 55.8 61.4 88.1	791 756 771 806 995	W338 W345 W349	TD#59 TD#59 TD#59	23 Apr. 65 23 Apr. 65 23 Apr. 65	34 40 46	0.6 0.6 0.5	98.5 97.1 97.7	1,080 1,070 1,080
W125	TD#54	7 Aug. 63	34	0.4	73.3 74.9	768	W291	2928R (TD#64)	8 Oct. 64	34	0.5	69.3	90-
W115 W224 W226 W229	TD#54 TD#54 TD#54 TD#54	7 Aug. 63 28 Oct. 63 28 Oct. 63 28 Oct. 63	46 34 40 46	0.3 0.3 0.3 0.3	74.9 51.9 50.5 49.8	744 794 782 794	W283 W280 W318 W326	2928R 2928R 2928R 2928R	8 Oct. 64 8 Oct. 64 30 Mar. 65 30 Mar. 65	40 46 34 40	0.5 0.5 0.8 0.6	62.9 69.8 102 97.6 97.2	865 888 1,080 1,060
W139 W138 W137	TD#59 TD#59 TD#59	8 Aug. 63 8 Aug. 63 8 Aug. 63	34 40 46	0.2 0.4 0.6	53.6 69.5 77.2	756 768 906	W322 W289	2928R 2928N	30 Mar. 65 9 Oct. 64	46 34	0.8	96.0	1,07
W155	2928R (TD#64)	8 Oct. 63	34	0.9	79.7	920	W275 W277	(TD#43) 2928N 2928N	9 Oct. 64 9 Oct. 64	40 45	0.4 0.3 0.7	100 96.2	1,073
W158 W156 W203	2928R 2928R 2928R	8 Oct. 63 8 Oct. 63 23 Oct. 63	40 46 34	0.8 1.0 0.8	80.6 81.6 81.0	920 920 981	W319 W323 W315	2928N 2928N 2928N	30 Mar. 65 30 Mar. 65 30 Mar. 65	34 40 45	0.7 0.5 0.7	99.6 99.2 101	1,080 1,060 1,060
W207 W211	2928R 2928R	23 Oct. 63 23 Oct. 63	40 46 34	0.6	78.6 80.5 86.4	981 932 946	W281	2917T (TD#29)	9 Oct. 64	46	0.8	107	1,080
W240 W241 W242	2928R 2928R 2928R	28 Jan. 64 28 Jan. 64 28 Jan. 64	40 46	0.3 0.4 0.4	86.5 86.4	946 970	W284 W306 W310	2917T 2917T 2917T	9 Oct. 64 20 Apr. 65 20 Apr. 65	52 40 46	1.1 0.1 0.1	116 98.7 97.9	990 1,150 1,170 1,170
W157	2928N (TD#43)	8 Oct. 63	34	0.3	71.1	920	W314 W288	2917T TD#134	20 Apr. 65 13 Oct. 64	52 34	0.1	98.7 99.1	1,01
W162 W160 W231 W228 W225	2928N 2928N 2928N 2928N 2928N 2928N 2928N	14 Oct. 63 14 Oct. 63 28 Oct. 63 28 Oct. 63 28 Oct. 63 28 Jan. 64	34 40 34 40 45 34	0.2 0.3 0.2 0.2 0.3 0.1	63.0 63.0 57.6 78.8 59.3 87.4	870 859 821 852 782 970	W 288 W 287 W 285 W 358 W 354 W 350	TD#134 TD#134 TD#134 TD#134 TD#134 TD#134	13 Oct. 64 13 Oct. 64 13 Oct. 64 3 May 65 3 May 65 3 May 65	40 46 34 40 45	1.1 1.3 0.2 0.2 0.1	98.1 105 94.0 93.7 94.1	1,040 1,080 1,100 1,110 1,110
W248 W249 W244	2928N 2928N 2928N	28 Jan. 64 28 Jan. 64	40 45	0.1	88.1 88.2	970 970	W282	TD#135	13 Oct. 64	34	1.3	111 112	1,12: 1,100
W163 W161 W159 W215	2917T (TD#29) 2917T 2917T 2917T	14 Oct. 63 14 Oct. 63 14 Oct. 63 25 Oct. 63	40 46 52 40	0.6 0.8 0.7 0.4	87.8 88.1 90.0 96.7	958 937 937 972	W276 W278 W359 W355 W351	TD#135 TD#135 TD#135 TD#135 TD#135	13 Oct. 64 13 Oct. 64 3 May 65 3 May 65 3 May 65	40 46 34 40 46	1.3 1.2 0.1 0.1 0.1	112 115 66.3 70.6 70.9	1,100 890 910 910

LABORATORY SOIL COLUMN INVESTIGATIONS

6.01 The Purpose of the Study.

Experience and information obtained from the field spreading basin study have pointed out the need for the accelerated accumulation of data on the effects of aerobic percolation through a soil system. While the field study is quite valuable, the lack of control over many of the variables of the spreading operation often creates great difficulties in the interpretation of data. Therefore, it was decided to set up in the laboratory multiple columns of media (soil and sand) which could be dosed at higher rates and heavier loads than occur under present conditions in the field. Closer control over this complex biological system is a primary consideration. It is advantageous to develop a laboratory setup in which special attention can be given to the optimization of the necessary parameters for the degradation of refractory compounds, in this case a highly aerobic soil system for rapid removal of synthetic detergents.

6.02 The Intermittent Laboratory Sand Filter.

A new design for a laboratory soil column incorporated the features necessary to permit operation of the column as a respirometer as well as a lysimeter. The column is shown schematically in Figure 6-1. It consists basically of sections of four-inch diameter, anodized aluminum tubing. Each column is a combination of two types of modules, a filter section and a sampling section. No section is more than eighteen inches long; hence the entire section can be placed in the laboratory autoclave to be sterilized if desired.

The filter section is a straight piece of tubing with an aluminum screen of about 8 mesh mounted near one end. Sand or other filter media are retained in this section. A sampling section is made from a connector piece used to join sections of four-inch aluminum irrigation pipe. One end of the sampling section is closed, so that the unit looks simply like a small coffee can. Mounted in the aluminum end plate of the sampling section are two 3-inch aluminum tubes. One tube is U-shaped and is a siphon which causes water collected in the sampling section to be emptied automatically and discharged out the bottom when the level of the water reaches some specific height. The second tube is a straight pipe which permits gas to flow through the sampling section even though it may be partially filled with water causing the siphon tube to be sealed off. A third tube is mounted through the side of the sampling section. This is a port which permits emptying of the water from the sampling section by not letting the water level rise to the height necessary to prime the siphon.

A complete laboratory column consists of alternate sampling and filter sections. A whole unit can be constructed of sand beds of arbitrary depths. Each sampling section permits the sampling of the percolate after it has passed through a given depth of sand without disturbing the sand bed or the pattern of flow through the bed. The columns can be modified at will by simply removing a section without disturbing the rest of the column.

In order to operate the columns as respirometers, it is necessary to make them airtight. This is accomplished by using O-ring seals at the joints between the modules and taping the joint on the outside with a plastic tape. Preliminary studies have shown that it is possible to run the columns with a co-current forced flow of air and water. However, the experiments discussed in this chapter were run with a natural ventilation only, that is, the air is drawn into the columns as a result of the intermittent dosing with the liquid feed.

Each time a sampling section discharges its liquid feed, a natural draft is produced drawing air into the column. Because of the multiple sampling sections in each column, a given volume of liquid causes the column to be "aerated" several times as it passes down the column.

6.03 Design of Columns Used in Study.

Five sand columns were fabricated and set for a preliminary laboratory experiment. These columns each consist of four filter sections alternated with sampling sections between. The uppermost sand bed is two inches deep, the second four inches, the third eight inches, and the bottom section is twelve inches for a total sand depth of 26 inches for each of the columns. The material used in the columns is a coarse sand having a mean size of about 1.3 mm and a geometric standard deviation of 1.21. This is a foundry grade sand and was quite clean when purchased. The table below summarizes the distribution of sand in the columns.

Nominal Sand Depth (inches)	Weight of Sand	l in Section (kg.)
(thenes)	(103.)	(Ny.)
2	1.36	0.62
4	2.76	1.25
8	5.28	1.76
12	8 19	3.72

To keep the sand from passing through the support screens a thin layer of pea gravel was placed on top of each filter support screen.

6.04 Method of Seeding the Columns.

In order to develop a rich biological flora in the sand beds each of the columns was fed with heavy doses of settled sewage from the Whittier Narrows Treatment Plant. A single column was erected initially and fed for about 25 days with settled sewage at hydraulic loads from 1 to 3 meters per day. After 20 days of operation of the first columns, four more columns were started and fed raw settled sewage at similar loads. These columns were operated for about 5 days and then all five columns were changed to a daily load of an artificial feed consisting of about 100

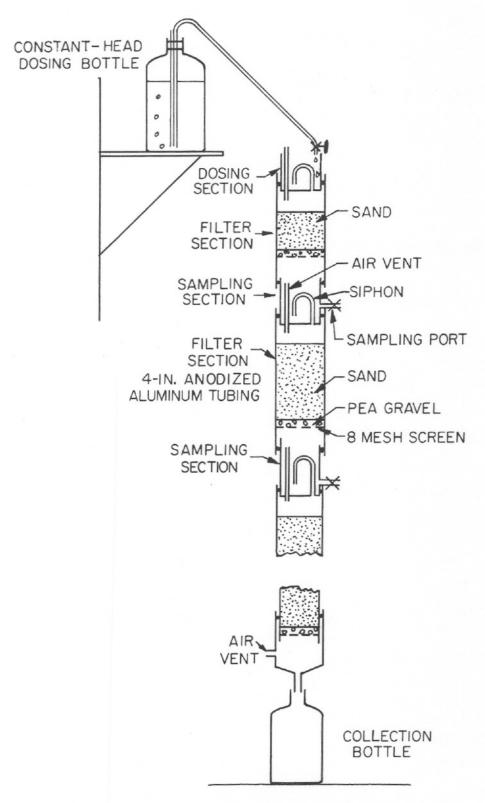


Fig. 6-1—Schematic of Laboratory Soil Column

nl of nutrient broth in 24 liters of Pasadena tap water. A loading of 24 liters per day on a diameter of 4 inches corresponds to a hydraulic load of about 3 meters per day. After 5 days of loading, the synthetic feed was changed to about 100 ml of nutrient broth plus a liter of settled sewage in 24 liters of tap water. The columns were loaded for 15 days with this feed and then the composition was changed to about 200 ml of nutrient broth plus 2 liters of settled sewage in 24 liters of tap water which was fed to the columns for about 9 days. Next the columns were dosed with a sterile feed consisting of about 300 ml of nutrient broth in 24 liters of tap water. This feed was applied for about 25 days. At the end of this seeding period, the first controlled experiments were started. The unusually long seeding period, about 80 days, was partially due to the need for developing loading procedures and sampling techniques and also waiting for an adequate supply of the new blend of synthetic detergent, the linear alkylate sulfonate (LAS).

The study described in this chapter was concerned principally with the comparative behavior of the synthetic detergents linear alkylate sulfonate (LAS) and the conventional alkylbenzenesulfonate (ABS) in an aerobic intermittent sand filter of special design. A careful study of the effects of multiple loadings at different detergent concentrations was carried out. Moreover, the effects of high oxygen demands attributable to ammonia and the influences of pH change on the quality of the percolate were observed. The ABS used in this study was supplied by the Soap and Detergent Association and is identified as CAL-RESEARČH 61R-8008 (55.3% ABS). Mr. George Cook of the California Chemical Company very kindly supplied the LAS, namely Soap and Detergent Association Blend No. 2, CALRESEARCH 64R-5808 (47.7% LAS).

6.05 Method of Operation

The LAS and ABS were added to artificial feeds consisting of nutrient broth, ammonium chloride or sodium nitrate, and tap water. The major criterion for the feed was to make it similar organically to a highly treated activated-sludge effluent with the exception that the feed would not be seeded with bacteria or other organisms.

The columns were divided into three groups. Two columns were fed LAS and two received ABS. The fifth column received the same basic feed but no synthetic detergent. It acted as the control. An arbitrary hydraulic loading of 1 gallon per day per column was chosen strictly on the basis of the convenience of using bottles that happened to be on hand in the laboratory. This is approximately a dosage of 0.49 meter per day (1.6 ft per day). The characteristics of the feed solutions are given in Table 6-1. The nutrient-broth stock solution has a COD of about 5.8 mg per ml and an organic nitrogen concentration of about 0.7 mg N per ml. The total COD of the feed solutions was not due to nutrient broth only, for there is an appreciable COD from the synthetic detergents. Measured COD concentrations of the ABS and LAS solutions indicate

Table 6-1-Characteristics of feed solutions Ingredients added to one gallon of tap water

Period (days)	Column	ABS (mg)	LAS (mg)	Nutrient broth* (ml)	NH ₄ Cl (mg)	NaNO: (mg)
0 to 129†	A1	20		25	500	
0 to 129†	A1 A2	40		50	1,000	
0 to 129†	L1		20	25	500	
0 to 129†	L2		40	50	1,000	
0 to 129†	Control			50 25	500	
30 to 353†	A1	20		25		750
30 to 353†	A2	40		50		1,500
30 to 353†	L1		20	25		750
30 to 353†	L2		40	50		1,500
30 to 353†	Control			25		750

† Days 72 to 140; 5 ml of about 1N Na2CO2 was added per gallon daily to buffer feed

solutions.

* Stock nutrient broth solution containing 3 g. beef extract and 5 g. peptone in 1 liter

Ingredients of feed solutions on a concentration basis

Period (days)	Column	ABS (mg/l)	LAS (mg/l)	COD (mg/l)	Org. N + NHa (mg N/l)	NO ₃ —N (mg N/l)
0 to 129	A1	5.3		49.5	39.3	
0 to 129	A2	10.6		99.0	78.7	
0 to 129	L1		5.3	51.6	39.3	
0 to 129	L2		10.6	103.3	78.7	
0 to 129	Control			38.4	39.3	
130 to 353	A1	5.3		49.5	4.6	32.7
30 to 353	A2	10.6		99.0	9.2	65.4
30 to 353	L1		5.3	51.6	4.6	32.7
30 to 353	L2		10.6	103.3	9.2	65.4
30 to 353	Control			38.4	4.6	32.7

that there are 2.1 mg COD per mg ABS and 2.5 mg

COD per mg LAS.

The feeds described in Table 6-1 were continuously applied to each of the columns on a daily basis. Table 6-2 contains a chronological history of unusual events, namely, days the columns were not fed, breakdowns, errors in loading, etc. Initially the feed was delivered to the columns by a peristaltic pump, but continuing breakdowns of the pump caused a changeover to constant-head bottles for the feed solutions. This method proved to be a much more reliable and more easily maintained feeding system.

The columns were loaded hydraulically by continuously applying the feed into a sample section at the top of each column by a pump or constant-head bottle. A volume of about 500 ml was necessary to activate the siphon. The rate of flow into the top of the column was such that it required about 14 to 18 hours to apply one gallon of feed. The siphon tubes discharge about 8 times in this period, sending about 500 ml through the column per discharge. For the clean columns the residence time was less than four minutes. Nearly all of this time could be accounted for in time required to fill and empty the various sampling sections down the column. The residence time in the coarse sand was negligible. After the development of a biological flora in the sands, the residence time increased to about 10 minutes. Because of the slight differences in the volume of the sampling sections, the residence time could vary, that is, a sampling section midway down the column may fill up just to a level slightly less than the height necessary to activate the siphon. This condition means that this section will not discharge until much later when the next load is fed into the top of the column.

6.06 Types of Experiments and Their Durations.

The experiment was monitored by collecting the entire effluent from each column and performing chemical analyses on these samples. The results reported in this section are primarily for the water that passed through the entire 26 inches of sand. The data have been broken down into periods described below.

Experiment I: This experiment covered the first 71 days of the column operation during which time the columns were loaded each day with two exceptions, Day 50 when none were fed and Day 55 when neither L1 nor A2 was loaded. For the first 15 days of this experiment surfactant concentration was measured in the effluent from each column by the methylene-blue assay. Ammonia and organic nitrogen analyses were also run daily for the first 15 days to assess the degree of nitrification taking place in the columns. Determinations of COD, nitrite, nitrate, pH, temperature, TDS, and TVS were run arbitrarily during the first 15 days. Subsequently, measurements of surfactants, COD, nitrogen components, and pH were monitored on a weekly basis.

Experiment II: Because the measured pH of the column effluents showed values below pH 6, it was decided to see if the performance of the sand filters could be improved by attempting to buffer the feeds. Days 72 to 140 were characterized by the addition of about 5 ml of about 1N Na₂CO₃ to the one-gallon feed for the columns. In all other respects, the feed

solutions were identical to those used during the tial 71 days of operation until Day 130 when ammonium chloride in the feed solutions was replaced by sodium nitrate. This change in the oxidation of the main source of the nitrogen in the column was part of the attempt to delineate the principal cause of the large drop in pH of the feed solution resulting from passage through the filters.

Experiment IIa: During the interval covered Days 108 to 129, the column effluents from the daily loading were recycled once more through the column. This change caused each column to receive twice the hydraulic load normally applied to it.

Experiment III: This was the terminal experiment performed in the sand filters. The addition of the sodium carbonate buffer was stopped on Day 140 For the period Days 141 to 353, the original hy. draulic loading pattern of one gallon per day of feed to each column was followed with the nitrogen in the feed solutions being principally in the form of nitrate During the interval covered by Days 289 to 293, the columns were re-seeded with micro-organisms by feeding each one gallon of a mixture of one-half primary effluent settled sewage) and one-half tap water. Normal feeding of the sterile solutions began again on Day 294. An error in preparing the stock solution of the ABS resulted in columns A1 and A2 receiving only one-half the expected concentration of ABS after Day 336 until the end of the experiment.

Table 6-2—CHRONOLOGY OF LABORATORY COLUMNS

Day	Date	Remarks
0	11 May 1964	Comparative Detergent Study began using artificial feed (see Table 6-2) under schedule of daily loading of columns.
50	30 June 1964	Columns not loaded.
55	5 July 1964	Columns L1 and A2 not loaded.
64	14 July 1964	Tube pump used to feed columns replaced with constant-head bottles.
72	22 July 1964	Feeds to columns buffered by adding 5 ml of approximately 1N Na ₂ CO ₃ .
84	3 August 1964	Column A2 received feed for column L2 and vice versa.
108	27 August 1964	New loading pattern initiated. Effluent from columns recycled so that a given feed passes through a column twice in 24 hour period.
116	4 September 1964	Columns not loaded.
128	16 September 1964	Top section of Column L2 (2-in sand bed) scarified and loosened up.
130	18 September 1964	Stopped recycling feeds. Replaced NH ₄ Cl with NaNO ₃ in all feeds. Returned to hydraulic load of 1 gallon per 24 hour period.
141	1 October 1964	Discontinued addition of Na ₂ CO ₃ to feeds.
211	8 December 1964	Each column loaded twice for sampling purposes.
218	15 December 1964	Each column loaded twice for sampling purposes.
229	26 December 1964	Each column loaded twice for sampling purposes.
232	29 December 1964	Each column loaded twice for sampling purposes.
282	17 January 1965	Columns not loaded.
289	24 February 1965	Primary effluent diluted 1:1 with tap water fed to all columns. One gallon mixture each day for 5 days.
294	1 March 1965	Resumed feeding columns with artificial feeds.
302	9 March 1965	Columns not loaded.
304	11 March 1965	Columns not loaded.
324	31 March 1965	Columns not loaded.
327	3 April 1965	Columns not loaded.
336	12 April 1965	Stock solution for ABS made at one-half strength by error so that Columns A1 and A2 received only 2.6 mg/l and 5.3 mg/l respectively.
354 - 360	30 April-6 May 1965	Columns not loaded.
361-364	7-10 May 1965	Columns loaded again for purposes of measuring moisture content of sands.
365	11 May 1965	Columns dissembled.

6.07 Results of Analyses.

The results of the experiments described in the previous section are presented in tabular form in this section. Tables 6-3 to 6-9 show the results of measured values of pH, nitrogen components, COD, chlorides, surfactants LAS and ABS by methylene-blue assay, and total alkalinity for each of the filters. For each experimental period, as well as the entire period of the study, the mean value of the parameters are given along with the number of samples measured and the range, that is the high and low value of each parameter. The quality of the final effluents after the 26

inches of percolation can be compared with the quality of the feed solutions for each period described in Table 6-1.

For comparison with the calculated concentrations of the constituents in the feed solutions presented in Table 6-1, the measured values of the feed quality are given for two feeds, Day 51 and Day 338, in Table 6-10. Tables 6-11 and 6-12 show the quality of the percolate as a function of depth for each of the columns. These samples were taken over nearly a monthly interval. Each value represents the composite of a one-gallon feed solution passing through the depth

Table 6-3—Chemical quality of effluents from laboratory sand columns

Period: 0 to 71 days

				Concentrations in I	nilligrams per lite	er	
	pH value	NO ₂ —N	NO3—N	NH2—N	COD	Surfactant	Alk (as CaCOs)
Column A-1 Mean High Low Number of samples	6.2 7.6 5.4	0.14 0.65 0.00 14	19.71 26.00 13.00 7	22.89 71.60 10.50 22	20.5 28.2 12.7	4.3 5.8 0.8 24	28.4
Column A-2 Mean High Low Number of samples	6.7 7.4 6.1 7	0.11 0.41 0.00 14	18.43 31.00 0.00 7	49.43 61.00 20.10 21	26.1 34.8 16.7 15	5.6 8.8 0.5 25	13.6
Column L-1 Mean High Low Number of samples	6.0 7.5 4.9	0.12 1.05 0.01 14	22.00 25.00 18.00 7	19.30 24.80 10.50 22	14.2 21.7 7.3	1.0 1.8 0.4 25	5.6
Column L-2 Mean High Low Number of samples	6.3 7.4 5.4	0.32 0.94 0.02	19.38 23.00 12.00 8	50.37 64.00 12.50 22	27.0 32.8 18.3	2.5 4.5 0.8 26	24.4
Control Mean High Low Number of samples	6.3 7.5 5.8	0.15 1.04 0.01 13	17.29 21.00 11.00 7	16.37 27.60 2.80 21	12.0 19.7 6.2 9	0.2 0.2 0.1 22	14.8

Table 6-4—Chemical quality of effluents from laboratory sand columns

Period: 72 to 140 days

				Concentration	s in milligrams per	liter		
	pH value	NO:-N	NO ₈ —N	NH-N	COD	Cl-	Surfactant	Alk (as CaCO ₁)
Column A-1 Mean	5.6 7.1 4.2	.25 1.51 0.00 9	26.00 41.00 9.00	12.82 26.40 2.80	20.8 39.5 9.2	149.5 166 133 2	5.1 7.7 3.3	45 107 2.1 7
Column A-2 Mean High Low Number of samples	6.6 7.3 5.1	0.30 1.07 0.03	27.27 34.00 13.00	40.60 54.50 8.50	29.0 42.2 20.9	237.5 255 220 2	7.9 9.8 5.9	52 98.5 9.4 7
Column L-1 Mean High Low Number of samples	6.2 7.4 4.4	0.25 0.59 0.00 9	25.82 38.00 12.00	13.80 23.00 2.20 10	16.2 26.3 7.5	133 1	0.7 1.2 0.5 11	45.1 80.3 4.3 7
Column L-2 Mean High Low Number of samples	6.1 7.0 4.0	0.55 1.82 0.00 9	27,27 72,00 9,00 11	46.75 60.00 10.00 10	30.5 48.0 22.8	245.5 249 242 2	2.6 4.6 1.6	54.9 136 1.1 7
Control Mean High Low Number of samples	6.0 7.4 4.5	0.27 0.74 0.00 9	25.73 38.00 13.00	12.04 19.60 4.00	11.0 22.1 2.0 11	144 1	0.2 0.2 0.1 5	48 126 5.3 7

of sand filter indicated and then having the analyses run on the entire one-gallon composite.

6.08 Discussion of the Results.

This section will deal with a discussion of the performance of the filters during each of the experimental periods described in Section 6.06, followed by general statements on the overall comparative behavior of the two synthetic detergents relative to each other and to the control.

For the first 71 days of operation, Experiment I, the data presented in Tables 6-1, 6-3, and 6-10 show

that all the columns had effluents with mean value pH ranging from 6.2 to 6.7 after being dosed feed solutions with pH ranging from 7.2 to 7.6 final nitrite-nitrogen concentration of the effluent four of the filters was about the same with the heal loaded LAS column showing a slightly higher than the others. The nitrate-nitrogen concentration the column effluents are very nearly equal for all filters irrespective of the fact that columns A2 L2 received double the ammonia-nitrogen concentration in their feeds, compared to the other three lumns. Perhaps this measure of the degree of niting

Table 6-5—Chemical quality of effluents from laboratory sand columns

Period: 108 to 130 days

				Concentra	ations in milligram	s per liter		
	pH value	NO ₂ —N	NO ₈ —N	NH2—N	COD	CI-	Surfactant	Alk (as CaOO)
Column A-1 Mean	4.7 5.8 4.2 5	0.01 0.04 0.00 3	30.20 35.00 23.00 5	7.00 7.80 5.60	14.8 19.3 9.2 5	133	4.4 5.6 3.3 5	2.3 2.3 2.1 2
Column A-2 Mean. High Low. Number of samples.	6.3 7.0 5.1 5	0.12 0.14 0.09	24.60 34.00 19.00 5	40.88 49.00 32.00 4	27.3 30.0 20.9	220	7.6 8.2 7.2 5	22.5 36.4 9.4
Column L-1 Mean High Low Number of samples	5.8 6.8 4.4 5	0.02 0.05 0.00 3	28.60 32.00 21.00 5	9.60 10.00 8.80 4	11.5 14.2 7.5	133	0.6 0.6 0.5 5	7.4 10.4 4.3
Column L-2 Mean High Low Number of samples	5.6 6.9 4.0 5	0.03 0.05 0.00 3	25.80 37.00 16.00 5	45.88 52.00 40.50	24.3 26.2 22.8	242	2.2 2.8 1.8	1.6 2.1 1.1
Control Mean High Low Number of samples	5.2 6.4 4.5 5	0.02 0.05 0.00 3	29.40 34.00 21.00 5	7.60 11.80 5.60 4	6.9 10.6 2.0 5	144	=	5.4 6.4 5.3

Table 6-6—Chemical quality of effluents from laboratory sand columns

Period: 141 to 353 days

				Con	ncentrations in	milligrams per li	ter		
	pH Value	NO ₂ —N	NO3-N	NH ₂ —N	COD	Surfactant	Alk (as CaCO2)	TS	TVS
Column A-1 Mean	7.0 7.3 6.7	0.13 0.30 0.08 12	38.83 44.00 34.00 12	1.74 10.30 0.30 14	16.2 27.0 11.2	2.3 3.3 1.1 15	106 123 83 11	790 934 701 3	327 447 228
Column A-2 Mean High Low No. of samples	7.2 7.5 6.8	0.37 0.74 0.10 12	61.33 72.00 44.00 12	1.40 5.40 .15	20.9 31.0 17.2	4.3 8.4 1.4	134 150 111 11	1338 1383 1294 2	5\$4 606 562 2
Column L-1 Mean High Low No. of samples	7.3 7.5 7.0	0.23 0.81 0.08 12	36.92 40.00 34.00 12	1.28 6.60 0.30 14	14.1 25.9 8.6 13	0.5 0.8 0.3	110 118 100 11	939 1167 711 2	433 451 415 2
Column L-2 Mean High Low No. of samples	6.9 7.4 6.7	0.69 1.65 0.28	63.75 72.00 49.00 12	1.79 6.90 0.40 14	23.5 28.6 15.3	1.2 1.9 0.8	107 118 94 12	1145 1323 967 4	530 603 455 4
Control Mean High Low No. of samples	7.3 7.5 6.9	0.15 0.24 0.00 12	37.25 40.00 32.00 12	0.88 4.60 0.30 14	10.3 25.5 3.8 14	0.6 1.8 0.2	110 124 97 11	933 1137 693 4	407 427 392 4

tion being independent of the initial loading is not surprising. Each of the columns, the lightly loaded ones, as well as the heavily loaded, showed an appreciable ammonia-nitrogen concentration in the final effluent, that is, nitrification of the feed solutions was by no means complete for any of the columns. Since the columns were of identical construction and were operated the same hydraulically, one might conclude that the degree of nitification was limited by the amount of oxygen supplied to each column and the duration of the percolate within the column. All these factors were presumably the same for all five columns.

The COD data show about a 75-percent reduction for columns A2, L2, and the Control, whereas column A1 has only about a 66-percent removal and column L1 excels with about an 88-percent removal of COD. However, the main difference in the quality of the effluents is apparent with respect to the much greater removals of the surfactant LAS. On the average for the 71-day study, the lightly loaded LAS column showed an 80-percent removal while the heavily loaded filter showed a 60-percent removal. By comparison, the two columns fed ABS at the same feed concentrations as the LAS columns, had an average removal of

Table 6-7—Chemical quality of effluents from laboratory sand columns

				Cor	eentrations in	milligrams per li	ter		
	pH Value	NO2-N	NO ₄ —N	NH;—N	COD	Surfactant	Alk (as CaCO2)	TS	TVS
Column A-1 Mean High Low No. of samples	7.0 7.3 6.7	0.12 0.30 0.08 11	38.91 41.00 34.00 11	1.06 2.45 0.30	15.5 19.2 11.9	2.5 3.3 1.6 12	105 123 83 10	790 934 701 3	327 447 228 3
'olumn A-2 Mean High Low No, of samples	7.1 7.5 6.8	0.37 0.74 0.10	61.27 72.00 44.00	1.11 2.10 0.15 11	21.1 31.0 17.2	3.8 8.4 2.5	134 150 111 10	1338 1383 1294 2	584 606 562 2
olumn L-1 Mean. High Low. No. of samples.	7.3 7.5 7.0 11	0.23 0.81 0.08	36.63 39.00 34.00	0.93 2.20 0.30	12.8 18.1 8.6 10	0.6 0.8 0.3 12	109 118 100 10	939 1167 711 2	433 451 415 2
olumn L-2 Mean High Low No. of samples	6.9 7.4 6.8	0.71 1.65 0.28	63.55 72.00 49.00 11	1.11 2.45 0.40	22.1 26.1 15.3	1.2 1.7 0.8 12	108 118 94 10	1145 1323 967 4	530 603 455 4
ontrol Mean. High Low No. of samples.	7.3 7.5 6.9	0.14 0.24 0.00	37.18 40.00 32.00	0.56 1.15 0.30	9.2 15.4 3.8	0.3 0.3 0.2 3	109 124 97 10	933 1137 693	407 427 392 4

Table 6-8—Chemical quality of effluents from laboratory sand columns

Period: 290 to 353 days

				Concentrations in	milligrams per lite	r	
	pH value	NO2-N	NO:-N	NH ₄ —N	COD	Surfactant	Alk (as CaCO ₂)
Column A-1							
Mean	7.1	0.17	38.00	4,20	18.1	1.5	114
	7.2			10.30	27.0	2.1	114
High	7.0				21.0		
Low	7.0			0.45	11.2	1.1	
No. of samples	2	1	1	3	3	3	1
Column A-2							
Mean	7.2	0.42	62.00	2.47	18.8	1.8	136
High				5.30	22.5	2.1	100
Low				0.70	17.2	1.4	
1.0W				00	17.2	1.1	
No. of samples	2	1	1	3	3	3	1
'olumn L-1							
Mean	7.3	0.20	40.00	2.55	18.2	0.8	116
High		0.20	10.00	6.60	25.9	1.3	110
					13.1	0.5	
Low				0.45	13.1	0.5	
No. of samples	2	1	1	3	3	3	1
olumn L-2							
Mean	6.8	0.45	66.00	4.12	28.0	1.4	107
High	7.0	0.10	00.00	6,90	29.3	1.9	10.
Low	6.7			2.30	26.2	1.0	
No of complete	2			2.30	20.2	1.0	
No. of samples	2	1	1	3	3	3	1
ontrol							
Mean	7.3	0.14	38.00	2.07	14.1	1.8	120
High	7.4	0.11	55,50	4.60	25.5		120
Low	7.2			0.30	6.2		
No of complet	2.2			0.30	0.2		
No. of samples	ú	1	1	3	3	1	1

Table 6-9—Chemical quality of effluents from laboratory sand columns Period: 0 to 353 days

					Concentrati	ions in milligr	ams per liter			***************************************
	pH value	NO ₂ —N	NO ₂ —N	NH ₂ —N	COD	CI-	Surfactant	Alk (as CaCO ₃)	TS	TVS
Column A-1 Mean	6.3 7.6 4.2	0.16 1.51 0.00 35	29.66 44.00 9.00 30	14.26 71.60 0.30 46	19.2 39.5 9.2 37	147 166 133 3	3.9 7.7 0.8 50	101 123 2.1 18	790 934 701 3	22: 40: 224:
Column A-2 Mean. High Low. Number of samples.	6.9 7.5 5.1	0.25 1.07 0.00 35	38.83 72.00 0.00 30	32.52 61.00 0.15 45	25.2 42.2 16.7 39	200 255 126 3	5.7 9.8 0.5 50	98 150 94 19	1,338 1,383 1,294 2	564 606 562
Column L-1 Mean High Low Number of samples	6.6 7.5 4.4 31	0.19 1.05 0.00 35	29.37 40.00 12.00 30	12.62 24.80 0.30 46	14.7 26.3 7.3 39	148 162 133 2	0.8 1.8 0.3 51	81 118 53 19	939 1,167 711 2	433 451 415
Column L-2 Mean High Low Number of samples	6.5 7.4 4.0 32	0.50 1.82 0.00 36	39.36 72.00 9.00 31	34.36 64.00 0.40 46	26.8 48.0 15.3 39	233 249 209 3	2.1 4.6 0.8 52	85 141 1.1 20	1,145 1,323 967 4	530 603 455 4
Control Mean	6.6 7:5 4.5 31	0.18 1.04 0.00 34	28,37 40,00 11,00 30	10.51 27.60 0.30 45	11.0 25.5 2.0 34	118 144 93 2	0.3 1.8 0.1 31	82 126 5.3	933 1,137 693 4	407 427 392 4

Table 6-10—Measured quality of feed solutions to laboratory

	D	ay 51			
	A1	A2	L1	L2	Control
pH NO; — N NO; — N NH; — N COD Alkalinity (as CaCO ₃) TVS. TVS. Surfactant (not measured)	7.6 0.00 1.80 42.0 59.7 112 818 310	7.2 0.01 1.30 88.0 109.0 128 753 292	7.6 0.00 2.00 44.0 59.4 113 730 255	7.4 0.01 2.30 100.0 112.0 111 678 292	7.3 0.04 1.80 47.0 48.8 123

	Do	ay 338			
pH	7.9 0.00 38.0 7.70 54.9 119 3.3	7.6 0.06 70.0 13.60 98.0 125 5.9	7.9 0.01 36.0 8.15 61.2 118 6.2	7.7 0.05 68.0 13.30 114.0 125 11.4	7.9 0.00 38.0 5.15 47.0

* Concentrations in milligrams per liter.

23 percent for the lightly loaded column and 47 percent for the heavily loaded column. No reason is available for the relatively better removal of surfactant for the heavily loaded ABS column.

Figure 6-2 graphically shows the relative surfactant concentrations for each of the columns. The concentration of surfactant reported for the Control column which was fed no surfactant indicates the reliability of the methylene-blue assay as performed in our laboratory for detecting surfactant. Over the first 71 days, the mean of the surfactant measurements on the Control effluent was 0.2 mg per liter. From Figure 6-2, covering the initial 50 days of the study, one can see that the LAS columns became adapted to the removal of LAS after about 22 days of feeding. The ABS

columns showed very little removal of surfactant during the first 50 days.

Experiment II covers the time interval from Day 72 to Day 140 and the measured data on filter operation are summarized in Tables 6-1, 6-4, and 6-5. It was observed that there was a significant reduction in pH during the first 71 days of column operation. In order to control the pH, about 5 ml of 1 N sodium carbonate was added to each one gallon of feed solution. After addition of the sodium carbonate, the feed solutions had a pH of about 8.5. From Tables 6-4 and 6-5 it can be seen that the mean values of the pH of the column effluents ranged from 5.6 to 6.6 over the entire period and from 4.7 to 6.3 over the interval Day 108 to 130 (Experiment IIa) when the column effluents after the first pass through the filters were recycled back down the sand columns a second time in each 24-hour interval.

For the entire period of Experiment II all column effluents had about the same amount of nitrite-nitrogen and nitrate-nitrogen without regard to the initial ammonia-nitrogen concentration of the feed solutions. In the interval of recycling, the total amount of nitrate-nitrogen in the effluent after the second pass through the filter increased, but the final concentration was slightly less than double the concentration of the effluent after a single pass. For the interval Day 72 to 140, percent nitrification (based on the reduction of ammonia-nitrogen) averaged 65 to 70 percent for columns A1, L1, and the Control, while for A2 it was about 48 percent and L2, 41 percent. This was a slightly better performance than for the unbuffered, but younger, system in Experiment I.

The average reduction of COD during Experiment II was about 70 percent for columns A2, L1, L2, and the Control; but column A1 continued to show poorer performance with only about a 52-percent COD re-

moval. With respect to surfactant removal, the LAS columns excelled with a removal of 94 percent for L1 and 75 percent for L2, whereas column A1 showed nearly no removal (only 4 percent) of ABS while A2 indicated a 25-percent removal.

After 142 days of operating the sand filters under controlled conditions with the sterile feeds, the behavior of the columns indicated a much superior capability for removing LAS compared to ABS. With respect to the total amount of ammonia-nitrogen converted to nitrate, the surfactant columns performed quite similarly to the Control column. Also, with respect to COD reduction with the exception of column A1 all columns showed the same percent reduction

Table 6-11—Quality of effluents as function of depth (sampling period: days 205 to 232)

	Depth of percolation (inches)	рН	NO2—N (mg/l)	NO3—N (mg/l)	NH3—N (mg/l)	Alkalinity (mg CaCO ₄ /l)	Surfactant (mg/l)	COD (mg/l)	TS (mg/l)	TVS (mg/l)
Al	2 614 26	7.1 6.8 7.0	0.09 0.11 0.17 0.09	34.00 37.00 37.00 41.00	6.60 4.50 1.38 2.30	116 97 96	4.8 4.0 2.8 1.8	49.3 35.4 20.0	884 721	417 -415
A2	2 6 14 26	7.3 7.1 7.0	0.75 0.97 0.60 0.47	59.00 61.00 65.00 67.00	10.65 8.40 3.62 1.15	149 133 121	10.0 7.4 5.4 2.6	67.9 43.9 35.5 20.9	911 921	324 547
1	2 6 14 26	7.1 7.1 7.3	0.13 0.13 0.14 0.17	34.00 34.00 37.50 39.00	5.60 4.05 1.25 2.20	117 101 97	3.1 0.8 0.8 0.3	42.5 27.2 20.4 10.9	795 709	367 -404
,2	6	7.0 6.9 6.8	0.45 0.62 0.96 0.54	57.00 51.00 60.00 72.00	9.25 7.85 1.10 1.55	141 116 108 111	6.4 1.9 2.4 0.8	65.5 50.9 26.4 16.5	969 902	347 -527
Control	2 6 14 26	7.1 6.9 7.1	0.43 0.17 0.18 0.15	34.00 36.00 38.00 39.00	1.40 0.55 1.95 1.15	125 106 98	0.4 0.3 0.2 0.3	25.0 13.6 9.8 8.6	866 -548	406 268
eed*	A1 A2 L1 L2 Control			32.7 65.4 32.7 65.4 32.7	4.6 9.2 4.6 9.2 4.6		5.3 10.6 5.3 10.6	49.5 99.0 51.6 103.3 32.7		

^{*} Calculated from added ingredients.

Table 6-12—Quality of effluents as function of depth (sampling period: days 338 to 353)

	Depth of percolation (inches)	pН	NO ₂ —N (mg/l)	NO ₃ —N (mg/l)	NH ₃ —N (mg/l)	Alkalinity (mg CaCO ₂ /l)	Surfactant (mg/l)	COD (mg/l)
AI	0 2 6 14 26	7.9 7.4 7.1 7.0 7.2	0.00 0.15 0.39 0.15 0.17	38.00 38.00 35.00 45.00 38.00	7.70 2.00 0.40 0.85 0.45	119 130 121 110 114	3.3 1.9 1.7 1.4	54.9 52.1 21.2 14.7 11.2
A2	0	7.6 7.3 7.3 7.1 7.2	0.06 1.11 0.99 0.52 0.42	70.00 60.00 68.00 70.00 62.00	13.60 6.45 0.85 0.30 0.70	125 160 151 131 136	5.9 4.4 2.7 2.1 1.8	48.0 51.4 34.9 24.4 15.8
L1	0	7.9 7.4 7.3 7.2 7.3	0.01 0.51 0.21 0.11 0.20	38.00 35.00 39.00 41.00 40.00	8.15 5.30 4.00 0.00 0.45	118 133 124 115 116	6.2 0.5 0.6 0.5 0.5	61.2 35.7 27.8 21.7 13.1
.2	0 2 6 14 26	7.7 7.1 7.1 6.9 7.0	0.05 0.31 1.52 0.45 0.45	68.00 57.00 68.00 73.00 66.00	13.30 11.15 8.70 5.60 3.15	125 148 147 115 107	11.4 4.2 2.9 1.5	114.0 69.8 53.7 39.1 26.2
Control	0	7.9 7.4 7.2 7.2 7.4	0.00 0.69 0.35 0.13 0.14	38.00 35.00 35.00 43.00 38.00	5.15 0.30 0.15 3.00 0.30	119 134 130 118 120	=======================================	47.0 17.6 14.9 15.5 6.2
Feed*	A1			32.7 65.4 32.7 65.4 32.7	4.6 9.2 4.6 9.2 4.6		2.6 5.2 5.3 10.6	49.5 99.0 51.6 103.3 38.4

^{*} Calculated from added ingredients.

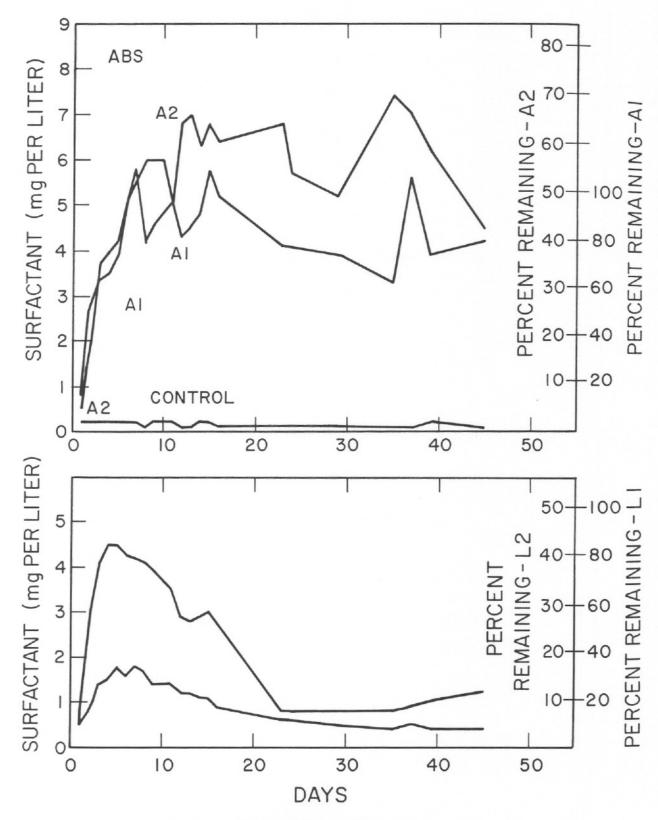


Fig. 6-2—ABS and LAS in Column Effluents as a Function of Time—Experiment 1.

after passage through 26 inches of sand. It is concluded that column A1 did not develop any of the proper organisms during the initial seeding of the filters to remove surfactant ABS from the feed solutions.

Experiment III was conducted to demonstrate the importance of nitrification in the reduction of pH of a buffered and unbuffered percolate. In Experiments I and II, a significant reduction of pH resulted during the passage of the feed solutions through the filter columns with the largest pH reduction occurring during Experiment IIa when the effluents from a single pass were recycled for a second pass through the columns in a 24-hour period. In Experiment III, the ammonium chloride in the feed solutions was replaced by sodium nitrate. The only nitrogen in the minus three oxidation state (ammonia) came from the nutrient broth, but this was small compared to the ammonia-nitrogen dose in the previous experiments.

The data presented in Tables 6-6, 6-7, and 6-8 show that the unbuffered feed solutions produced effluents with a pH in the range 6.1 to 7.3. In addition, alkalinities of the final effluents showed means from 104 to 134 mg/l (as $CaCO_3$) by comparison to alkalinities of effluents during Experiment IIa that were as low as 2.0 mg/l (as $CaCO_3$).

It is reasonable to suppose that the reduction in pH and drop in alkalinity during Experiments I and II resulted from the oxidation of the ammonia to nitrate. This biochemical reaction may be written as:

$$NH_{4^{+}} + 2O_{2} \rightarrow NO_{3^{-}} + 2H^{+} + H_{2}O$$
 (6-1)

The above equation shows a net production of free oxygen ions where 2 mols of hydrogen ions are produced for each mol of nitrogen (N) that is oxidized to nitrate. The alkalinity of the feed solutions is principally in the form of bicarbonate, so that the reduction of alkalinity can be expressed by

$$H^+ + HCO_3^- \rightarrow H_2CO_3$$
 (6-2)

Equations 6-1 and 6-2 indicate that each mol of ammonia-nitrogen oxidized to nitrate-nitrogen is capable of neutralizing 2 mols of bicarbonate alkalinity, that is, the oxidation of 14 mg of ammonia as N can reduce the alkalinity by 100 mg (as CaCO₃). These numbers compare very favorably with the alkalinity reductions and ammonia oxidation reported for Experiments I and II.

During the latter part of Experiment III, the filter columns were reseeded with a mixture of settled sewage and tap water. This reseeding was done primarily to see if an improvement in the removal of ABS would result. Because of an error in the preparation of the ABS feed solutions and the short duration of the column experiments after the seeding, no definite conclusions from this rejuvenation are merited

Table 6-6 summarizes the data for the entire period of Experiment III. The nitrite-nitrogen concentrations

average higher than any of the previous experiments suggesting that some of the nitrate-nitrogen may have been reduced. Enough information is not available to elaborate on this point other than it may well be possible to control the oxidation-reduction potential of the percolate in the sand columns by attempting to supply oxygen to the micro-organisms in the combined form as nitrate. More will be made of this point in the final chapter of this report.

Using the measured values of the ammonia-nitrogen reported in Table 6-10 (Day 338) for the influents and the mean values of the effluents in Table 6-6, the columns show an 80- to 90-percent reduction in ammonia-nitrogen. Columns A1, L1, and the Control have about the same amount of nitrate-nitrogen in the influent and effluents, while columns A2 and L2 have mean values in the effluents slightly less than influents. However, the differences are really not significant.

With respect to COD reduction, on a calculated basis using the influent concentrations from Table 6-1, the columns show 70- to 80-percent reduction. The more heavily loaded surfactant columns, A2 and L2, show the better removal.

6.09 Results of the Full Study.

The average of measured parameters describing effluent quality are summarized in Table 6-9. With respect to oxidation of ammonia-nitrogen and overall removal of COD, the columns receiving the surfactants behaved similarly to the column receiving no surfactant. With respect to surfactant, it is quite apparent that LAS is much more readily removed than ABS. To demonstrate the relative removal on a quantitative basis a material balance was calculated for the four surfactant columns.

Table 6-13 is a summary of the analyses of sand cores that form each of the sand sections making up the filter columns. For each analysis a composite sample was made by using a thin-walled brass tube (½-inch I.D., 3 inches long) as a coring device. Several cores were taken from the top of each sand section. Hence, the entire depth of sand was sampled for the 2-inch sand beds, nearly the full depth for the 4-inch beds, and only the upper part for the 8 and 12-inch sections.

To make a material balance of the surfactants, it is necessary to know the total amount of surfactant applied to each column, the relative concentrations of the influent and effluent, and the amount of surfactant adsorbed to the sand and organic matter composing the filter media. Using the data in Tables 6-1 and 6-2 for the full 353 days of the study, one can calculate that columns A1, L2, and the Control were given 350 doses and columns A2 and L1, 349 doses of their respective feed solutions. These numbers result from the failure to load the columns on several days, the addition of settled sewage for five days, and a multiple loading of the columns several

times for sampling purposes. From the number of doses to each column, and the amount of surfactant in each dose, one can calculate the total load over 353 days.

Total Weight of Surfactant Added (gm)

 L_2 A2 L_1 AI 14.00 7.986.83 13.62

Taking the data from section 6.03 for the weight of sand in each filter bed and the apparent weight of surfactant (methylene-blue assay) desorbed from the sand beds after the 353 days of column operation, the following calculations result:

Table 6-13—Results of analyses of sand cores taken from the laboratory filter at the end of the study

Sand section	Moisture content	Volatile solids	Organic N +NH3-N	COD	Surfactant
A1 2 4 4 8 12 12 12 12 12 12 12 12 13 14 15 15 16 16 17 17 18 17 18 17 18 18 18 18 18 18 18 18 18 18 18 18 18	0.085	5.6	0.21	1.8	5.2
	0.056	5.1	0.20	2.2	9.3
	0.024	3.7	0.12	0.86	2.4
	0.024	3.8	0.034	0.48	1.7
A2 2	0.13	9.1	0.32	4.4	25.8
	0.073	7.1	0.52	4.0	5.1
	0.042	5.5	0.13	2.0	6.5
	0.032	4.9	0.092	1.1	5.3
L1 2 4	0.30	13.0	0.21	1.8	5.4
	0.061	5.1	0.16	1.4	2.2
	0.024	4.1	0.073	0.92	1.6
	0.024	3.3	0.042	0.80	0.77
L2 2	0.11 0.071 0.025	6.6 6.8 4.2 4.1	0.31 0.34 0.15 0.045	3.7 3.5 1.9 0.63	9.8 5.2 2.5 0.80
Control 2 4	0.15	7.3	0.25	2.8	5.0
	0.050	5.1	0.13	1.1	1.7
	0.029	4.4	0.079	0.86	1.0
	0.027	4.0	0.035	0.55	1.2
New sand		3.0		0.14	0.80

Total Weight of Adsorbed Surfactant (mg)

Sand						
Section		A1	A2	L1	L2	Control
2 in.		3.2	16.0	3.4	6.1	- estima
4 in.		10.6	6.4	2.8	6.5	3.1 2.1 1.8
			11.4	2.8	4.5	2.1
12 in.		6.3	19.7	2.9	3.0	4.5
T	otal	24.3	53.5	11.9	20.1	11.5

Using a sample of clean sand from the supply used to pack the columns, an extraction for surfactant performed. This clean sand had an apparent surface tant content of 0.80 mg ABS per kg dry soil, so that for a column containing 7.35 kg of sand the back ground apparent surfactant is about 5.9 mg. The num bers for the total weight of adsorbed surfactant for each of the columns given above should have 5.9 mg subtracted from them in order to arrive at the true apparent amounts of adsorbed surfactant.

Corrected Weights of Adsorbed Surfactant (mg)

AI	A2	L1	L2	Control	
18.4	47.6	6.0	14.2	5.6	

The weight of adsorbed surfactant is practically nerligible compared to the total amount of surfactant applied to the columns. Therefore the behavior of the columns with respect to removal and degradation is adequately described by the changes in the concentration between the influent and effluent concentrations since the quantity adsorbed to the sand is so

As shown in Table 6-9, the overall performance of the columns resulted in a removal of 26 percent of the ABS from the lightly loaded filter and 46 percent of the ABS from the more heavily loaded filter. For the columns receiving LAS, the lightly loaded filter showed an 85 percent removal while the heavier dosed column removed 80 percent. Table 6-6 shows that the percentage removals were higher in the last days of the study, being about 57 to 60 percent for the ABS columns and reaching 89 to 91 percent for the LAS columns.

Weight loss at 105°C., gm. water per gm. dry sand.
 Weight loss at 600°C., gm. volatile solids per kg. dry sand.
 Organic nitrogen + ammonia-nitrogen, gm. N per kg. dry sand.
 Chemical oxygen demand, gm. COD per kg. dry sand.
 Surfactant by methylene-blue assay, mg. surfactant per kg. dry soil.

REMOVAL AND BIODEGRADATION OF ABS AND LAS IN SOIL SYSTEMS (FIELD AND LABORATORY)

7.01 The Problem.

In the light of the decisions by the detergent manufacturers to convert production from the biologically resistant alkyl benzene sulfonates (ABS) to the biologically "soft" linear alkylate sulfonates (LAS) by the end of 1965, some of the problems discussed in this chapter may seem passé and academic. Nevertheless, the knowledge gained by this detailed investigation of ABS will assist in a better understanding of the principles and mechanisms involved in the adsorption and degradation of other exotic organic compounds. Although ABS may soon be of historical interest only, it has served to alert water-quality-control engineers and scientists to problems associated with all types of biologically resistant compounds.

This chapter describes the results of field investigations relative to ABS at the Whittier Narrows and Rio Hondo Test Basins. In addition, the reader is referred to Chapter 6, where the removal of ABS and LAS in laboratory sand columns is discussed.

Prior to the inception of the investigations described in this report, it was evident that the surface-active substances categorized as alkyl benzene sulfonates (ABS) would be critical parameters of pollution. It was recognized that ABS occurred in well-stabilized wastewater effluents in the general magnitude of 3 to 5 mg/l, that the USPHS Drinking Water Standards of 1962 placed on ABS a recommended (i.e., nonmandatory) limit of 0.5 mg/l, and that ABS persisted in some ground waters interminably, traveling many miles with no apparent diminution in concentration or foaming potential. For these reasons, ABS was likely to be the most-critical constituent of treated wastewater with respect to the quality of ground water downstream from the spreading basins.

The fact that ABS had been reported to persist in some ground-water regimes for months and years did not mean necessarily that ABS would be a problem in the Whittier Narrows Water Reclamation project. Indeed, there was some evidence and considerable hope that ABS could be removed effectively by soil under optimum conditions during spreading operations. It was advisable, however, to assess the factors controlling ABS removal and to determine the optimum conditions for the Whittier Narrows project.

Specifically, it was the intent of this phase of the investigation to attempt to answer the following questions:

a. Will the ABS that has survived treatment by the activated-sludge process and foam fractionation be removed during percolation through the upper 2 to 8 feet of soil in conventional spreading operations?

b. Will this removal be permanent, or will a chromatographic effect occur, whereby the ABS will be desorbed and leached to greater depths or longer distances?

c. Is the removal achieved by physical adsorption alone or does biochemical degradation occur, either initially or subsequently?

d. What rate mechanisms are involved and how are they affected by methods of basin operation, soil characteristics, temperature, and other natural variables?

e. If biodegradation does take place, what organisms are involved and what are the pathways of decomposition?

f. How can a spreading basin best be operated to achieve maximum permanent removal of ABS and to prevent ABS pollution of deep or distant ground waters?

These are important but perplexing questions. Quite obviously, they cannot all be answered on the basis of this investigation. Nevertheless, the data gathered to date provide some very interesting opportunities for hypotheses and lead to some provocative observations. Before reporting these results, however, it is well to reflect on some of the work done by others in relation to this problem.

7.02 A Critical Review of Current Knowledge on the Behavior of ABS and LAS in Ground Water.

Literature on the occurrence and behavior of ABS in raw wastewaters, during treatment processes, and in surface and subsurface waters is prodigious. An exhaustive review of such literature is far beyond the scope of this report. It should be pointed out, however, that a comprehensive bibliography on synthetic detergents in water and wastes (5) was printed in October 1962. Furthermore, summaries of the general problem have been presented by Ward (6), the Soap and Detergent Association (7), Coughlin, et al. (8), Walton (9), Patton (10) and others. With respect to the fate of ABS in ground water, however, the literature is not voluminous and certainly not very definitive.

There are, to be sure, frequent references to the detection of ABS in ground-water supplies, with computations or conjecture as to the origin of the ABS, the distance traveled, and the elapsed time or duration of ABS persistence. It has been reported, for example, that ABS has traveled two or more miles underground, that it has penetrated to depths of 100 feet or more, and that it has persisted in ground water for three years or longer. Literature describing these situations has been cited by Coughlin, et al. (8), Walton (9), Patton (10), Ewing, et al. (11), a survey by LA Section ASCE (18), a report to the California State Legislature (19), and others.

Published reports on fundamental studies dealing with the mechanisms and rates of ABS removal by soil are relatively sparse. Within the United States, they are confined largely to the work done by the USPHS at Cincinnati (12, 13, 14), by Klein et al. at

Berkeley (15, 16), and by Ewing et al. in Illinois

(11).

The Cincinnati group has demonstrated rather convincingly that the ABS in septic-tank effluent can be removed from levels of 5-35 mg/l down to less than 0.5 mg/l by intermittent aerobic filtration through unsaturated soils. By using ABS tagged with sulfur-35, they found that there was apparent degradation as well as adsorption of the ABS, i.e., the tagged sulfur appeared as sulfate in the effluent. Soil columns were dosed once a day with pre-selected volumes of anaerobic septic-tank effluent at the following rates and loadings:

Number of Columns Used	Hydraulic Loading m/day	BOD $Loading$ $g/m^2/day^*$	ABS Loading g/m²/day*	Duration of Ponding, Minutes†
4	0.122	16.35	1.34	10-24
5	0.203 0.406	$27.2 \\ 54.4$	$\frac{2.24}{4.47}$	30-46 44

Calculated at 134 mg/l of BOD and 11 mg/l of ABS.
 † After one month of sewage dosing.

In practice, conventional intermittent sand filters are dosed at hydraulic rates of 0.15 m/day or less, and at BOD loadings of 20 to 40 g/m²/day. The high rates of loading for these USPHS tests, and the short duration of ponding, were made possible by the fact that Ottawa silica sand was used as the test soil. The effective sizes varied from 0.1 to 0.3 mm in various columns, with uniformity coefficients of about 1.7 in all cases.

The ABS removal was high (80-90 percent) at the start of the test, dropped to as low as 20 percent at the 0.3 meter depth of percolation after 10 to 15 days of operation, and then recovered to 70-80 percent after 2 to 5 months. The initial high rate of removal was attributed to adsorption by the sand itself. The poor removal at 10 to 15 days was undoubtedly due to the saturation of the adsorptive capacity of the sand prior to the development of adequate biological growths. The improved removal in the period from one to five months of operation resulted from a buildup of organisms capable of adsorbing and degrading ABS. The variation in effective size of the sand had no apparent influence on removal efficiencies; nor did the differences in hydraulic or organic loadings. The deeper filters gave better total removal of ABS, but most of the removal came in the first meter of depth.

It is essential to note that these USPHS experiments were performed under highly aerobic conditions. Although the septic-tank effluent used for dosing was anaerobic, the columns were ponded with effluent for brief periods only, and they were exposed to the open air for more than 23 hours out of each day. Since they drained freely, there was ample opportunity for air (and oxygen) to penetrate downward from the surface through the soil interstices.

The Berkeley studies were considerably different and cannot be compared directly with the Cincinnati experiments, as explained below. The first annual report of the Berkeley investigations (15) describes tests with sterile saturated soil columns and with biologically seeded columns, both saturated and unsaturated. All columns during all tests received a

sterilized feed solution containing BOD dilution water plus nutrient broth to give 20 mg/l of BOD. The feed solution also contained 1.0 mg/l of ABS and a very small amount of ABS tagged with sulfur-35.

The sterile saturated columns (indicative of ground, water flow in confined aquifers) revealed that the adsorptive capacity of the five soils tested (based on desorption from the soil at the end of the test) ranged from 1.64 mg/kg for a sandy soil (Oakley) to 13.0 mg/kg for a fine sandy loam (Hanford). All soils exhibited an early breakthrough of ABS, the more permeable and least adsorptive soils requiring the least period of time.

For the biologically active but saturated columns. Oakley sand was used. Prior to the test, the columns. were deaerated with CO2 gas, saturated with water, and then dosed with water containing one or ten percent sewage. The saturated anaerobic seeded columns were allowed to rest for 16 to 64 hours before receiving sterile feed solution with ABS. There was no opportunity for aerobic cultures to develop. In the first test of 48 hours duration (but only 6 hours of actual percolation time) the average hydraulic load. ing was 3.45 m/day based on the full 48-hour period or 27.5 m/day during the actual time of filtration. These are very rapid rates for most ground-water movement in confined aquifers. A second test with a heavier pre-dose of seed material ran steadily for 52 hours at a hydraulic loading of 6.47 m/day. Neither of these tests gave any indication that ABS adsorption was increased by the presence of bacteria in an anaerobic environment, nor was there any evidence of ABS degradation.

A biologically active unsaturated column (B) was compared with a similar saturated one (A), both using Oakley sand and identical sterile feed solutions. The saturated column was flushed with CO2 and soaked overnight with a 10-percent sewage mixture containing nutrient broth equivalent to 20 mg/l of BOD. The unsaturated column was similarly soaked with dilute sewage but drained immediately to allow air to be entrapped in the soil. Column A was operated for 20 days with an actual percolation time of 53 hours. The hydraulic loading, based on 20 days, was 1.31 m/day. Column B was operated for 50 days with only 43 hours of percolation time and 1151.5 hours of holding or resting time. The hydraulic loading based on 50 days was 0.46 m/day, a rate comparable to the highest one used by the USPHS but much higher than that employed in practice. The corresponding BOD loading was 9.19 g/m²/day, and the ABS loading was only 0.459 g/m²/day. These organic parameters are much lower than the USPHS loadings reported above. No evidence of biodegradation of ABS was found in saturated column A, based on sulfur-35 assay; but in column B some biodegradation of adsorbed ABS occurred during the holding or resting periods when the column was drained to entrain more air. The percentage of ABS biodegraded, however, was infinitesimal compared with the total amount dosed onto or drained from the column. Most degradation appeared to be confined to the upper 2 or 3 cm of soil.