Staff Report of the

CALIFORNIA ENVIRONMENTAL PROTECTION AGENCY

REGIONAL WATER QUALITY CONTROL BOARD CENTRAL VALLEY REGION

# METAL CONCENTRATIONS, LOADS, AND TOXICITY ASSESSMENT IN THE SACRAMENTO/SAN JOAQUIN DELTA: 1993-1995

December 1998

53.1

REPORT COMPILED BY: Stephen L. Clark Student Intern

Contributors: CVRWQCB Staff: Jerry Bruns, Valerie Connor, Janice Cooke, Bill Croyle, Chris Foe, Michelle McGraw, Shelly Morford, and Sue Yee State of California

California Environmental Protection Agency

# REGIONAL WATER QUALITY CONTROL BOARD CENTRAL VALLEY REGION

Edward Schnabel, Chair Steven Butler, Vice Chair Chuck Ahlem, Member Jane Papazian, Member Craig Pedersen, Member William Porter, Member

Gary M. Carlton, Executive Officer

3443 Routier Road, Suite A Sacramento, California 95827-3003

> Phone: (916) 255-3000 CalNet: 8-494-3000

# DISCLAIMER

This publication is a technical report by staff of the California Regional Water Quality Control Board, Central Valley Region. No policy or regulation is either expressed or intended.

# Forward

This project has been funded by the Bay Protection Toxic Clean-up Program and by the Central Valley Regional Water Quality Control Board under contract number FG 2305 ES with the California Department of Fish and Game to conduct a survey of metals in the Sacramento/San Joaquin Delta Estuary. This document was prepared through agreement number 2-088-250 with the California State Water Resources Control Board. The contents of this document do not necessarily reflect the views and policies of the California Department of Fish and Game nor of the State Water Resources Control Board nor does mention of trade names or commercial products constitute endorsement or recommendation for use.

ii

# TABLE OF CONTENTS

TABLE OF CONTENTS	iii
LIST OF TABLES	iv
LIST OF FIGURES	vii
LIST OF APPENDICES	X
EXECUTIVE SUMMARY	xi
INTRODUCTION	1
Basin Description	1
Sources of Metals	1
Metal Toxicity	2
Water Quality Criteria/Objectives	3
Bay Protection Toxic Cleanup Program	4
MATERIALS AND METHODS	6
Quality Assurance Program	6
Sample Locations	6
Sample Collection and Storage	6
Metal Analyses	6
Toxicity Samples	7
Were Low Detection Limits Obtained Using Ultra -Clean Techniques?	7
AA Methods (Trace Metal Lab)	7
AA Methods (Mussel Watch Lab)	7
Were Water Quality Objectives Exceeded?	8
Was Metals Related Toxicity Identified in the Delta?	8
Statistical methods and definitions of toxicity	8
Were Metal Loading Patterns Characteristic of Hydrological Conditions?	9
Water Years 1993, 1994, and 1995	9
Flow Rates	9
Load Calculations	9
What Source(s) of Metals Were Identified During Metals Source Pilot Study?	10
RESULTS AND DISCUSSION	12
Quality Assurance/Quality Control	12
Hydrological Conditions	12
Were Low Detection Limits Obtained Using Ultra -Clean Techniques?	12
Were Water Quality Objectives Exceeded?	12
Was Metals Related Toxicity Identified in the Delta?	13
Were Metal Loading Patterns Characteristic of Hydrological Conditions?	15
What Trends in Metal Concentrations Were Identified?	17
What Source(s) of Metals Were Identified During Metals Source Pilot Study?	22
ACKNOWLEDGMENTS.	23
SUMMARY OF RECOMMENDATIONS	24
LITERATURE CITED	25
APPENDICES	216

# LIST OF TABLES

**?** 

Table	1.	Sites and dates of sampling in the Delta and Lower Sacramento River Basin47
Table	2.	Analytical information for four programs monitoring metals in the Sacramento
		River Watershed
Table	3.	Summary of Water Year 1993-1994 metal concentration data and related water
		quality objectives from the San Joaquin River at Antioch
Table	4.	Summary of Water Year 1994-1995 metal concentration data and related water
		quality objectives at Duck Slough
Table	5.	Summary of Water Year 1994 metal concentration data and related water
		quality objectives from French Camp Slough
Table	6.	Summary of Water Year 1993-1994 metal concentration data and related water
		quality objectives from the Sacramento River at Hood
Table	7.	Summary of Water Year 1993-1994 metal concentration data and related water
		quality objectives from Middle River at Bullfrog Landing
Table	8.	Summary of Water Year 1993-1995 metal concentration data and related water
		quality objectives from the Mokelumne River
Table	9.	Summary of Water Year 1994 metal concentration data and related water
		quality objectives from the Old River at Tracy Blvd
Table	10	. Summary of Water Year 1994 metal concentration data and related water
		quality objectives from Paradise Cut
Table	11.	. Summary of Water Year 1994-19945 metal concentration data and related
		water quality objectives from Prospect Slough
Table	12.	. Summary of Water Year 1993-1994 metal concentration data and related water
		quality objectives from the Sacramento River at Rio Vista
Table	13.	. Summary of Water Year 1995 metal concentration data and related water
		quality objectives from Skag Slough74
Table	14.	. Summary of Water Year 1994 metal concentration data and related water
		quality objectives from the San Joaquin River at Stockton
Table	15.	. Summary of Water Years 1994-1995 metal concentration data and related water
		quality objectives at Ulatis Creek
Table	16.	. Summary of Water Year 1993-1995 metal concentration data and related water
		quality objectives from the San Joaquin River at Vernalis
Table	17.	. Summary of Water Year 1995 metal concentration data and related water
		quality objectives from the Sacramento River at Greene's Landing
Table	18.	. Number of Dissolved (0.45 $\mu$ m) metal analyses and exceedances of water
		quality objectives during water years 1993-1995
Table	19.	. Summary of 1993-1994 toxicity monitoring data
Table	20.	. Summary of 1994-1995 toxicity monitoring data
Table	21	. Summary of Dissolved (0.45 $\mu$ m) metal analyses from 1993 to 1995 with
		notes on levels of concern in the literature
Table	22	. Summary of lead concentrations reported to have effect on algae and diatoms92
Table	23	. Summary of lead concentrations reported to have effect on invertebrates

iv

Table	24. Summary of lead concentrations reported to have effect on fish	
Table	25. Summary of arsenic concentrations reported to have effect on algae	
Table	26. Summary of arsenic concentrations reported to have effect on invertebrates96	
Table	27. Summary of arsenic concentrations reported to have effect on fish	
Table	28. Summary of chromium concentrations reported to have effect on algae and	
	diatoms	
Table	29. Summary of chromium concentrations reported to have effect on	
	invertebrates	
Table	30. Summary cf chromium concentrations reported to have effect on fish100	)
Table	31. Summary of nickel concentrations reported to have effect on algae and	
	diatoms101	l
Table	32. Summary of nickel concentrations reported to have effect on invertebrates102	2
Table	33. Summary of nickel concentrations reported to have effect on fish103	3
Table	34. Summary of copper concentrations reported to have effect on fish104	1
Table	35. Summary of copper concentrations reported to have effect on invertebrates105	5
Table	36. Summary of copper concentrations reported to have effect on algae100	5
Table	37. Summary of zinc concentrations reported to have effect on fish107	7
Table	38. Summary of zinc concentrations reported to have effect on invertebrates108	3
Table	39. Summary of zinc concentrations reported to have effect on algae109	)
Table	40. Summary of cadmium concentrations reported to have effect on fish110	)
Table	41. Summary of cadmium concentrations reported to have effect on invertebrates112	l
Table	42. Summary of cadmium concentrations reported to have effect on algae112	2
Table	43. Comparison of Metal Load Estimates in the Sacramento River at Greene's	
	Landing from January through April during a Dry Year (1994) and a Wet Year	
	(1995)111	3
Table	44. Comparison of Metal Load Estimates in the Sacramento River at River Mile	
	44 from January through April during a Dry Year (1994) and a Wet Year	
	(1995)114	4
Table	45. Comparison of Metal Loads to the Delta Contributed by Sources which	
	Drain to the Yolo Bypass and Sacramento River, January-April 199511:	5
Table	46. Total recoverable and dissolved (0.45 $\mu$ m) metal concentrations in samples	
	collected from all stations monitored in 1993, 1994 and 1995110	5
Table	47. Total recoverable and dissolved (0.45 $\mu$ m) metal concentrations in samples	
	collected at Greene's Landing from January through March of 1993, 1994	
	and 199511	7
Table	48. BPTCP: Summary of regression coefficients for total recoverable and	_
	dissolved (0.45 μm) metals, flow, and TSS during Water Year 199411	8
Table	49. BPTCP: Summary of regression coefficients for total recoverable and	_
	dissolved (0.45 $\mu$ m) metals, flow, and TSS during Water Year 199511	9
Table	50. BPTCP: Summary of regression coefficients for total recoverable and	
	dissolved (0.45 $\mu$ m) metals, flow, and TSS during Water Years 1994 and	~
	199512	0

,		
Tabl	le 51. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved (0.45 μm) metals, flow, and TSS during Water Year 1994	
Tabl	le 52. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved (0.45 μm) metals, flow, and TSS during Water Year 1995	۶
Tabl	le 53. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved (0.45 μm) metals, flow, and TSS during Water Years 1994-1995	ĩ
Tabl	<ul> <li>le 54. BPTCP: Summary of total recoverable metals regressed against other metals for samples collected from Greene's Landing during WY94, WY95, and the</li> <li>combined WY94/95124</li> </ul>	

vi

.

.

.

,

. .

**a** ',

# LIST OF FIGURES

Figure	1. Map of the Sacramento-San Joaquin River Delta and its major tributaries	.126
Figure	2. Sacramento River Flow at Freeport from January 1993 to September 1995	.127
Figure	3. Total Recoverable Copper vs. Flow during WY94	.128
Figure	4. Total Recoverable Zinc vs. Flow during WY94	.129
Figure	5. Total Recoverable Chromium vs. Flow during WY94	.130
Figure	6. Total Recoverable Lead vs. Flow during WY94	.131
Figure	7. Total Recoverable Nickel vs. Flow during WY94	132
Figure	8. Total Recoverable Copper vs. TSS during WY94	.133
Figure	9. Total Recoverable Zinc vs. TSS during WY94	134
Figure	10. Total Recoverable Chromium vs. TSS during WY94	135
Figure	11. Total Recoverable Lead vs. TSS during WY94	136
Figure	12. Total Recoverable Nickel vs. TSS during WY94	.137
Figure	13. Dissolved (0.45 μm) Copper vs. Flow during WY94	.138
Figure	14. Dissolved (0.45 μm) Chromium vs. Flow during WY94	.139
Figure	15. Dissolved (0.45 μm) Nickel vs. Flow during WY94	.140
Figure	16. Dissolved (0.45 μm) Copper vs. TSS during WY94	141
Figure	17. Dissolved (0.45 μm) Chromium vs. TSS during WY94	142
Figure	18. Dissolved (0.45 μm) Nickel vs. TSS during WY94	.143
Figure	19. Chromium, Total Recoverable vs. Dissolved (0.45 µm) during WY94	.144
Figure	20. Lead, Total Recoverable vs. Dissolved (0.45 µm) during WY94	.145
Figure	21. Total Recoverable Copper vs. Flow during WY95	.146
Figure	22. Dissolved (0.45 μm) Copper vs. Flow during WY95	.147
Figure	23. Total Recoverable Zinc vs. Flow during WY95	148
Figure	24. Dissolved (0.45 µm) Zinc vs. Flow during WY95	149
Figure	25. Total Recoverable Chromium vs. Flow during WY95	150
Figure	26. Dissolved (0.45 μm) Chromium vs. Flow during WY95	.151
Figure	27. Total Recoverable Lead vs. Flow during WY95	.1.52
Figure	28. Dissolved (0.45 μm) Lead vs. Flow during WY95	.153
Figure	29. Total Recoverable Cadmium vs. Flow during WY95	.154
Figure	30. Dissolved (0.45 µm) Cadmium vs. Flow during WY95	.155
Figure	31. Total Recoverable Nickel vs. Flow during WY95	.156
Figure	32. Dissolved (0.45 μm) Nickel vs. Flow during WY95	.157
Figure	33. Total Recoverable Arsenic vs. Flow during WY95	.158
Figure	34. Dissolved (0.45 µm) Arsenic vs. Flow during WY95	.159
Figure	35. Total Recoverable Copper vs. TSS during WY95	.160
Figure	36. Total Recoverable Zinc vs. TSS during WY95	.161
Figure	37. Total Recoverable Cadmium vs. TSS during WY95	.162
Figure	38. Total Recoverable Cadmium vs. Flow during WY94	.163
Figure	39. Copper, Total Recoverable vs. Dissolved (0.45 μm) during WY95	.164
Figure	40. Lead, Total Recoverable vs. Dissolved (0.45 μm) during WY95	.165
Figure	41. Total Recoverable Copper vs. TSS, WY94-WY95	.166

Figure	42.	Total Recoverable Zinc vs. TSS, WY94-WY95	167
Figure	43.	Total Recoverable Chromium vs. TSS, WY94-WY95	168
Figure	44.	Total Recoverable Nickel vs. TSS, WY94-WY95	169
Figure	45.	Total Recoverable Copper vs. Flow, WY94-WY951	170
Figure	46.	Total Recoverable Zinc vs. Flow, WY94-WY95	171
Figure	47.	Total Recoverable Chromium vs. Flow, WY94-WY95	172
Figure	48.	Total Recoverable Nickel vs. Flow, WY94-WY95	173
Figure	49.	Dissolved (0.45 µm) Chromium vs. TSS, WY94-WY95	174
Figure	50.	Dissolved (0.45 µm) Lead vs. TSS, WY94-WY95	175
Figure	51.	Dissolved (0.45 µm) Nickel vs. TSS, WY94-WY95	176
Figure	52.	Dissolved (0.45 µm) Chromium vs. Flow, WY94-WY95	177
Figure	53.	Dissolved (0.45 µm) Lead vs. Flow, WY94-WY95	178
Figure	54.	Dissolved (0.45 µm) Nickel vs. Flow, WY94-WY95	179
Figure	55.	Total Zinc vs. Total Copper, WY94	180
Figure	56.	Total Chromium vs. Total Copper, WY94	181
Figure	57.	Total Lead vs. Total Copper, WY94	182
Figure	58.	Total Nickel vs. Total Copper, WY941	183
Figure	59.	Total Chromium vs. Total Zinc, WY94	184
Figure	60.	Total Lead vs. Total Zinc, WY94	185
Figure	61.	Total Nickel vs. Total Zinc, WY94	186
Figure	62.	Total Lead vs. Total Chromium, WY941	187
Figure	63.	Total Nickel vs. Total Zinc, WY94	188
Figure	64.	Total Nickel vs. Total Lead, WY94	189
Figure	65.	Flow vs. TSS, WY94	190
Figure	gure 66. Flow and TSS Pattern in the Sacramento River at Greene's Landing from		
		January through March 19941	91
Figure	67.	Total Zinc vs. Total Copper, WY951	192
Figure	68.	Total Chromium vs. Total Copper, WY95	193
Figure	69.	Total Cadmium vs. Total Copper, WY95	194
Figure	70.	Total Nickel vs. Total Copper, WY951	195
Figure	71.	Total Chromium vs. Total Zinc, WY95	96
Figure	72.	Total Cadmium vs. Total Zinc, WY951	97
Figure	73.	Total Nickel vs. Total Zinc, WY95	198
Figure	74.	Total Cadmium vs. Total Chromium, WY95	99
Figure	75.	Total Nickel vs. Total Chromium, WY95	200
Figure	76.	Flow vs. TSS, WY95	201
Figure	77.	Precipitation and TSS Pattern at Greene's Landing from January Through	
		Mid February 1995	202
Figure	78.	Flow and TSS Pattern from January to March 1995	203
Figure	79.	Flow vs. TSS, WY95 Without Pre- and First Flush Values	204
Figure	80.	Total Zinc vs. Total Copper, WY94/WY95	205
Figure	81.	Total Chromium vs. Total Copper, WY94/WY95	206
Figure	82.	Total Lead vs. Total Copper, WY94/WY95	207

ବ

viii

83. Total Nickel vs. Total Copper, WY94/WY95	
84. Total Chromium vs. Total Zinc, WY94/WY95	
85. Total Lead vs. Total Zinc, WY94/WY95	
86. Total Nickel vs. Total Zinc, WY94/WY95	
87. Total Lead vs. Total Chromium, WY94/WY95	
88. Total Nickel vs. Total Chromium, WY94/WY95	
89. Total Nickel vs. Total Lead, WY94/WY95	
90. Flow vs. TSS, WY94/WY95	
	<ol> <li>83. Total Nickel vs. Total Copper, WY94/WY95</li> <li>84. Total Chromium vs. Total Zinc, WY94/WY95</li> <li>85. Total Lead vs. Total Zinc, WY94/WY95</li> <li>86. Total Nickel vs. Total Zinc, WY94/WY95</li> <li>87. Total Lead vs. Total Chromium, WY94/WY95</li> <li>88. Total Nickel vs. Total Chromium, WY94/WY95</li> <li>89. Total Nickel vs. Total Lead, WY94/WY95</li> <li>90. Flow vs. TSS, WY94/WY95</li> </ol>

# LIST OF APPENDICES

Appendix A:	Description of Site Locations	.216
Appendix B:	Raw Metal Analysis Data	219
Appendix C:	Quality Assurance/Quality Control Methods and Results	.244
Appendix D:	Metals Source Pilot Study	.253

х

ŝ

**EXECUTIVE SUMMARY** The Sierra Nevada, Klamath, Cascade, and Coast range mountains surrounding the Central Valley are rich in geological deposits of metal laden ores. Historic mining activity resulted in open mines and exposed tailings which leach metals into the Sacramento River and its tributaries. Runoff from mining operations has resulted in exceedances of water quality objectives, fish kills, and elevated metal concentrations in sediment and tissues of aquatic organisms (Nordstrom *et al.*, 1977; Wilson *et al.*, 1981; SWRCB, 1990; Montoya and Pan, 1992; Fujimura *et al.*, 1995; Saiki *et al.*, 1995; Cain *et al.*, 1998). In addition, metals in the upper and middle regions of the watershed have been linked to impacts in aquatic life using toxicity tests (Connor *et al.*, 1993; Bailey *et al.*, 1994; Connor *et al.*, 1994). However, metal concentrations and toxicity have not been well characterized in the Sacramento-San Joaquin River Delta.

The Bay Protection and Toxic Cleanup Program (BPTCP) was created to identify toxic hot spots, develop sediment quality objectives, and remediate toxic hot spots in California. The Central Valley Regional Water Quality Control Board utilized BPTCP funds to determine if metals threatened beneficial uses in the Delta. The current study had four objectives: 1) to determine if metal concentrations (i.e., arsenic, cadmium, chromium, copper, lead, nickel, and zinc) could be measured in the Sacramento-San Joaquin Delta during low and high flow periods using ultra clean methods with detection limits low enough to evaluate compliance with water quality objectives; (2) to define the extent of water quality objective exceedances in the Delta for metals; 3) to define the extent of metal associated toxicity throughout the Delta using the EPA three species toxicity tests; and 4) to determine the metal loading patterns into the Delta from the Yolo Bypass and Sacramento River (at Greene's Landing) during low and high flow periods. To address these objectives, fixed stations were monitored over multiple seasons and storm events. However, much of the sampling effort was focused during the winter to complement ongoing monthly metals monitoring by the Sacramento County Ambient Monitoring Program. The biotoxicity project is discussed in separate reports (Deanovic et al., 1996 & 1998). Because significant loads were identified entering the Delta during storm events, a study (Metals Source Pilot Study) was conducted during a single winter storm event to better characterize the source(s) of the loads.

Water samples were collected for metal analyses during the relatively normal 1993 water year (October 1992-September 1993: WY93), critically dry 1994 water year (October 1993-September 1994: WY94), and high flow 1995 water year (October 1994-September 1995: WY95). Flows in the combined discharge of the Sacramento River and Yolo Bypass peaked at 135,000 CFS on 28 March during WY93 and at 334,000 CFS on 13 March during WY95. As a result of the low rainfall during WY94, flows at Freeport did not exceed 30,000 CFS and the Yolo Bypass had measurable flows above 1000 CFS on only four days.

## Were low detection limits obtained using ultra clean techniques? Yes

• Evapoconcentration prior to analysis of field collected samples resulted in the detection of arsenic, cadmium, chromium, copper, lead, nickel, and zinc down to the low to mid parts per trillion range, well below values set for water quality objectives.

• Analysis of laboratory and field blanks indicated samples could be collected relatively free of metal contamination.

# Were water quality objectives exceeded for metals during the study? No

• USEPA National Ambient Water Quality Criteria and the USEPA Proposed California Toxics Rule Criteria were never exceeded in any of 549 samples collected from 15 Delta stations during critically dry, normal, and wet water years.

• The site-specific numeric water quality objectives for arsenic, copper, silver, and zinc were not exceeded in the Delta.

a)

#### What trends in metal concentrations were identified?

• During the critically dry WY94, total recoverable concentrations of chromium, copper, lead, nickel, and zinc increased with increasing flow conditions and increased sediment load in the Sacramento River at Greene's Landing.

• TSS or flow could be used to predict general levels (high versus low) of total recoverable copper, chromium, lead, nickel, and zinc during the drought-like conditions in WY94. Furthermore, these metals tracked each other very closely during this period such that high total recoverable zinc concentrations coincided with high total recoverable copper, chromium, lead, and nickel concentrations.

• During the high flow WY95, total recoverable cadmium, chromium, copper, and zinc concentrations at Greene's Landing were still significantly related to TSS indicating these metals were bound to suspended sediment particles during both dry and wet years.

• During the high flow WY95, total recoverable cadmium, chromium, copper, nickel, and zinc were inter-related and lead was not associated with any other metal. Using the inter-related nature of TSS and the grouped metals (i.e., copper, zinc, chromium, and cadmium), one could begin to utilize TSS levels as a general indicator for levels of these metals (e.g., high versus low concentrations).

• The value of these relationships is in designing when to collect samples if one is interested in sampling for high metal concentrations. For some metals, high flow events would be expected to produce high total recoverable metal concentration.

### Was metals related toxicity identified in the Delta? No

• Fifty eight samples exhibited toxicity during the study. Metals were never implicated in the Toxicity Identification Evaluation (TIEs) studies conducted on samples collected from the Delta which were toxic. However, TIEs could not be performed on all samples which exhibited toxicity due to budgetary constraints.

# Were metal loading patterns characteristic of hydrological conditions? Yes

• Depending on the metal, Sacramento River loads increased from approximately 460%

to 5,300% from the critically dry WY94 to the wet WY95. This indicates that high flow water years can greatly increase metal inputs to the Delta when compared to dry years.

• Sediment loading patterns in the Sacramento River and Yolo Bypass were nearly identical to the load patterns for copper and zinc during the wet WY95, with greater loads in the Bypass. These metals, as well as chromium, appeared to be transported into the Delta bound to sediment particles.

Constituent .	Bypass Load	River Load
Copper (kg)	296,000	144,000
% of total	67	33
Zinc (kg)	727,000	394,000
% of total	65	35
Chromium (kg)	472,000	155,000
% of total	74	26
Lead (kg)	64,700	54,400
% of total	54	46
Cadmium (kg)	1,550	1,660
% of total	48	52
Nickel (kg)	911,000	201,000
% of total	82	18
Arsenic (kg)	22,400	20,800
% of total	52	48
Sediment (metric tons)	2,500,000	1,300,000
% of total	66	34

#### What source(s) of metals were identified during the March 1995 high flow pilot study?

• Metal loading from historic mines in the Lake Shasta region could not be assessed because reservoir releases were maintained low to minimize downstream flooding.

• Areas of significant load contributions during the study included Cottonwood Creek in the upper Watershed and Cache Creek in the lower Watershed.

• Additional inputs of metals which resulted in high loads occurred between the Bend River bridge and Ord Ferry bridge and between County Road A-8 and Colusa. Both regions receive runoff from undammed creeks during major storm events.

Based on a lack of metals related toxicity and no exceedances of water quality objectives for metals in this study, future metals monitoring (excluding mercury) in the Delta as Regional Board special studies is not a high priority. However, staff recommend that ambient monitoring programs such as the Coordinated Monitoring Program, Regional Monitoring Program, Sacramento River Watershed Program, and CALFEDs Coordinated Monitoring and Research Program continue to include water column metals monitoring and that sediment testing and tissue analyses be included.

# **INTRODUCTION**

#### **BASIN DESCRIPTION**

The Sacramento-San Joaquin Delta Estuary is ecologically, aesthetically, and economically significant to the State of California. The area comprises over 700 miles of interconnected waterways and encompasses 1,153 square miles (Central Valley Regional Water Quality Control Board, 1994). The Delta, together with San Francisco Bay, is the largest estuary on the west coast of North America. It is fed by three main rivers, the Sacramento, the San Joaquin, and the Mokelumne, with a combined average unimpaired flow of about twenty-two million acre-feet per year. The Sacramento-San Joaquin Delta serves California as a significant water resource. Recognized beneficial uses include fisheries and wildlife habitat, agricultural supply, recreation, navigation, industrial process and municipal and domestic supply. Two statistics are presented below to help illustrate the environmental significance of the estuary to the people of California. First, over two-hundred-eighty species of birds and over fifty species of fish inhabit the freshwater portion of the estuary (San Francisco Estuary Project, 1992; Herbold and Moyle, 1989). This is considerably more than any other water body in the State of California (San Francisco Estuary Project, 1992). Second, over half of all the drinking water for the State of California is pumped from the Delta (San Francisco Estuary Project, 1992). The Sacramento River contributes over 80% of the drinking water to the Delta, but is also a major conveyance route for contaminants from upstream sources to the Delta.

#### SOURCES OF METALS

The Sierra Nevada, Cascade, Klamath, and Coast range mountains surrounding the Central Valley are rich in geological deposits of metal laden ores. Historic mining activity resulted in open mines and exposed tailings which leach metals into the upper Sacramento River Watershed and its tributaries. Relatively few historic mining operations contributed the majority of metals to regional waters. Runoff from mining operations in the upper Watershed has resulted in exceedances of water quality objectives, fish kills, and elevated metal concentrations in sediment and tissues of aquatic organisms (Nordstrom *et al.*, 1977; Wilson *et al.*, 1981; SWRCB, 1990; Montoya and Pan, 1992; Fujimura *et al.*, 1995; Saiki *et al.*, 1995; Cain *et al.*, 1998). Since the implementation of acid mine drainage controls on Iron Mountain Mine (IMM), exceedances of water quality standard exceedances in Keswick Reservoir have been reduced (Heiman, pers. comm.). However, limited water-quality standard exceedances in Keswick Reservoir have been reported as recently as January, 1997 (Alpers, written comm.). The spatial and temporal patterns of metal dispersion from mines are variable (Alpers, written comm.). Although mine drainage is a significant contributor of metals to the system, metals also enter from other sources.

Discharges from agriculture areas are important sources of metals laden runoff to the lower Sacramento River. Agricultural drains discharged an estimated 74% of the total chromium load, 75% of total nickel load, and 17% of the total copper load in the Sacramento Valley in 1985 (Montoya *et al.*, 1988; CVRWQCB, 1989). Agricultural applications of the pesticide copper sulfate [i.e., hydroxide and sulfate (basic and pentahydrate)] reached 6,471,596 lbs. in California

during 1993 (Department of Pesticide Regulation, 1995). This quantity represents a 17% increase from 1991 applications (Department of Pesticide Regulation, 1993). Of the total applied during 1993, 1,808,043 lbs. of copper were applied on rice crops (Department of Pesticide Regulation, 1995). This quantity represents a 21% increase from 1991 applications (Department of Pesticide Regulation, 1993). By far, the majority of the rice cultivation in California occurs in the Sacramento River Watershed. Copper levels measured in agricultural drainage of the Sacramento River Watershed during 1985 were significantly higher during the rice growing season (May-June) compared to January-April levels (Montoya et al., 1988; CVRWQCB, 1989). Copper use on orchards is also increasing, but the potential for off site movement has not been investigated. United States Geological Survey (USGS) load estimates for the dissolved and colloidal forms of copper during July and September 1996 and May-June 1997 show increases on the Sacramento River between Colusa and Verona where water enters from the Colusa Basin Drain, Sacramento Slough, and other tributaries carrying agricultural return flows (Alpers, written comm.). Furthermore, data collected for the USGS National Water Quality Assessment (NAWQA) program on the Sacramento River indicate loads of copper into the Colusa Basin Drain during June 1997 were slightly less than that from Iron Mountain Mine via Spring Creek during the same sampling period (Alpers, written comm.). However, the transport, fate, and biotic effects of copper from the drains into the softer waters of the Sacramento River are not completely understood.

Another important source of metal input to the system is urban runoff which carries metals from transportation and homeowner uses into regional waters. Urban runoff has been estimated to contribute approximately 94% of the lead, 8-9% of the copper, cadmium, and zinc, and 14-16% of the nickel and chromium total loads in the Sacramento River Watershed (Montoya *et al.*, 1988; CVRWQCB, 1989). The American River in the lower Sacramento River Watershed receives urban runoff containing metals from several sources in the Sacramento metropolitan area. Total recoverable copper, lead, and zinc concentrations increased from upstream to downstream monitoring stations on the American River when concentrations were averaged from July 1994 to 1995 (Larry Walker Associates, 1996). Although increased concentrations were observed, they were minor and well below water quality objectives and were at least in part associated with wet weather urban inflows. Of concern to the Central Valley Regional Water Quality Control Board (CVRWQCB) are the effects metal sources may have on aquatic life throughout the Watershed, including the Delta.

## METAL TOXICITY

The most sensitive beneficial use when metals are considered is the protection of aquatic life. In order to understand the scope of metal impacts in the Delta, the spatial and temporal extent of effects in the upper Watershed must first be characterized. The Basin Plan of the Central Valley Regional Water Quality Control Board contains a narrative toxicity objective which states that all waters must be maintained free of toxic substances in concentrations that cause detrimental physiological responses in aquatic organisms (Central Valley Regional Water Quality Control Board, 1994). The Basin Plan also states that compliance with this narrative objective can be

evaluated in a number of ways, including the use of the US EPA three species bioassay protocols and by comparing metal concentrations with available objectives and criteria. The Regional Board uses both approaches to evaluate threats posed by elevated metal concentrations. These bioassays measure changes in growth, survival, and/or reproduction of three species from three diff\_rent phyla and trophic levels. Regional Board staff have relied on the use of the three species bioassays since 1986 to assess compliance with the Basin Plan's narrative toxicity objectives. Toxicity testing results have indicated metal related toxicity in the Shasta Mining District.

Studies conducted from 1991-1992 to monitor toxicity and metal concentrations in discharges from major reservoirs identified relatively few incidents of toxicity (Goetzl and Stephenson, 1993; Connor *et al.*, 1994). Results may have been influenced by climate conditions, such as the ongoing drought, as well as mine remediation projects. Significant toxicity to the freshwater alga *Selenastrum* was detected in the Sacramento River downstream from the Keswick Dam. Toxicity was detected in 75% of the samples collected from Keswick Reservoir (Connor *et al.*, 1994). When compared to 18 other sites sampled throughout the Watershed, samples collected downstream from Keswick Dam exhibited the highest frequency of toxicity and the greatest number of exceedances of cadmium, copper, and zinc water quality objectives (Goetzl and Stephenson, 1993). There was a positive relationship between *Selenastrum* was detected in a similar study conducted in 1993 (Bailey *et al.*, 1994).

In conclusion, metal analyses and toxicity testing conducted since 1988 provide some indication of metals impacting aquatic life in the Sacramento River from mining. However, no studies have been undertaken in the Delta to determine the overall importance of metals and toxicity on aquatic resources.

# WATER QUALITY CRITERIA/OBJECTIVES

The CVRWQCB is not only interested in characterizing toxicity to aquatic organisms, but also in characterizing regional waters for compliance with numeric water quality objectives. However, in the past it was difficult to use monitoring data to evaluate compliance with existing metal water quality objectives because either the detection limits were too high (e.g., above actual instream concentrations) or the quality assurance and control were not rigorous. Further difficulty has been encountered because of changes in water quality objectives in California. During 1995, criteria used to protect aquatic life from inorganic constituents were promulgated in the California Inland Surface Waters Plan. These objectives were based on the US EPA National Ambient Water Quality Criteria. However, values for the Inland Surface Waters Plan were expressed as total recoverable metal, while the US EPA criteria were expressed as dissolved metal (Marshack, 1995). The Inland Surface Waters Plan was repealed in 1994 as a result of a legal challenge, leaving California without enforceable numerical water quality objectives for priority toxic pollutants in surface waters as required for each state by the Clean Water Act, except for certain site-specific numeric water quality objectives in the Water Quality Control Plan for the

CVRWQCB. The Water Quality Control Plan contains numeric water quality objectives for several metals in the Sacramento River, including arsenic, barium, cadmium, copper, cyanide, iron, manganese, silver, and zinc. In 1997, the US EPA proposed to promulgate water quality criteria for priority toxic pollutants for California's inland surface waters by developing the California Toxics Rule. In addition to the site-specific water quality objectives in the Water Quality Control Plan, criteria currently used as guidance for the CVRWQCB to protect freshwater aquatic life from inorganic constituents are the US EPA Proposed California Toxics Rule and the US EPA National Ambient Water Quality Criteria. As of 1998, both criteria are expressed as dissolved metals (Marshack, 1998). Therefore, additional metal monitoring was needed to better assess compliance.

# BAY PROTECTION AND TOXIC CLEANUP PROGRAM

In 1989, the California Water Code was amended to create the Bay Protection and Toxic Cleanup Program (BPTCP). The three primary goals of the program are to: 1) identify toxic hot spots; 2) develop sediment quality objectives; and 3) remediate toxic hot spots, either through cleanup efforts, mitigation or prevention. Section 13391.5 of the Water Code defines toxic hot spots as: "....[L]ocations in enclosed bays, estuaries, or adjacent waters in the 'contiguous zone' or the 'ocean' as defined in Section 502 of the Clean Water Act (33. U.S.C. Section 1362), the pollution or contamination of which affects the interests of the State, and where hazardous substances have accumulated in the water or sediment to levels which (1) may pose a substantial present or potential hazard to aquatic life, wildlife, fisheries, or human health, or (2) may adversely affect the beneficial uses of the bay, estuary, or ocean waters as defined in the water quality control plans, or (3) exceeds adopted water quality or sediment quality objectives."

The BPTCP identifies five conditions that are used to define toxic hot spots.

- 1. Exceedance of water quality objectives
- 2. Toxicity associated with a toxic pollutant
- 3. Exceedance of tissue contaminant levels
- 4. Impairment of resident organisms
- 5. Degradation of populations or communities associated with toxic pollutants

Using Bay Protection Toxic Cleanup Program funds, the Central Valley Regional Water Quality Control Board conducted a study from May 1993 to December 1996 to characterize toxicity, metal concentrations, and metal loads in the Delta. The overall focus of this study was to determine if there were metal impacts in the Delta, and if so, identify whether the impacts were a result of transport or *in situ* processes. Prior to this study, there had been ongoing monitoring efforts in the Delta for many years. However, the monitoring was deficient in three general areas. First, as stated above, the monitoring focused on chemical analyses with a lack of rigorous quality assurance and high detection limits. Second, the monitoring efforts did not incorporate measurements of multiple metals and organic compounds. In addition, toxicity tests were not conducted concurrently with monitoring therefore prohibiting an assessment of the contribution

â

metals had on aquatic life in the Delta. Furthermore, the situation of multiple metals working in an additive manner to cause toxicity is potentially important in the Delta because of the high load and diversity of inputs. Third, most of the annual metal load to the Delta is associated with major storm events. Past monitoring within the Delta had not adequately characterized metal level, and loads to the Delta during storm events.

The current study had four objectives: 1) to determine if metal concentrations (i.e., arsenic, cadmium, chromium. copper, lead, nickel, and zinc) could be measured in the Sacramento-San Joaquin Delta during low and high flow periods using ultra clean techniques with detection limits low enough to evaluate compliance with water quality objectives; 2) to define the extent of water quality objective exceedances in the Delta for metals; (3) to define the extent of metal associated toxicity throughout the Delta using the EPA three species toxicity tests; and (4) to determine the metal loading patterns into the Delta from the Yolo Bypass and Sacramento River (at Greene's Landing) during low and high flow periods. To address these objectives, fixed stations were monitored for metals and biotoxicity over multiple seasons and storm events. However, much of the sampling effort was focused during the winter to complement ongoing monthly monitoring by the Sacramento County Ambient Monitoring Program. The biotoxicity project is discussed in separate reports (Deanovic *et al.*, 1996; 1998). Because significant loads were identified entering the Delta during storm events, a pilot study was conducted ("Metals Source Pilot Study") during a single winter storm event to better characterize the source(s) of the loads (Appendix D).

)

# MATERIALS AND METHODS

# QUALITY ASSURANCE PROGRAM

The purpose of the Quality Assurance Program was to ensure the data were generated under conditions that accurately reflected the quality of the water sample. Standardized procedures were followed in all aspects of research. These methods are described in the Project Quality Assurance Plan designed for this project (Connor *et al.*, 1995). Both accuracy and precision were addressed in the quality assurance/quality control (QA/QC) document. A full description of the QA/QC methods and data can be found in Appendix C.

# SAMPLE LOCATIONS

Water samples were collected for metal analyses and toxicity assessments during the 1993 (October 1992-September 1993), 1994 (October 1993-September 1994), and 1995 (October 1994-September 1995) water years. Sampling sites for metal analyses included main river inputs to the Delta, back sloughs and small upland drainages, areas receiving urban runoff, and points along the path of water movement across the Delta (Fig.1; Table 1). In addition, samples for were collected for a pilot study ("Metals Source Study") designed to identify sources of metals loads into the Delta and upstream to Shasta Dam during a single storm event (Fig. D-1; Table D-1). Additional sampling sites were selected for toxicity assessments (Deanovic *et al.*, 1996; 1998). Detailed site descriptions are provided in Appendix A and D.

### SAMPLE COLLECTION AND STORAGE

#### Metal Analyses

Samples for total recoverable and dissolved metals analyses were collected by Regional Board staff. All samples were collected from beneath the water surface by boat, from a bridge, or from the bank in a rapidly moving section of the water course. The samples were collected by inserting cleaned bey-a-line tubing through 25 feet of PVC pipe (Goetzl and Stephenson, 1993). The use of the pipe allowed the sampling point to be about 20 feet from the shore and thus minimized edge effects. All samples were pumped from the point of collection (using a peristaltic pump) through 25 feet of acid-cleaned tubing directly into an analysis bottle containing acid. The tubing ended in a dust free sampling box which contained the sampling bottles. The bottles were handled without opening the box through gloved port holes. The tubing and the box were employed to minimize the exposure of the samples to airborne contamination. The exception to this procedure was the sampling conducted during high flow events. This sampling used an acid washed one gallon borosilicate glass composite sampler instead of a glove box for sample collection. All analysis bottles were double bagged except while being filled. All samples collected for determining the concentration of dissolved metals were filtered through a 0.45 micron polypropylene MSI cartridge filter attached to the end of the tubing. At each site water conditions, sampling conditions, water temperature, pH, and EC were recorded. After collection, all samples were triple bagged and placed in a dust free container until shipped to the Moss

Landing Mussel Watch Lab. The details of the sampling equipment and procedures are fully described in Goetzl and Stephenson (1993).

# **Toxicity Samples**

Bioacsay surveys were conducted from May 1993 to December 1996 in the Delta. Site locations, method of water collection, and sample storage are contained in Deanovic *et al.*, (1996) and (1998). Bioassays were run on all water samples collected from the Delta for metal analyses. However, additional sites were only tested for toxicity. If toxicity was detected and no samples were collected for metal analyses, then sub-samples were taken from the bioassay water and placed in a one liter polyethylene bottle (containing nitric acid) for determination of total recoverable and dissolved (filtered with a Gelman A/E glass fiber filter, nominal pore size of 0.45  $\mu$ m) metal concentrations.

WERE LOW DETECTION LIMITS OBTAINED USING ULTRA CLEAN TECHNIQUES?

Total recoverable and dissolved (0.45  $\mu$ m filtered) metal concentrations were analyzed by the California Department of Fish and Game Mussel Watch Laboratory and at the Moss Landing Marine Lab Trace Metals Laboratory, using ultra-clean facilities and graphite furnace atomic absorption spectrophotometry (Goetzl and Stephenson, 1993). Twenty percent of the samples were split samples analyzed by the Trace Metals Laboratory. Samples were analyzed using an evapo-concentration technique to obtain low detection limits (Goetzl and Stephenson, 1993; Goetzl *et al.*, 1994, 1995). The essence of this procedure is that a sample is concentrated twenty-five fold by evaporation followed by an acid-treatment to re-dissolve the sample. This procedure can achieve detection limits in the parts per trillion range.

# Atomic Absorption Methods (Trace Metal Lab)

Samples were analyzed by flameless Atomic Absorption (AA) on a Perkin-Elmer Zeeman 5000 Atomic Absorption Spectrophotometer equipped with an HGA 500 graphite furnace at the Salinas facility of Moss Landing Marine Laboratories. Due to high concentrations, a few samples were analyzed using flame AA on a Perkin-Elmer 603 AAS. Samples and standards were prepared in a laminar-flow clean bench inside the trace metal lab. To ensure accurate results, the samples were analyzed using the stabilized-temperature platform technique. The characteristic mass for each element was computed to ensure the proper functioning of the Zeeman AA. Samples may be analyzed using a matrix modifier made up from ultra-clean chemicals. When no modifier is used, high-char temperatures allow interfering matrix components of the sample to be volatilized prior to atomization. Single spike additions to samples allow a check for recovery when standards are linear. Finally, the SLRS-2 (1993-94 samples) or SLRS-3 (1994-95 samples) river water standard reference material was evapoconcentrated and analyzed with each set of samples.

### AA Methods (Mussel Watch Lab)

The Mussel Watch Lab is located at the Moss Landing Marine Laboratories in Moss Landing, California. Samples were analyzed by furnace AA on a Perkin-Elmer Zeeman 3030 Atomic

Absorption Spectrophotometer with an AS60 auto-sampler and HGA 500 graphite furnace. Samples, blanks, matrix modifiers, and standards were prepared using clean techniques inside a clean lab. Milli-Q water and ultra-clean chemicals were used for all standard preparations. To ensure accurate results the samples were analyzed using the stabilized-temperature platform technique. Matrix modifiers were used when the components of the matrix interfered with adsorption. Matrix modifiers were used for arsenic in all samples and for lead in 1993-94 samples. Blanks and a standard reference material (SLRS2 river water) were evapoconcentrated and analyzed with each set of samples.

÷

0

# WERE WATER QUALITY OBJECTIVES EXCEEDED?

Compliance with site-specific numeric water quality objectives described in the Water Quality Control Plan was assessed for samples collected from the Delta (CVRWQCB, 1994). In addition, the more stringent US EPA Proposed California Toxics Rule and the US EPA National Ambient Water Quality Criteria (expressed as four day average criteria) to protect freshwater aquatic life (Marshack, 1998) were compared to hardness corrected dissolved metal concentrations to determine whether exceedances occurred in the Delta during the study.

# WAS METALS RELATED TOXICITY IDENTIFIED IN THE DELTA?

Standardized U.S. EPA freshwater bioassay protocols were used for this study (U.S. EPA, 1994). The three organisms used in the laboratory assays were: (1) a primary producer, the green algae *Selenastrum capricornutum*; (2) a primary consumer, the zooplankton *Ceriodaphnia dubia*; and (3) a secondary consumer, the fathead minnow, *Pimephales promelas*. A complete description of the methodologies applied in testing ambient water samples for toxicity can be found in Deanovic *et al.*, (1996; 1998). When toxicity was detected in a sample, follow-up toxicity identification evaluation (TIE) procedures coupled to analytical chemistry were implemented to help determine the cause. Briefly, samples were tested for toxicity following several manipulations designed to render certain chemical/elemental constituents in the sample non-toxic. In addition, methods were applied to recover the chemical/elemental causes of the observed toxicity. A complete description of TIE procedures can be found in U.S. EPA (1991; 1992) and Bailey *et al.*, (1996).

# Statistical Methods and Definition of Toxicity

Toxicity was defined as a statistically significant difference (p<0.05) between a sample and the laboratory control. Bartlett's Test for homogeneity of variance was run on all fish growth and mortality, *Ceriodaphnia* reproduction, and algal growth data. When the data variance was homogeneous, the samples were compared to the controls using Analysis of Variance and Dunnett's mean separation tests. If the data variance was not homogeneous, then comparisons were made against the control using Kruskal-Wallis and Dunn's non-parametric multiple comparison. *Ceriodaphnia* survival was compared against the control with a Fisher's Exact Test. No statistical analyses were conducted on TIE results. Acute toxicity was defined as a statistically significant difference in mortality within 96 hours between an ambient water and

laboratory control sample. *Selenastrum* toxicity was defined during the 1993-1994 monitoring as a statistically significant difference in cell counts between an ambient sample and a laboratory control. Due to the low frequency of statistically significant toxicity when ambient samples were compared to laboratory control samples, cell counts in the 1994-1995 samples were also compared to other field samples collected on the same day to determine if the relative level of cell counts differed among stations. Consult Deanovic *et al.*, (1996) and (1998) for additional information regarding the statistics applied for the toxicity test results.

# WERE METAL LOADING PATTERNS CHARACTERISTIC OF HYDROLOGICAL CONDITIONS?

### Water Years 1993, 1994, and 1995

Water year 1993 (October 1992-September 1993) was classified as a relatively normal water year in the Sacramento Basin. Precipitation in the region during water year 1993 was 149 percent of the long-term average while runoff was about 125 percent of the 1961-1990 median based on five representative streamflow records (Mullen *et al.*, 1994). Water year 1994 (October 1993-September 1994) was classified as critically dry and is identified in this report as a "dry year". Precipitation in the region during water year 1994 was 36 percent of the long-term average while runoff was about 69 percent of the 1961-1990 median based on five representative streamflow records (Friebel *et al.*, 1995). During such dry years, the Sacramento River serves as the primary source of water transport from the Sacramento Basin to the Delta. Conversely, water year 1995 (October 1994-September 1995) was characterized by high flows which resulted in water transport to the Delta via the Sacramento River and the Yolo Bypass. Although summary hydrologic conditions for the region are not available for water year 1995, combined flows for the Sacramento River and Yolo Bypass peaked at 334,000 CFS and 16 inches of rain fell in the City of Sacramento in January (Foe and Croyle, 1998). Therefore, water year 1995 was classified as a "wet year" for the purposes of this study.

#### Flow Rates

)

Daily water discharge rates from the Sacramento River at Greene's Landing and the Yolo Bypass at Prospect Slough were obtained from USGS flow gauges (Mullen *et al.*, 1994; Friebel *et al.*, 1995; Markham *et al.*, 1996; California Data Exchange Center, 1998).

#### Load Calculations

Bulk daily metal loads (kg/day) at Prospect Slough and the Sacramento River at Greene's Landing were calculated for arsenic, cadmium, chromium, copper, lead, nickel, and zinc from January through April 1994 and 1995. Mercury loads were not included in this report but can be found in Foe and Croyle (1998). Two methods were employed to calculate loads. First, regression analyses were performed to determine if significant relationships existed between flow and total recoverable concentrations of each individual metal (Steel and Torrie, 1960). When the variance appeared to greatly increase/decrease with increasing flow, the data were log transformed and a comparison of residuals was conducted. If the variance in the data was then similar with increasing flow, then a best fit line was applied to the log transformed data. When regression

analyses were significant, models were developed for each metal using a linear regression with flow as the independent variable and total recoverable concentration as the dependent variable. Daily flows were entered into the linear regression equation to obtain daily predicted metal concentrations. Daily predicted concentrations ( $\mu g/l$ ) were then multiplied by daily flow to obtain model generated estimates of metal load. This method was used to provide a rough estimate of loads when significant relationships existed between flow and metal concentrations, however transformation of the data may affect concentrations by 5-25%. Alternative methods are available which provide a more rigorous estimate of load (Cohn *et al.*, 1989; Helsel and Hirsch, 1992). These methods were not applied here since the objective was to provide a rough estimate of load fluctuations between wet and dry years and the sample collection design could not be properly applied to the models.

A second method was applied when a regression was not significant. Loads were calculated individually for Prospect Slough and Greene's Landing by multiplying daily flow readings by the average metal concentration ( $\mu$ g/l) measured in all field samples at each of the two sites ("Average Concentration Method"):

Daily Load (kg) = [Avg. metal concentration ( $\mu$ g/l)] x (2.445 x 10<sup>-3</sup>) x [Flow (CFS)]

Total load was estimated by summing the daily loads for each period. Due to the uncertainties in flow measurements ( $\pm$  10%) and the uncertainty involved with the regression analyses, the number of significant figures for load calculations was set at three for the purposes of load comparisons. Loads were also calculated using data from the Sacramento Coordinated Water Quality Monitoring Program's Ambient Monitoring Program (AMP), using the Average Concentration Method and regression models. This permitted a comparison of load estimates calculated for two independent monitoring efforts on the Sacramento River at Greene's Landing and River Mile 44. However, AMP monitoring relied on different collection methods, sample frequencies, sample locations, and temporal pattern of sampling than those of this study (Larry Walker and Associates, 1996).

# WHAT SOURCE(S) OF METALS WERE IDENTIFIED DURING THE METALS SOURCE PILOT STUDY?

3

Water samples were collected for a one-time pilot study during a major storm event in March 1995 to assess the relative metal load contribution from sources upstream of the Delta, primarily in the Sacramento River Watershed. Sampling methods followed those described above with sampling dates reported in Table D-1. The study was designed to assess metal loads, therefore only total recoverable concentrations were quantified. No toxicity samples were collected and the lack of dissolved metals analyses prohibited an assessment of water quality objective exceedances. Although the objective of the pilot study was to track sources of metals during a high flow event, the data could not be used to quantify the load contribution from mines in the area of Lake Shasta and Keswick Reservoir because discharges from the reservoirs were

maintained at low levels to minimize downstream flooding. This resulted in samples downstream of the reservoirs which were negligibly affected by runoff from mines. A full description of the results of the Metals Source Pilot Study can be found in Appendix D.

.

)

# **RESULTS AND DISCUSSION**

# QUALITY ASSURANCE/QUALITY CONTROL PROGRAM

Field blanks collected on nine occasions indicated negligible contamination with no metals detected above 1  $\mu$ g/l (Table C-1). Field duplicates were collected on 64 occasions with a resulting average difference between two laboratories of 16% (Table C-2). Analysis of laboratory blanks resulted in 65% of the individual metals data quantified as below the detection limits for the methods applied in this study (Table C-3). Intra-laboratory precision results ranged from 2 to 20%, depending upon the metal (Goetzl *et al.*, 1994, 1995). A more complete description of the quality assurance and quality control results can be found in Appendix C.

# HYDROLOGICAL CONDITIONS

Water samples for chemical analyses were collected and toxicity assessments were performed during the relatively normal 1993 water year (WY93), critically dry 1994 water year (WY94), and high flow 1995 water year (WY95). Flows in the combined discharge of the Sacramento River and Yolo Bypass peaked at 135,000 on 28 March during WY93 and at 334,000 CFS on 13 March during WY95 (Mullen *et al.*, 1994; Markham *et al.*, 1996). As a result of the low rainfall during WY94, flows at Freeport did not exceed 30,000 and the Yolo Bypass had measurable flows above 1,000 CFS on only four days (Fig. 2; Friebel *et al.*, 1994).

WERE LOW DETECTION LIMITS OBTAINED USING ULTRA CLEAN TECHNIQUES? Evapoconcentration of field collected samples resulted in the detection of arsenic, cadmium, chromium, copper, lead, nickel, and zinc down to the low to mid parts per trillion range (Table 2). This method improved upon other analytical methods and resulted in detection limits which were among the lowest of four programs monitoring metals in the Sacramento River Watershed (Table 2; Larry Walker and Associates, 1996; Sacramento Regional County Sanitation District, 1996). The advantage of a lower detection limit is metals can be quantified at concentrations which are well below values set for water quality objectives. Furthermore, these lower detection limits minimize the frequency of non-detects, permit the detection of metals at and below actual instream values, and provide for a more accurate estimate of metal loads (Goetzl and Stephenson, 1993).

S.

# WERE WATER QUALITY OBJECTIVES EXCEEDED?

Site-specific numeric water quality objectives in the Water Quality Control Plan for the CVRWQCB were compared to dissolved metal concentrations (0.45 µm filtered) in samples collected from 15 Delta stations during WY94 and WY95 to determine if the exceedances occurred (CVRWQCB, 1994; Tables 3-17). The site-specific numeric water quality objectives for arsenic, copper, silver, and zinc in the Delta were not exceeded.

Dissolved metal concentrations were compared to the more stringent USEPA National Ambient Water Quality Criteria and the USEPA Proposed California Toxics Rule Criteria (Tables 3-17). With the exception of As, criteria for the metals quantified in this study are water hardness dependent. No water quality criteria were exceeded for 549 individual Delta metal analyses (Table 18).

## WAS METALS RELATED TOXICITY IDENTIFIED IN THE DELTA?

Waters sampled from the Delta region were tested for toxicity during WY94 and WY95 using the EPA three species toxicity tests to determine if aquatic life was impacted. Deanovic *et al.*, (1996) and Deanovic *et al.*, (1998) contain a full description of the results. In brief, 34 and 58 (including relative reductions in algal cell counts) toxic events were detected during WY94 and WY95, respectively (Table 19 & 20).

Approximately 7% of the samples collected from the Delta region tested toxic to *Ceriodaphnia* during WY94, while samples were toxic 14% of the time during WY95. Most of the toxicity (e.g., 68%) to *Ceriodaphnia* occurred in samples collected from back-sloughs and small upland drainages. Toxicity Identification Evaluations were performed on toxic samples during both years to determine if the cause of toxicity could be determined. Typically, toxicity was related to pesticides, including organophosphates, carbamates, and unknown metabolically activated compounds. Metals were never implicated in TIE studies conducted on the samples which exhibited toxicity (Table 19 & 20). However, TIEs were not performed on all toxic samples due to budgetary limitations.

On 329 occasions *Selenastrum* toxicity tests were performed on samples collected from the Delta during WY94 to WY95. The number of toxic events remained fairly constant at about 1% for both water years (Table 19). However, nearly 30% of the ambient samples exhibited reductions in cell counts relative to other ambient samples collected on the same day in WY 95 (Table 20). As with *Ceriodaphnia*, the majority of the events with reduced cell counts occurred in the back-sloughs and small upland drainages (Table 20). TIE tests on Delta samples which exhibited toxicity implicated non-polar organics as causative toxicants and, as with the *Ceriodaphnia* TIEs, no examples of metal related toxicity were found.

*Pimephales* toxicity tests were conducted on 216 Delta samples, with the bulk of the testing during WY94 (Table 19). Approximately 9% of the samples were toxic in WY94 with toxicity in all water categories except urban runoff receiving waters. No TIEs were conducted on these samples so the causative agents remain unknown.

The EPA Three Species may not necessarily be the most sensitive organisms to metals. Tables were created documenting the most sensitive 10-15 literature reports for algae, invertebrates, and fish. Dissolved metal concentrations were selected as this is the form most bioavailable to aquatic organisms during water column exposure. Effect levels from the literature values were then compared to the highest dissolved concentration measured in the Delta for each metal to

assess the potential for effects in species other than the three species used in the EPA toxicity tests applied in this study (Reyes, 1994; Table 21).

Dissolved lead peaked at 3.87  $\mu$ g/l (at 5-mile Slough; hardness = 80 mg/l) and averaged 0.31  $\mu$ g/l over the combined water years (Table 21). No algal responses would be expected at these concentrations (Table 22). Unicellular invertebrates, such as ciliates, had reduced oxygen uptake after only four minutes exposure to 0.75  $\mu$ g/l lead (Table 23; Slabbert and Morgan, 1982). Three-spine stickleback, a freshwater fish, had increased mortality in response to 0.2  $\mu$ g/l dissolved lead exposure after five days (Table 24; Jones, 1938). More recent work indicates carp enzyme systems are sensitive to lead down to 1.1  $\mu$ g/l (Table 24; Nakagawa *et al.*, 1995).

2

2

The average dissolved concentration of arsenic was 1.28  $\mu$ g/l and the highest concentration was 3.03  $\mu$ g/l (Table 21; at 5-mile Slough; hardness = 80 mg/l). Phytoplankton exhibited altered photosynthetic productivity following long-term exposure to 1.5  $\mu$ g/l arsenic, however exposure for 109 days at this concentration in the basin is highly unlikely (Table 25; Wangberg *et al.*, 1991). Fifty percent of *Daphnia duplex* were immobilized following exposure to 0.5  $\mu$ g/l arsenic for as little as one day (Table 26; Lilius *et al.*, 1995). Fish did not respond to arsenic exposure until concentrations exceeded 27  $\mu$ g/l (Table 21).

Dissolved chromium concentrations reached 5.39  $\mu$ g/l (hardness = 98 mg/l) at Duck Slough and averaged 1.34  $\mu$ g/l from 1993-1995 (Table 21). Algal responses occurred from 2  $\mu$ g/l to 5.2  $\mu$ g/l and included altered biomass and incipient growth inhibition (Table 28; Bringmann, 1975; Shabana *et al.*, 1986). *Selenastrum* responses were not reported until 20  $\mu$ g/l (Table 28; Pillard *et al.*, 1987). The most sensitive response of any aquatic invertebrate in the USEPA Aquire Database was decreased survival in an euglenoid down to 1  $\mu$ g/l (Table 29; Yonge *et al.*, 1979). Environment Canada (1994) reported toxicity in some zooplankton species at chromium concentrations of 0.5  $\mu$ g/l. Cytogenetic alterations and changes in growth were reported in carp at 0.05  $\mu$ g/l and 1.5  $\mu$ g/l, respectively (Table 30; Al-Sabti *et al.*, 1994; Mao and Wang, 1990).

Greene's Landing had the highest measured dissolved nickel concentration of 26  $\mu$ g/l (hardness = 44 mg/l) and the average for the study was 2.72  $\mu$ g/l (Table 21). Blue-green algae exhibited mortality at concentrations down to 1.2  $\mu$ g/l (Table 31; Bringmann and Kuhn, 1978). The EC<sub>50</sub> for *Selenastrum capricornutum* exposed for four days to nickelous chloride was 6.3  $\mu$ g/l (Table 45; Blaise *et al.*, 1986). Mortality was recorded for *Ceriodaphnia dubia* down to 3.8  $\mu$ g/l (Table 32; Kszoz *et al.*, 1992). No fish responses were reported in this concentration range (Table 33).

The maximum dissolved concentration of copper measured in this study was 9.48  $\mu$ g/l (at Greene's Landing; hardness = 62 mg/l) which has been shown to have effects on fish, invertebrates, and algae (Table 21). Freshwater fish responses ranged from avoidance to death (Table 34; Reyes, 1994). This concentration was lethal to several species of water flea for exposure durations down to two days (Table 35; Reyes, 1994). Algal responses ranged from altered photosynthetic output to decreased growth and altered metabolism (Table 36; Reyes, 1994).

The highest dissolved zinc concentration measured during monitoring was 70.2  $\mu$ g/l (at 5-mile Slough; hardness = 80 mg/l) (Table 21). This concentration is high enough to have potential effects on aquatic life. The most sensitive fish response in the literature was avoidance of solutions containing 5.6  $\mu$ g/l zinc sulfate by rainbow trout (Table 37; Sprague, 1964b). Invertebrates, such as the aquatic sowbug, experienced mortality at 10  $\mu$ g/l (Table 38; Migliore & DeNicola Guidici, 1990). Algae exhibit population declines (as measured by declines in cell numbers) when exposed to concentrations down to 5  $\mu$ g/l (Table 39). This concentration is slightly above the mean concentration when both water years were averaged. Exposures of *Selenastrum* for seven days at 5  $\mu$ g/l, as opposed to the four day exposures in this study, resulted in inhibited cell growth.

Cadmium concentrations peaked at 0.55  $\mu$ g/l (at Greene's Landing; hardness = 72 mg/l) and averaged 0.3  $\mu$ g/l in this study (Table 21). Exposure of rainbow trout to comparable concentrations for 18 months resulted in reduced survival (Table 40; Birge *et al.*, 1981). Other more short term effects include albinism in catfish (Table 40; Westerman and Birge, 1978). Invertebrates, such as copepods and water fleas, are reported to respond at this concentration range with increased mortality (Table 41). Algal responses to cadmium are reported to occur in the parts per billion range (Table 42; Reyes, 1994).

Some of the potential responses of algae, invertebrates, and fish described above would obviously be affected by the duration of exposure, which is difficult to assess from the composite Delta samples. Furthermore, some of the dissolved metal could be biologically unavailable because of high organo-iron complexes present in the Delta. However, the maximum dissolved concentrations of metals reported in this report may be an underestimation of actual instream maxima. For example, total recoverable metal concentrations measured during the metals source pilot study were, by far, the highest measured during the three water years (Appendix D). No dissolved concentrations were measured during the source study. Furthermore, none of the water samples collected during the metals source study were tested for toxicity due to the project objectives. It is possible that high total recoverable concentrations in the metals source pilot study coincided with higher dissolved metal concentrations than those presented in Table 21.

# WERE METAL LOADING PATTERNS CHARACTERISTIC OF HYDROLOGICAL CONDITIONS?

The objectives of the metal loads component of this study were to: (1) estimate loads on the mainstem lower Sacramento River from January to April during a critically dry and a wet year and determine how they vary with hydrological conditions and (2) determine the spatial partitioning of loads during a wet year when water enters the Delta from the Yolo Bypass and lower Sacramento River. The emphasis of this study on high flows was designed to complement ongoing monthly metals monitoring by the Sacramento County Ambient Monitoring Program. Load calculations were based on a regression relationship and/or the Average Concentration (AC) method (see methods). More rigorous load evaluation methods are available (Cohn *et al.*, 1989).

However, the intent here was to provide rough load estimates and the two methods selected were considered adequate for this purpose.

Regression models for WY94 consistently estimated lower loads at Greene's Landing during WY94 when compared to the AC method (Table 43). When significant, the regression model approach was considered to be more robust because it tested for statistical fitness whereas the AC approach lacked statistical analyses. The load estimate for cadmium during the dry WY94 was the lowest of all metals, with 698 kg contributed to the Delta over the four month time period (Table 43). Zinc load was the highest of all metals, ranging from 37,900 to 50,700 kg depending upon the method selected.

2

Water years were compared using the regression model for WY94 and the AC method for WY95. Increased flows and higher total recoverable metal concentrations for most metals combined to result in increases in metal loads ranging from approximately 240% to 2,400% (Table 43). This is somewhat of an invalid comparison because much of the water entering the Delta during WY95 was in the Bypass and, therefore, this load contribution would not be included in these values. When total loads into the Delta from the Sacramento River Watershed (e.g., Greene's Landing + Yolo Bypass) for WY95 are compared to WY94, percent increase in loads ranges from 460% for cadmium to 5,300% for chromium (Table 43 & 44). To put these percentages in the context of the amount of metals added to the Delta, cadmium loads increased from 698 kg in WY94 to 1,660 kg in WY95 while nickel loads increased from 13,700 kg to 1,110,000 kg. Chromium loads also increased markedly from 10,500 kg to 627,000 kg. These data indicate high flow years contribute significantly more metal loads to the Delta when compared to a critically dry year.

In an effort to determine if similar load patterns emerged with an independent data set, loads were calculated in the same manner using the Sacramento County Ambient Monitoring Program (AMP) data collected during the same water years. The same pattern emerged when WY94 and WY95 were compared but, with the exception of cadmium, the magnitude of increased loads for WY95 was lower than those estimated for this study (Table 45). A similar pattern of lower load prediction for most metals was found when estimates for each method (e.g., average concentration and model) were compared (Table 45). For example, load calculations using the Ambient Monitoring Program data ranged from 18% to 102% of estimates in this study. As with the metal concentration comparisons among these two studies, much of the difference can be attributed to the frequency of sample collection. Sampling frequency for this study was much greater than that of the AMP due to the programmatic questions each study addressed. The increased sample frequency in this study resulted in samples which were collected across a wider spectrum of flow conditions within the time period of interest, which is important for accurate predictions of loads.

Metal loads were calculated for the lower Sacramento River and Yolo Bypass during high flow to using the BPTCP data to characterize the contribution differences between these two sources of Delta water. Since the regression relationships between total recoverable metal concentrations and flows were not significant for WY95, comparisons between the two sources were based on

the AC method. Bypass water carried between 48% and 82% of the total load of the measured metals whereas the Sacramento River contributed between 18% and 52% (Table 44). Combined loads for these two sources varied from 3,210 kg of cadmium to 1,120,000 kg and 1,110,000 kg of zinc and nickel, respectively. Dividing loads by the number of days from January to April provides an estimate of the average daily load entering the Delta from the Sacramento River Watershed during high flow conditions. Average daily loads of cadmium, zinc, and nickel which entered the Delta from January through April of 1995 was estimated at 31 kg, 10,700 kg, and 10,700 kg, respectively.

Interesting patterns developed when the load contributions were compared for the lower Sacramento River and Yolo Bypass. Foe and Croyle (1998) estimated the sediment load entering the Delta from the Sacramento River and the Bypass to be 1,300,000 (34%) and 2,500,000 (66%) metric tons, respectively, from January through April 1995. The percentages of copper and zinc from the two sources was nearly identical to those of sediment (Table 44). The Bypass contributed 74% of the chromium as well. These three metals were significantly related to TSS during this water year (see trends in metal concentration section below), indicating that they were either bound to sediment particles diverted into the Bypass or bound to sediment sources within the Bypass. The bulk of nickel loads entering the Delta from the Sacramento River Watershed were carried in the Bypass as well, but this contribution had no relationship to sediment loads. Nickel is common in the geological deposits of the western valley and may enter the Bypass from local sources. Arsenic, cadmium, and lead loads were generally equal in the Bypass and lower Sacramento River.

#### WHAT TRENDS IN METAL CONCENTRATIONS WERE IDENTIFIED?

Metal analyses conducted in this study were essential for assessing exceedances of water quality objectives, performing meaningful toxicity tests, and calculating loads. Another important use for the metals analyses data can be in the determination of relationships between metal concentrations and other water quality and hydrological parameters. The following paragraphs describe relationships which occurred during this study between metal concentrations, flow, and total suspended solids. In addition, some metals seemed to be inter-related, such that high concentrations in one usually coincided in high concentrations in others. These relationships can be useful for determining the best time to collect water quality samples. For example, if certain events (e.g., high flow storm events) can be used to predict when metal concentrations may be among the highest levels for the year in a particular area, monitoring plans can be developed to capture the data of interest by knowing when to expect peak flows. The information is not intended to be used as a predictive tool for metals concentrations in place of actual in-stream monitoring. On the contrary, the information is intended to improve our understanding of when, where, and possibly why we could expect metals concentrations to be high such that appropriate monitoring designs can be developed for future studies.

Four hundred and four water samples were collected from 37 stations for analysis of dissolved and total recoverable metal concentrations (Appendix B). When total recoverable and dissolved

concentrations were independently averaged for all samples collected, a trend of increasing chromium, copper, nickel, and zinc concentrations was observed from WY93 and WY94 to WY95 (Table 46). Clearly, the data are highly variable within each year due to the large spatial and temporal scale of the sampling effort. This typically would result in data which are not significantly different. The data were not analyzed statistically due to large differences in the number of samples collected among years. However, the results indicate that extended periods of unusually high flows can result in marked increases in the average concentration of chromium, copper, nickel, and zinc. Other metals did not exhibit a consistently strong association with peak flows. For example, total recoverable and dissolved arsenic showed a trend of decreasing average concentration from WY94 to WY95. Cadmium, on the other hand, had a distinctly different profile with total recoverable concentrations increasing and dissolved concentrations essentially remaining unchanged during the three water years. Average total recoverable lead concentrations decreased slightly from the WY93 to WY94, then increased by more than three fold in WY95. while the average dissolved concentration increased from WY93 to WY95. It should be noted that averaging the metal analyses for all stations can be problematic because of different sample collection frequencies at each station and different stations monitored among water years. Ideally, statistical analyses of the data would be performed to ascertain if significant relationships existed in the data set. Again, the experimental design employed in this study resulted in great variability about the mean which prohibits the identification of significant relationships. The data should however be used for the basis of follow-up studies which should incorporate a more statistically balanced sampling design.

An analysis of average metal concentrations was performed at Greene's Landing on the lower Sacramento River to determine if the trends among water years held true within a station extensively sampled during the same period. Similar to when concentrations from all stations were averaged, the average total recoverable and dissolved chromium, lead, nickel, and zinc showed a trend of increased concentrations from WY93 to WY94, WY94 to WY95, and WY93 to WY95 (Table 47). Average dissolved concentrations of cadmium behaved in a similar fashion as the entire data set, with no changes among water years. However, average total recoverable cadmium concentrations had a different pattern with a decrease from WY94 to WY95. Average dissolved copper concentrations were also inconsistent with the combined data with no difference between WY93 and WY 94, but matched the trends for the combined data from WY94 to WY95. Arsenic was not measured at Greene's Landing during WY94 and therefore changes during water years could not be compared at this station. With the exception of dissolved cadmium concentrations, the concentration of the monitored metals appear to be closely tied to flow or other parameters related to flow when high flow conditions are compared to normal or drought conditions. However, the reverse trend (e.g., decreased concentrations with decreased flows) does not hold true when comparing drought conditions to normal hydrological conditions.

Dissolved and total recoverable metal concentrations collected from the Sacramento River at Greene's Landing were regressed against each other, flow at Freeport, and total suspended solids (TSS) for WY94, WY95, and combined the WY94 and WY95 (WY94/95) to determine if these factors were interrelated. The number of significant relationships between dissolved metals, total

÷

recoverable metals, flow, and TSS declined from 13 in the critically dry WY94 to eight in the high flow WY95 (Tables 44 and 49). When data from water year 1994 and 1995 were combined, 16 of 35 regression analyses were significant (Table 49).

During the dry WY94, total recoverable concentrations of chromium, copper, lead, nickel, and zinc at Greene's Landing were significantly associated with total suspended solids and flows (Table 48; Figs. 3-12). These significant relationships indicate these metals were bound to suspended sediments. These metal laden suspended sediments are in turn closely associated with flows during this critically dry year, such that the total recoverable metal concentrations increase with increasing flows. Conversely, dissolved chromium, copper, and nickel were also closely tied to flow conditions but did not exhibit significant relationships with total suspended solids (Table 48; Figs. 13-18). Filtration (0.45 $\mu$ m) of samples as done in this study would permit the passage of colloid-associated metals into the dissolved fraction. The lack of significant relationships between dissolved metals and TSS may be due to the presence of other suspended solids in the TSS measurements. Total recoverable metal concentrations could not be used to predict dissolved concentrations due to a lack of significant relationships (Table 48; Figs. 19 & 20). Both total recoverable and dissolved cadmium concentrations were unrelated to flow and TSS, which is consistent with the lack of a trend reported in Tables 46 and 47. Therefore, concentrations of several metals would be expected to increase with increasing flow conditions and/or increased sediment load in the Sacramento River during dry conditions.

These conclusions did not necessarily hold true at Greene's Landing during the wet WY95. Of particular interest is the absence of significant relationships between flows and total recoverable and dissolved metal concentrations in WY95 when compared to WY94 (Tables 48 and 49; Figs. 21-34). When compared to the dry WY94, the breakdown in the relationships in WY95 may be related to, but are not limited to: (1) an increase in tributary input of suspended sediments in the system during this exceptionally wet year; (2) contribution of suspended sediments, flow, and metals from sources further into the watershed; (3) resuspension of deeply buried sediments in the waterways; (4) transportation of larger particles which may have different affinities for metal contaminants than those which occur in the system during dry years; (5) stripping of algae from rocks and transport downstream due to scour during high flows; and (6) flushing of planktonic communities from lakes and rivers during high flow conditions. The major sources of suspended sediments in the lower watershed during a dry water year are the Sacramento, Feather, and American Rivers, whereas smaller tributaries on the western and eastern valley slopes may contribute significantly to the total suspended solids during a wet year. The different geological sources of these sediments may result in different binding affinities for the metals and could therefore disrupt the relationships between total recoverable metals, total suspended solids, and flow. However, this is conjecture at this point and would require further study to clarify the role of small tributary sediments during high flow conditions.

Although the relationships between flow and metal concentrations broke down during high flows found in WY95, total recoverable copper, zinc, and cadmium concentrations at Greene's Landing were still significantly related to TSS indicating these metals are bound to suspended sediment

particles during both dry and wet years (Table 49; Figs. 35-37). The level of significance for this relationship with cadmium ( $R^2 = 0.92$ ) is drastically different than in WY94, again possibly pointing toward further evidence that additional sources of suspended sediments enter the system during high flows (Table 49; Fig. 38). In contrast to WY94, total recoverable and dissolved concentrations for some metals (i.e., copper and lead) were related in WY95 (Table 49; Figs. 39 & 40). Therefore, as dissolved concentrations of lead increased at Greene's Landing, one could predict that total recoverable copper concentrations would increase as well.

Significant relationships between total recoverable copper, zinc, chromium, and nickel at Greene's Landing reemerged again when data from the two water years were combined (Table 50; Figs. 41-48). Consistent with WY94 and WY95, total recoverable concentrations of these metals were significantly associated with suspended sediments and flow for WY94/95 (Table 50; Figs. 41-48). One could apply the relationships between flow and total recoverable concentrations of these metals as a predictive tool. Although the relationships are significant, there is considerable variability about the regression line, especially during high flows (Fig. 46). Therefore, predicting total recoverable concentrations from flow would have a wide margin of error. Dissolved chromjum, lead, and nickel also were significantly related to flow, but only dissolved lead was significantly related to TSS (Table 50; Figs. 49-54). This finding indicates the dissolved forms of chromium and nickel increased over the sampling period with increasing flow, but the metals were not significantly related to suspended sediments. Dissolved chromium and lead were associated with the total recoverable form. This relationship was also significant for copper and nickel, but the dissolved forms of these two metals were not associated with suspended sediments. Therefore, the relationships among dissolved concentration, total recoverable concentration, flow, and TSS are often metal dependent, different when extreme water years are compared and when water years are combined. Additional research would be required to determine if consistent relationships occurred during dry and wet years and blind studies may be necessary to determine the accuracy of using these relationships as a predictive tool for metal concentrations in the Sacramento River.

Relationships found between flow, TSS, and metals during this study should not be applied to times of the year other than when winter flows occur because the relationships may not apply. For example, the Sacramento County's Ambient Monitoring Program (AMP) collected similar concentration and flow data throughout the year from the Sacramento River about eight miles upstream of Greene's Landing (Larry Walker & Associates, 1996). Many of the relationships between flow, TSS, and metals were not significant during the dry WY94 (Table 51), indicating the relationships reported during winter flows do not necessarily hold true at other times of the year. However, relationships between TSS and total recoverable copper, zinc, chromium, and cadmium held true during WY95 for both sampling efforts (Table 52). When water years were combined for both data sets, little overlap in significant relationships between metals, flow, and TSS occurred (Tables 50 and 53). These contrasting data sets provide a good example of the differences which may be encountered during environmental monitoring with two different approaches: a systematic sampling effort with samples collected approximately every two weeks versus a program with samples collected many times during set events.

In comparing individual metals to flows and TSS, some associations were apparent (e.g., total recoverable copper, zinc, chromium, and nickel were associated with flow and TSS at Greene's Landing in the combined WY94 and WY95). To better understand this grouping of metals, total recoverable concentrations of each metal was plotted against other individual metals for individual and combined water years (Tables 54; Figs. 55-65, 67-75, & 80-89). During the dry WY94, significant relationships existed between total recoverable copper and chromium, lead, nickel and zinc (Tables 54; Figs. 55-58). Zinc was also significantly related to chromium, lead, and nickel (Tables 54; Figs. 59-6!). When all of the combinations of metal relationships were examined, copper, chromium, lead, zinc, and nickel appeared to be inter-related (Tables 54; Figs. 62-64). Interestingly, these metals were all significantly related to flow and TSS during this water year (Table 48). Flow and TSS were also significantly related to each other during WY94 and seemed to track closely track each other (Figs. 65 & 66). Cadmium was the only metal which did not have significant relationships with the other metals or flow and TSS. It would appear that TSS or flow could be used to formulate rough predictions of copper, chromium, lead, nickel, and zinc concentrations during the drought-like conditions in WY94. Furthermore, these metals would be expected to track each other very closely such that high zinc concentrations could be used to predict high copper, chromium, lead, and nickel concentrations.

A different pattern emerged at Greene's Landing during the wet WY95: cadmium, chromium, copper, nickel, and zinc were inter-related and lead was not associated with any other metal (Table 54; Figs. 67-75). Although none of these metals had significant relationships with flow during this water year, copper, zinc, chromium, and cadmium were significantly related to TSS (Table 49). Furthermore, the relationship between flow and TSS was not significant during WY95 (Fig. 76). This could be explained by several outlier points on the plot. Three low flow and low TSS values occurred at the beginning of January 1995 (Fig. 76). This was followed by a first flush event with high flows, precipitation, and TSS (Figs. 76-78). This high TSS pulse followed a peak of almost three inches in rainfall which was then followed by peak flows of nearly 100,000 CFS (Figs. 77-78). Conditions prior to, and including the pulsed event, appeared to cause the breakdown in the relationship between flow and TSS during WY95. Therefore, the data points were removed and the data was re-plotted resulting in a significant relationship (Fig. 79). The rapid changes in flow conditions induced by heavy rainfall could explain the lack of relationships between flow and the grouped metals. Using the inter-related nature of TSS and the grouped metals (i.e., copper, zinc, chromium, cadmium, and nickel), one could begin to predict high TSS concentrations would result in high concentrations of these metals during periods of very high flows.

Copper, zinc, chromium, lead, and nickel were again inter-related when both water years were combined (Table 54; Figs. 80-89). With the exception of lead, these metals were again significantly related to flow and TSS (Table 50). In addition, flow and TSS were also significantly related (Fig. 90). As illustrated for WY94, TSS and/or flow would appear to be useful in predicting concentration of copper, zinc, chromium, and nickel. Clearly, however, further study would be required to determine how accurate such predictions would be.

Furthermore, these relationships vary with water year as is apparent for WY94 and WY95 and should only be applied to different water years for the purpose of testing the "goodness of fit" of the relationship under different hydrological conditions. A more appropriate use of these relationships is in the design of monitoring plans for metals. For example, if a study is designed to quantify metals when concentrations are high, the relationships above indicate knowledge of flow conditions in the river can be used to optimize sampling such that concentrations would be expected to be high.

# WHAT SOURCE(S) OF METALS WERE IDENTIFIED DURING THE METALS SOURCE PILOT STUDY?

Given that concentrations of many metals peaked with high flow conditions, a special pilot study was undertaken to track sources of metal loads up the Sacramento River Watershed during one of the largest storms of the year in 1995. Due to the limited budget for the study and the focus on metal loads, analyses were performed for total recoverable concentrations only. Samples were not collected for the determination of toxicity or exceedances of water quality objectives (e.g., dissolved metal analyses). Although the objective of the pilot study was to track sources of metal loads during a high flow event, the data could not be used to quantify the load contribution from mines in the area of Lake Shasta and Keswick Reservoir because discharges from the reservoirs were maintained at low levels to minimize downstream flooding. This resulted in samples downstream of the reservoirs which were negligibly affected by runoff from mines. However, some previously reported and some unknown sources of metals were identified during the study. A complete description of the results from this study can be found in Appendix D.
## ACKNOWLEDGMENTS

The authors wish to thank the U.S. Bureau of Reclamation for use of their monitoring facility at Greene's Landing. We would like to thank Linda Deanovic, Kristy Cortright, Karen Larsen, Emilie Reyes, and Tom Kimball of the U.C. Davis Aquatic Toxicology Laboratory for conducting the toxicity tests. Mark Stephenson, Jon Goetzl, and the staff of the California Department of Fish and Game's Mussel Watch Laboratory were instrumental in conducting the metal analyses for this project. Metal load estimates for the Sacramento County Ambient Monitoring Program were calculated by Claus Suverkropp and made available for this report. The authors would also like to thank Charlie Alpers, Joseph Domagalski, Brian Finlayson, Tom Grovhoug, Dennis Heiman, Mark Stephenson, Rick Sugarek, and Claus Suverkropp for their detailed and thoughtful review of draft forms of this report.

Į

## SUMMARY OF RECOMMENDATIONS

1. Continue to rely on the metal analysis protocols and QA/QC guidelines implemented in this project for determining metal concentrations in the surface waters of the Central Valley.

2. Continue using the US EPA Three Species Assays to identify toxicity in field samples. However, findings from a comprehensive literature search indicate other species may be more sensitive to metals. If future biomonitoring studies indicate a species is in decline in the Delta, efforts should be made to determine if the species could be affected by ambient metal concentrations.

3. Conduct a special study to determine if there is a problem with accumulation of metals in the tissues of aquatic organisms, and determine if bioaccumulation is/is not resulting in biomagnification. If accumulation is occurring, determine if the high total loads measured during wet years, such as WY95, play a role in any identifiable bioaccumulation problem.

4. Relative to other sources, determine the contribution of mines, urban runoff, and agricultural discharges on the overall metal loads entering the Delta. Included in this study should be a description of how the contribution varies seasonally and with major storm events.

5. Ambient monitoring programs such as the Coordinated Monitoring Program, Regional Monitoring Program, Sacramento River Watershed Program, and CALFEDs Coordinated Monitoring and Research Program continue to include water column metals monitoring and incorporate sediment testing and tissue analyses.

6. Additional recommendations specific to the Metals Source Pilot Study can be found in Appendix D. Several metals appear to be closely associated with suspended sediment particles. Special studies should initiated to determine if erosion controls can reduce suspended sediment and total recoverable metal concentrations in regions which were sources of high suspended sediment and metal concentrations during the study.

24

## Literature Cited

Abbassi, S.A., P.C. Nipaney, and R. Soni. 1985. Environmental consequences of the inhibition of the hatching of pupae of <u>Aedes aegypti</u> by mercury, zinc, and chromium - The abnormal toxicity of zinc. Int. J. Environ. Stud. 24: 107-114.

Academy of Natural Sciences. 1960. The sensitivity of aquatic life to certain chemicals commonly found in industrial wastes. Philadelphia, Pennsylvania.

Adams, E.S. 1975. Effects of lead and hydrocarbons from snowmobile exhaust on brook trout (*Salvelinus fontinalis*). Trans. Am. Fish. Soc. 104(2):363-373

Alkahem, H.F. 1994. The toxicity of nickel and the effects of sublethal levels on haemetological parameters and behavior of the fish, *Oreochromis niloticus*. J. Univ. Kuwait Sci. 21(2):243-251

Allen, P. 1993. Changes in tissue GSH concentrations as indicators of acute cadmium or lead toxicity. Fresenius Environ. Bull. 2(10):582-587.

Al-Sabti, K., M. Franko, B. Andrijanic, S. Knez and P. Stegnar. 1994. Chromium induced micronuclei in fish. J. Appl. Toxicol. 14(5):333-336

Anderson, R.L., C.T. Walbridge, and J.T. Fiandt. 1980. Survival and growth of <u>Tanytarsus dissimilus</u> (Chironomidae) exposed to copper, cadmium, zinc, and lead. Arch. Environ. Contam. Toxicol. 9: 329-335.

Andros, J.D. and R.R. Garton. 1980. Acute lethality of copper, cadmium, and zinc to northern squawfish. Trans. Am. Fish. Soc. 109: 235.

Angandi, S.B. and P. Mathad. 1994. Effect of chromium and nickel on *Scenedesmus quadricauda*. Phykos 33(1/2):99-103

Applegate, V.C., J.H. Howell, A.E. Hall Jr. and M.A. Smith. 1957. Toxicity of 4,346 chemicals to larval lampreys and fishes. Spec. Sci. Rep.-Fish. No. 207, U.S. Fish Wildlife Service, U.S.D.I, Washington, D.C.:157 p.

Arthur, J.W. and E.N. Leonard. 1970. Effects of copper on <u>Gammarus pseudo-limnaeus</u>, <u>Physa integra</u>, and <u>Campeloma decisum</u> in soft water. Jour. Fish. Res. Board Can. 27: 1277.

Azeez, P.A. and D.K. Banerjee. 1987. Influence of light on chlorophyll A content of blue-green algae treated with heavy metals. Bull. Environ. Contam. Toxicol. 38(6):1062-1069

)

Bailey, H.C., C. DiGiorgio, L. Deanovic, and D.E. Hinton. 1994. Master Contract North Valley Study. Quarterly Report prepared for the California State Water Resources Control Board. March 1994.

Bailey, H.C., C.L. DiGigorgio, K. Kroll, G. Starrett, M. Miller, and D.E. Hinton. 1996. Development of procedures for identifying pesticide toxicity in effluents and ambient waters: carbofuran, diazinon, chlorpyrifos. Environmental Toxicology and Chemistry. 15(6): 837-845.

Bales, S.S., M. Letourneau, A. Tessier, & P.G.C. Campbell. 1983. Variation in zinc adsorption and transport during growth of *Chlamydomonas variabilis* (Chlorophyceae) in batch culture with daily addition of zinc. Can. J. Fish. Aquat. Sci. 40: 895-904

Bartlett, L., F.W. Rabe, W.H. Funk. 1974. Effects of copper, zinc, and cadmium on *Selenastrum capricornutum*. Water Res. 8: 179-185

Becker, C.D. and M.G. Wolford. 1980. Thermal resistance of juvenile salmonids sublethally exposed to nickel, determined by the critical thermal maximum method. Environ. Pollut. 21(30):181-189

Belanger, S.E., J.L. Farris, & D.S. Cherry. 1989. Effects of diet, water hardness, and population source on acute and chronic copper toxicity to *Ceriodaphnia dubia*. Arch. Environ. Contam. Toxicol. 18(4): 601-611.

Benoit, D.A. 1976. Toxic effects of hexavalent chromium on brook trout *(Salvelinus fontinalis)* and rainbow trout *(Salmo gairdneri)*. Water Res. 10: 497

Benoit, D.A., and G.W. Holcombe. 1978. Toxic effects of zinc on fathead minnows Pimephales promelas in soft water. J. Fish. Biol. 13: 701-708.

Berglind, R., G. Dave and M.L. Sjobeck. 1985. The effects of lead on deltaaminolevulinic acid dehydratase activity, growth, hemoglobin content and reproduction in *Daphnia magna*. Ecotoxicol. Environ. Safety 9(2):216-229

Bertram, P.E. and B.A. Hart. 1979. Longevity and reproduction of <u>Daphnia pulex</u> (deGeer) exposed to cadmium-contaminated food or water. Environ. Pollut. 19: 295.

Biegert, E.K. and V. Valkovic. 1980. Acute toxicity and accumulation of heavy metals in aquatic animals. Period. Biol. 82(1):25-31.

Birge, W.J. 1978. Aquatic toxicology of trace elements of coal and fly ash. In: J.H. Thorp and J.W. Gibbons (Eds.), Dept. Energy Symp. Ser., Energy and Environmental Stress in Aquatic Systems, Augusta, GA 48:219-240

Birge, W.J. Black, and B.A. Ramey. 1981. The reproductive toxicology of aquatic contaminants. In: Hazard Assessment of Chemicals: Current developments. Vol. 1. Saxeena, J. and F. Fisher (Eds.). Academic Press, New York, NY pp.59-115

Birge, W.J., J.A. Black and A.G. Westerman. 1979. Evaluation of aquatic pollutants using fish and amphibian eggs as bioassay organisms. In: S.W. Nielsen, G. Migaki and D.G. Scarpelli. Symp. Animals Monitors Environ. Pollut. 1977, Storrs, CT 12:108-118.

Birge, W.J., J.E. Hudson, J.A. Black, and A.G. Westerman. 1978. Embryo-larval bioassays in inorganic coal elements and in situ biomonitoring of coal waste effluents. In: Symp. U.S. Fish Wild. Serv., Surface Mining Fish Wildlife. Needs in Eastern U.S., W. VA:97-104

Black, M.C., J.R. Ferrell, R.C. Horning and L.K. Martin, Jr. 1996. DNA strand breakage in freshwater mussels (*Anodonta grandis*) exposed to lead in the laboratory and field. Environ. Toxicol. Chem. 15(5):802-808

Blaise, C., R., Legault, N. Beringham, R Van Coillie and P. Vasseur. 1986. A simple microplate algal assay technique for aquatic toxicity assessment. Toxic. Assess. 1:261-281

Blaylock, B.G. and M.L. Frank. 1979. A comparison of the toxicity of nickel to the developing eggs and larvae of carp (*Cyprinus carpio*). Bull. Environ. Contam. Toxicol. 21:604-611

Borgmann, U., O. Kramar and C. Loveridge. 1978. Rates of mortality, growth and biomass production of *Lymnaea palustris* during chronic exposures to lead. J. Fish. Res. Board Can. 35(8):1109-1115

Borgmann, V., W.P. Norwood, I.M. Babirad. 1991. Relationship between chronic toxicity and bioaccumulation of cadmium in *Hyalella azteca*. Can J. Fish Aquat. Sci. 48: 1055-1060

Borgmann, U., W.P. Norwood and C. Clarke. 1993. Accumulation, regulation and toxicity of copper, zinc, lead and mercury in *Hyalella azteca*. Hydrobiologica. 259:79-89

Boutet, C. and C. Chaisemartin. 1973. Specific scientific properties of metallic salts in *Austropotamobius pallipes* pallipes and *Orconectes limosus*. C.R. Soc. Biol. (Paris) 167(12):1933-1938

Braginskiy, L.P. and E.P. Scherban. 1978. Acute toxicity of heavy metals to aquatic invertebrates at different temperatures. Hydrobiol. Jour. 14(6): 76.

Bringmann, G. 1975. Determination of the biologically harmful effect of water pollutants by means of the retardation of cell proliferation of the blue algae *Microcystis*. Gesundheits-Ing. 96:238

Bringmann, G and R. Kuhn. 1959. Water toxicology studies with protozoans as test organisms. Gesund. -Ing. 80:239-242

Bringmann, G and R. Kuhn. 1978. Testing of substances for their toxicity threshold: Model organisms *Microcystis (Diplocystis) aeriginosa* and *Scenedesmus quadricauda*. Mitt. Int. Ver. Theor. Angew. Limnol. 21:275-284

Bringmann, G and R. Kuhn. 1980. A comparison of the toxicity thresholds of water pollutants to bacteria, algae and protozoa in the cell multiplication inhibition test. Water Res. 14(3):231-241

Bringmann, G and R. Kuhn. 1981. Comparison of the effects of harmful substances on flagellates as well as ciliates and on halozoic bacteriophagious and saprozoic protozoa. Gas-wasserfach, Wasser- Abwasser 122:308-313

Bringham, G, R. Kuhn and A. Winter. 1980. Determination of biological damage from water pollutants to protozoa. III. Saprozioc Flagellates. Z. Wasser-Abwasser-Forsch. 13(5):170-173

Brkovic-Popovic, I. and M. Popovic. 1977. Effects of heavy metals on survival and respiration rate of tubificid worms: Part II - effects on survival. Environ. Pollut. 13: 93-98

Buhl, K.J. and S.J. Hamilton. 1990. Comparative toxicity of inorganic contaminants released by placer mining to early life stages of salmonids. Ecotoxicol. Environ. Safety 20(3):325-342

Cain, D.J., J.L. Carter, S.U. Fend, S.N. Luoma, C.N. Alpers and H.E. Taylor. 1998. Metal exposure to a benthic invertebrate, *Hydropsyche californica*, in the Sacramento River downstream of Keswick Reservoir, California. U.S. Geological Survey. Preliminary Report. August 20

Cairns, J. Jr. and A. Scheier. 1968. A comparison of the toxicity of some industrial waste components tested individually and combined. Prog. Fish-Cult. 30:3.

Cairns, J. Jr., and R.E. Sparks. 1971. The use of bluegill breathing to detect zinc. PB211332. National Technical Information Service, Springfield, VA.

Cairns J. Jr. 1978. Effects of temperature on aquatic organism sensitivity to selected chemicals. Bulletin 106. Virginia Water Resources Research Center. Blacksburg, Virginia

Cairns, J. Jr., et al. 1978. Effects of temperature on aquatic organism sensitivity to selected chemicals. Bulletin 106. Virginia Water Resources Research Center, Blacksburg, Virginia.

Calderon Llanten, C.E. and H. Greppin. 1993. Toxicity tests on Zn, Cu and Pb with *Chlorella rubescent*, Chod: using Super(31)P NMR. Arch Aci. 46(2):249-258

California Data Exchange Center. 1998. Accessed by internet at http://cdec.water.ca.gov/ Data maintained by California Department of Water Resources, Division of Flood Control, Sacramento, CA

California Department of Pesticide Regulation. 1993. Pesticide Use Report Annual 1991 Indexed by Chemical. Dept. Pesticide Regulation, Information Systems Branch. Sacramento, CA.

California Department of Pesticide Regulation. 1995. Pesticide Use Report Annual 1993 Indexed by Chemical. Dept. Pesticide Regulation, Information Systems Branch. Sacramento, CA.

Call, D.J. et al. 1983. Toxicity and metabolism studies with EPA priority pollutants and related chemicals in freshwater organisms. PB83-263665. National Technical Information Service, Springfield, VA.

Canterford, G.S. and D.R. Canterford. 1980. Toxicity of heavy metals to the marine diatom, *Ditylum brightwellii* (West): Correlation between toxicity and metal speciation. J. Mar. Biol. Assoc. U.K. 60(1):227-242

Capelo, S., M.F. Vilhena, M.L.S. Simoes Goncalves and M.A. Sampayo. 1993. Effect of lead on the uptake of nutrients by unicellular algae. Water Res. 27(10):1563-1568

Carlson, A.R. and T.H. Roush. 1985. Site-specific water quality studies of the Straight River, Minnesota: Complex effluent toxicity, zinc toxicity, and biological survey relationships. EPA-600/3-85-005. National Technical Information Service, Springfield, VA.

Carlson, A.R., et al. 1982. Cadmium and endrin toxicity to fish in waters containing mineral fibers. EPA-600/3-82-053. National Technical Information Service, Springfield, VA.

Central Valley Regional Water Quality Control Board. 1989. A mass loading assessment of major point and non-point sources discharging to surface waters in the Sacramento Valley, California 1985. Staff report, Standards, Policy and Special Studies Section. Sacramento, CA. March

Central Valley Regional Water Quality Control Board. 1994. Water Quality Control Plan (Basin Plan): Sacramento River and San Joaquin River Basins, Third Edition. Staff Report, Sacramento CA. Chakomakos, C., et al. 1979. The toxicity of copper to cutthroat trout (Salmo clarki) under different conditions of alkalinity, pH, and hardness. Environ. Sci. Technol. 13: 213.

Chapman, G.A. 1975. Toxicity of copper, cadmium and zinc to Pacific Northwest salmonids. U.S. APE, Corvallis, Oregon.

Chapman, G.A. 1982. Letter to Charles E. Stephan. U.S. EPA, Corvallis, Oregon. December 6.

Chapman, G.A. and D.G. Stevens. 1978. Acutely lethal levels of cadmium, copper, and zinc to adult male coho salmon and steelhead. Trans. Am. Fish. Soc. 107: 837.

Christensen, E.R. and P.A. Zielski. 1980. Toxicity of arsenic and PCB to a green alga *(Chlamydomonas)*. Bull. Environ. Contam. Toxicol. 25(1):43-48

Clements, W.H., D.S. Cherry, & J. Cairns, Jr. 1989. The influence of copper exposure on predator-prey interaction in aquatic insect communities. Freshwater Biol. 21: 483-488.

Cohn, T.A., L.L. DeLang, E.J. Gilroy, R.M. Hirsch and D.K. Wells. 1989. Estimating constituent loads. Water Resources Research 22(5): 937-942

Conway, H.L. 1978. Sorption of arsenic and cadmium and their effects on growth, micronutrient utilization, and photosynthetic pigment composition of <u>Asterionella</u> formosa. Jour. Fish. Res. Board Can. 35: 286.

Cowell, B.C. 1965. The effects of sodium arsenite and silvex on the plankton populations in farm ponds. Tran. Am. Fish. Soc. 94:371

Colwell, F.S., S.G. Hornor, D. S. Cherry. 1989. Evidence of structural and functional adaption in epilithon exposed to zinc. Hydrobiologia. 171: 79-90

Connor, V., L. Deanovic, and E. Reyes. 1994. Basin Plan Metal Implementation Plan Development Project - Bioassay Results 1991-1992. Staff Report, Standards, Policies and Special Studies Unit, Central Valley Regional Water Quality Control Board, Sacramento, CA December 1994.

Connor, V., L. Deanovic, H. Nielsen, and H. Bailey. 1995. Quality Assurance Project Plan: Delta Monitoring 1993-1994. Staff Report, Standards, Policies and Special Studies Unit, Central Valley Regional Water Quality Control Board, Sacramento, CA. Cote, R. 1983. Toxic aspect of copper on the biomass and productivity of phytoplankton in the Saguenay River, Quebec. Hydrobiologia. 98: 85.

Daday, A., A.H. Mackerras and G.D. Smith. 1985. The effect of nickel on hydrogen nietabolism and nitrogen fixation in the cyanobacteria *Anabaena cylindrica*. J. Gen. Microbio. 131:231-238

Dave, g. 1984. Effects of copper on growth, reproduction, survival, and hemoglobin in Daphnia magna. Comp. Biochem. Physiol. 78C: 439.

Dave, G. and R. Xiu. 1991. Toxicity of mercury, copper, nickel, lead and cobalt to embryos and larvae of zebrafish, *Brachydanio rerio*. Arch. Environ. Contam. Toxicol. 21:126-134

Davis, P.H, J.P. Goettl Jr., J.R. Sinley and N.F. Smith. 1976. Acute and chronic toxicity of lead to rainbow trout (*Salmo gairdneri*) in hard and soft water. Water Res. 10(3):199-206

Deanovic, L.A., H.C. Bailey, T.W. Shed, and D.E. Hinton. 1996. Sacramento-San Joaquin Delta Bioassay Monitoring Report: 1993-1994. First Annual Report to the Central Valley Regional Water Quality Control Board, Sacramento, CA.

Deanovic, L.A., H.C. Bailey, T.W. Shed, and D.E. Hinton. 1998. Sacramento-San Joaquin Delta Bioassay Monitoring Report: 1994-1995. Second Annual Report to the Central Valley Regional Water Quality Control Board, Sacramento, CA.

DeFoe, D.L. 1982. Arsenic (V) Test Results. U.S. EPA, Duluth MN (Memo to R.L. Spehar, US EPA Duluth MN)

Den Dooren de Jong, L.E. 1965. Tolerance of *Chlorella vulgaris* for metallic and nonmetallic ions. Antonie Leeuwenhoek J. Microbiol. Serol. 31:301-313

De Nicola Guidici, M., L. Migliore, & S.M. Guarino. 1987. Sensitively of *Asellus aquaticus* (L.) and *Proasellus coxalis dollf*. (Crustacea, isopoda) to copper. Hydrobiologia 146: 63-69.

De Nicola Guidici, M., L. Migliore, C. Gambardella, and A Marotta. 1988. Effect of chornic exposure to cadmium and copper on *Asellus aquaticus* (L.) (Crustacea, isopoda) to copper. Hydrobiologia 157(3): 265-269.

Dowden, B.F. and H.J. Bennett. 1965. Toxicity of selected chemicals to certain animals. J. Water Pollut. Control Fed. 37(9):1308-1316

Eaton, J.G., et al. 1978. Metal toxicity to embryos and larvae of seven freshwater fish species - I. cadmium. Bull. Environ. Contam. Toxiolo. 19: 95.

Elder, J.F. and A.J. Horne. 1978. Copper cycles and CuSO4 algicidal capacity in two California lakes. Environ. Manage. 2: 17.

Environment Canada. 1994. Priority Substances List Assessment Report: Chromium and its Compounds. Canadian Environmental Protection Act. Quebec, Canada

Fezy, J.S., D.F. Spencer and R.W. Greene. 1979. The effect of nickel on the growth of the freshwater diatom *Navicula pelliculosa*. Environ. Pollut. 20(2):131-137

Finlayson, B.J. and D.C. Wilson. 1989. Evaluation of lethal levels, release criteria,, and water quality objectives for an acid-mine waste. In: Aquatic Toxicology: Eleventh Volume. ASTM Special Technical Publication 1007. pp. 189-203.

Foe, C. and W.C. Croyle. 1998. Mercury concentrations and loads from the Sacramento River and from Cache Creek to the Sacramento-San Joaquin Delta Estuary. Staff report submitted to the Central Valley Regional Water Quality Control Board, Sacramento, CA

Friebel, M.F., K.L. Markham, S.W. Anderson and G.L. Rockwell. 1995. Water Resources Data California Water Year 1994, Volume 4, Northern Central Valley Basin and the Great Basin from Honey Lake Basin to Oregon State Line. U.S. Geological Survey Water Data Report CA 94-4, 452 p.

Fujimura, R.W., C. Huang and B. Finlayson. 1995. Chemical and toxicological characterization of Kerswick Reservoir sediments. California Department of Fish and Game, Aquatic Toxicology Laboratory Report prepared for the State Water Resource Control Board.

Gachter, R., et al. 1973. Complexing capacity of the nutrient medium and its relation to inhibition of algal photosynthesis by copper. Schweiz. Z. Hydrol. 35: 252.

Garvey, J.E., H.A. Owen, & R.W. Winner. 1991. Toxicity of copper to *Chlamydomonas reinhardtii* (Chlorophyceae) and *Ceriodaphnia dubia* (Crustacea) in relation to changes in water chemistry. Aquat. Toxicol. 21: 157-170

Gendusa, A.C. 1990. Toxicity of chromium and flouranthene from aqueous and sediment sources to selected freshwater fish. Ph. D. Dissertation, University of North Texas:138

Genter, R.B., D.S. Cherry, E.P. Smith & J. Cairns, Jr. 1988A, Attached algal abundance altered by individual and combined treatments of zinc and pH. Environmental Toxicology & Chemistry 7: 723-733.

32

Goetzl, J., M. Stephenson, M. Puckett, G. Ichikawa, K. Paulson, J. Kanihan, and P Browne. 1994. Quality Assurance/Quality Control Document Bay Protection and Toxic Cleanup Program 1993-1994 Delta Metals. Staff Report, California Department of Fish and Game. Prepared for the State Water Resources Control Board.

Goetzl, J., J. Kanihan, M. Stephenson, M. Puckett, G. Ichikawa, and N. Morgan. 1995. Quality Assurance/Quality Control Document Bay Protection and Toxic Cleanup Program 1994-1995 Delta Metals. Staff Report, California Department of Fish and Game. Prepared for the State Water Resources Control Board.

Ghosh, K. and S. Jana. 1988. Effects of combinations of heavy metals on population growth of fish nematode *Spinicauda spinacauda* in aquatic environment. Environ. Ecol. 6(4):791-794

Grande, M. and S. Anderson. 1983. Lethal effects of hexavalent chromium, lead and nickel on young stages of Atlantic salmon (*Salmo salar L.*) in soft water. Vatten 39(4):405-416

Harrison, F.L. et al. 1981. Effects of copper on adult and early life stages of the freshwater clam, <u>Corbicula manilensis</u>. UCRL-52741. National Technical Information Service, Springfield, Virginia.

Harrison, F.L. et al. 1984. The toxicity of copper to the adult and early life stages of the freshwater clam, <u>Corbicula manilensis</u>. Arch. Environ. Contam. Toxicol. 13: 85.

Harry, H.W. and D.V. Aldrich. 1963. The distress syndrome in *Taphius glabratus* as a reaction to toxic concentrations of inorganic ions. Malacologia 192:283-289

Hart, B.A. and B.D. Schaife. 1977. Toxicity and bioaccumulation of cadmium in Chlorella pyrenoidosa. Environ. Res. 14: 401.

Hatakeyama, S. and Y. Sugaya. 1989. A freshwater shrimp (*Paratya compressa improvisa*) as a sensitive test organism to pesticides. Environ. Pollut. 59(4) 325-336.

Hatakeyama, S. and M. Yasuno. 1981. Effects of cadmium on the periodicity of parturation and brood zsize of <u>Moina macrocopa</u> (Cladocera). Environ. Pollut. (Series A) 26: 111.

Helsel, D.R. and R.M. Hirsch. 1992. Statistical Methods in Water Resources. Series: Studies in Environmental Science Vol 49. Elsevier, Amsterdam, 527p.

Herbold, B. and P.B. Moyle. 1989. The Ecology of the Sacramento-San Joaquin Delta: A Community Profile. U.S. Fish and Wildlife Service Biol. Rep. 85(7.22). 106p.

Heumann, H-G. 1987. Effects of heavy metals on growth and ultrastructure of *Chara* vulgaris. Protoplasma 136(1): 37-48.

Hofslagare, O., S. Sjoberg and G. Samuelsson. 1994. The effect of arsenate and arsenite on photosynthesis in *Scenedesmus obliquus*. A potentiometric study in a closed CO2 system. Chem. Spec. Bioavail. 6(4):95-102

Holcombe, G.W., et al. 1984. Methods for conducting snail (<u>Aplexa hypnorum</u>) embryo through adult exposures: effects of cadmium and reduced pH levels. Arch. Environ. Contam. Toxicol. 13: 627.

Horning, W.B. and T.W. Neiheisel. 1979. Chronic effect of copper on the bluntnose minnow, Pimephales notatus (Rafinesque). Arch. Environ. Contam. Toxicol. 8:545.

Hughes, J.S. 1973. Acute toxicity of thirty chemicals to striped bass (<u>Morone saxatilis</u>). Presented at the Western Association of State Game and Fish Comissioners, Salt Lake City, UT. July.

Hutchinson, T.C. 1973. Comparative studies of the toxicity of heavy metals to phytoplankton and their synergistic interactions. Water Pollut. Res. Can. 8:68-90

Hutchinson, T.C. and P.M. Stokes. 1975. Heavy metal toxicity using algal bioassays. In: Water Quality parameters. Barabas, S. (Ed.). ASTM STP 573. American Society for Testing and Materials, Philadelphia, PA. 320-343

Irmer, U., I. Wachholz, H. Schafer and D.W. Lorck. 1986. Influence of lead on *Chlamydomonas reinhardii* Danegard (Volvocales, Chlorophyta): Accumulation, toxicity and ultrastructural changes. Environ. Exp. Bit. 26(2):97-105

Jana, S.R. and N. Bandyopadhyaya. 1988. Effect of heavy metals on some biochemical parameters in the Freshwater fish, *Channa punctatus*. Aquat. Sci. Fish. Abstr.. 18(4):5486-1Q18: Environ. Ecol. 5(3):488-493.

Jana, S. and S.S. Sahana. 1989. Sensitivity of the freshwater fishes *Clarias batrachus* and *Anabas testudineus* to heavy metals. Environ. Ecol. 7(2):265-270

Jones, J.R.E. 1938. The relative toxicity of salts of lead, zinc and copper to the stickleback (*Gasterosteus aculeatus L.*) and the effect of calcium on the toxicity... J. Exp. Biol. 15(3):394-407.

Jones, J.R.E. 1939. The relation between the electrolytic solution pressures of the metals and their toxicity to the stickleback (*Gastrerosteus aculeatus L.*). J. Exp. Bio. 16 (4):425-437

34

Kapu, M.M. and D.J. Schaeffer. 1991. Planarians in toxicology. Responses of asexual *Dugesia dorotocephala* to selected metals. Bull. Environ. Contam. Toxicol. 47(2): 302-307

Keller, A.E. and S.G. Zam. 1991. The acute toxicity of selected metals to the freshwater mussel, *Anodonta imbecilis*. Environ. Toxicol. Chem. 10(4):539-546

Khangarot, B.S. 1991. Toxicity of metals to a freshwater tubificid worm, *Tubifex tubifex* (Muller). Bull. Environ. Contam. Toxicol. 46:906-912

Klass, E., et al. 1974. The effects of cadmium on population growth of the green alga <u>Scendesmus quadricauda</u>. Bull. Environ. Contam. Toxicol. 12: 442.

Klauda, R.J. 1985. Influence of delayed initial feeding on mortality of striped bass larvae exposed to arsenic and selenium. Am. Fish. Soc. Annu. Meeting No. 115:92-93

Kszos, L.A., A.J. Stewart and P.A. Taylor. 1992. An evaluation of nickel toxicity to *Ceriodaphnia dubia* and *Daphnia magna* in a contaminated stream and in laboratory tests. Environ. Toxicol. Chem. 11 (7):1001-1012

Kumar, D., and H.D. Kumar. 1985. Heavy metal toxicity in the Cyanobacterium *Nostoc linckia*. Aquat. Bot. 22(2):101:105

Lalande, M. and B. Pinel-Alloul. 1986. Acute toxicity of cadmium, copper, mercury and zinc to <u>Tropocyclops prasinus mexicanus</u> (Cycolpoida, Copepoda) from three Quebec lakes. Environ. Toxicol. Chem. 5: 95-102.

Lanzer-DeSouza, M.E. and N.M.M. DaSilva. 1988. Influence of pollutants on aquatic crustacean *Decapoda palamonidae*. Iheringia Ser. Misc. 2:13-30 (Spanish, English abstr.)

Larry Walker Associates. 1996. Sacramento Coordinated Water Quality Monitoring Program 1995 Annual Report. Prepared for the Sacramento Regional County Sanitation District, Sacramento County Water Agency and the City of Sacramento.

Laube, V.M., C.N. McKenzie and D.J. Kushner, 1980. Strategies of response to copper, cadmium and lead by blue-green and a green alga. Can. J. Microbiol. 26(11):1300-1311

Lawrence, S.G., M.H. Holoka 7 R. D. Hamilton. 1989. Effects of cadmium on a microbial food chain, *Chlamydomonas reinhardtii* and *Tetrahymena vorax*. Sci Total Environ. 87/88: 381-395.

Lilius, H., T. Hastbacka and B. Isomaa. 1995. A comparison of the toxicity of 30 reference chemicals to *Daphnia magna* and *Daphnia pulex*. Environ. Toxicol. Chem. 14(12):2085-2088

Lillebo, H.P., S. Shaner, D.Carlson, N. Richard, and P. DuBowy. 1988. Criteria based on direct toxicity to aquatic organisms. In: Water Quality Criteria for Selenium and Other Trace Elements for Protection of Aquatic Life and its Uses in the San Joaquin Valley.

Lind, D., K. Alto, and S. Chatterton. 1978. Regional copper study. Draft Report, Minnesota Environmental Quality Board St. Paul, MN 54p.

Maeda, S., R. Inoue, T. Kozono, T. Tokuda, A. Ohki and T. Takeshita. 1990. Arsenic metabolism in a freshwater food chain. Chemosphere 20(1-2):101-108

Mao, S. and C. Wang. 1990. The effect of some pollutants on SCE of grass carp cells. Ocean. Limnol. Sin./ Haiyang Yu Huzhao 21(3):205-211

Markham, K.L., S.W. Anderson, G.L. Rockwell, and M.F. Friebel. 1996. Water Resources Data California; Water Year 1995, Volume 4, Northern Central Valley Basin and the Great Basin from Honey Lake Basin to Oregon State Line. U.S. Geological Survey. Water Data Report CA 95-4, 428 p.

Marr, J.C.A, H.L. Bergman, J. Lipton, and C. Hogstrand. 1995. Differences in relative sensitivity of naive and metals-acclimated brown and rainbow trout exposed to metals representative of the Clark Fork River, Montana. Can. J. Fish. Aquat. Sci. 52: 2016-2030.

Marshack, J.B. 1995. A Compilation of Water Quality Goals. Staff Report submitted to the California Regional Water Quality Control Board, Central Valley Region. Sacramento, CA April 1996.

Marshack, J.B. 1998. A Compilation of Water Quality Goals. Staff Report submitted to the California Regional Water Quality Control Board, Central Valley Region. Sacramento, CA March 1998.

Marshall, J.S. 1978. Population dynamics of <u>Daphnia galeata mendotae</u> as modified by chronic cadmium stress. Jour. Fish. Res. Board Can. 35: 461.

Marshall, J.S., D.L. Mellinger, and J.I. Parker. 1981. Combined effects of cadmium and zinc on a Lake Michigan zooplankton community. Int. Assoc. Great Lakes Res. 7: 215-223.

McGreachy, S.M. and D.G. Dixon, 1990. Effect of temperature on the chronic toxicity of arsenate to rainbow trout (*Oncorhynchus mykiss*). Can J. Fish. Aquat. Sci. 47(11):2228-2234

McKim, J.M. et al. 1978. Metal toxicity to embryos and larvae of eight species of freshwater fish - II. copper. Bull. Environ. Contam. Toxicol. 19: 608.

McLeay, D.J. 1975. Sensitivity of blood cell counts in juvenile coho salmon (<u>Oncorhynchus kisutch</u>) to stressors including sublethal concentrations of pulp mill effluent and zinc. J. Fish. Res. Board Can. 32: 2357-2364.

Meindl, U. and G. Roderer, 1990. Influence of inorganic and triethyl lead on nuclear migration and ultrastructure of *Micrasterias*. Ecotoxicol. Environ. Safety 19(2):192-203

Meyer, W., G. Harisch and A.N. Sagredos. 1986. Biochemical and histochemical aspects of lead exposure in dragonfly larvae (*Odonata: Anisoptera*). Ecotoxicol. Environ. Safety.11(3):308-319

Meyer, W., M. Kretschmer, A. Hoffmann and G. Harisch. 1991. Biochemical and histochemical observations on effects of low level heavy metal load (lead, cadmium) in different organ systems of the freshwater crayfish, *Astacus astacus*. Ecotoxicol. Environ. Safety. 21(2):137-156

Migliore, L., and M.De Nicola Giudici. 1990. Toxicity of heavy metals to *Asellus aquaticus* (L.) (Crustacea, Isopoda). Hydrobiologica 203(3):155-164

Monahan, T.J. 1976. Lead inhibition of chlorophycean microalgae. J. Phycol. 12:358

Montoya, B.L. F.J. Blast and G.E Hans. 1988. A Mass Loading Assessment of Major Point and Non-point Sources Discharging to Surface Waters in the Central Valley. Draft Staff Report, Standards, Policies and Special Studies Unit, Central Valley Regional Water Quality Control Board. Sacramento, California.

Montoya, B.L. and X. Pan. 1992. Inactive Mine Drainage in the Sacramento Valley, California. Staff Report, Standards, Policies and Special Studies Unit, Central Valley Regional Water Quality Control Board. Sacramento, California.

Mount, D.I. 1982. Memorandum to Charles S Stephen. U.S. EPA, Duluth, Minnesota. June 7

Moore, M.V. & R.W. Winner. 1989. Relative sensitivity of *Ceriodaphnia dubia* laboratory tests and pond communities of zooplankton and benthos to chronic copper stress. Aquatic Toxicol. 15: 311-330.

Mount, D.I. 1982. Memorandum to Charles E. Stephan. U.S. EPA, Duluth, Minnesota. June 7.

Mount, D.I. and T.J. Norberg. 1984. A seven-day life-cycle cladoceran toxicity test. Environ. Toxicol. Chem. 3(3):425-434

Mullen, J.R., M.F. Friebel, K.L. Markham and S.W. Anderson. 1994. Water Resources Data California; Water Year 1993, Volume 4, Northern Central Valley Basins and the Great Basin from Honey Lake Basin to Oregon State Line. U.S. Geological Survey. Water Data Report CA-93-4, 437 p.

Munzinger, A. & M.-L. Guarducci. 1988. The effect of low zinc concentrations on some demographic parameters of *Biomphalaria glabrata* (Say), mollusca:gastropoda. 12:51-61

Munziger, A. 1990. Effects of nickel on *Daphnia magna* during chronic exposure and alterations in the toxicity to generations pre-exposed to nickel. Water Res. 24(7):845-852

Munzinger, A. & F. Monicelli. 1991. A comparison of the sensitivity of three *Daphnia* magna populations under chronic heavy metal stress. Ecotoxicol. Environ. Safety 22: 24-31

Naddy, R.B., T.W. LaPoint and S.J. Klaine. 1995. Toxicity of arsenic, molybdenum and selenium combinations to *Ceriodaphnia dubia* 

Nakagawa, H., T. Sato and H. Kubo. 1995. Evaluation of chronic toxicity of water lead to carp *Cyprinus carpio* using its blood 5-aminolevulinic acid dehydratase. Fish. Sci. 61(6):956-959

Nebeker, A.V., et al. 1984. Effects of copper, nickel and zinc on the life cycle of the caddisfly <u>Clistoronia magnifica</u> (Limnephilidae). Environ. Toxicol. Chem. 3: 645.

Nebeker, A.V., C. Savonen and D.G. Stevens. 1985. Sensitivity of rainbow trout early life stages to nickel chloride. Environ. Toxicol. Chem. 4: 233-239

Nichols, J.W., G.A. Wedemeyer, R.L. Mayer, W.W. Dickhoff, S.V. Gregory, W.T. Yasutake et al. 1984. Effects of freshwater exposure to arsenic trioxide on the parr-smolt transformation of coho salmon (*Oncorhynchus kisutch*). Environ. Toxicol. Chem. 3:143-149

Nielsen, H., L. Deanovic, and H. Bailey. 1995. Quality Assurance Project Plan: Delta Monitoring 1994-1995. Staff Report, Standards, Policies and Special Studies Unit, Central Valley Regional Water Quality Control Board, Sacramento, CA Nordstrom, D., F. Jenne and R. Avereth. 1977. Heavy metal discharges into Shasta Lake and Keswick Reservoirs on the upper Sacramento River, California: A reconnaissance during low flow. U.S. Geological Survey. Water Resources Investigations 76-49. Menlo Park, California. March

Novak, A., B.S. Walters and D.R.M. Passion. 1980. Toxicity of contaminants to integrate food organisms. Prog. Fish. Res. 1980, Great Lakes Fish. Lab., U.S. Fish Wild. Serv., Ann Arbor, MI:2 p.

Oladimeji, A.A., S.U. Qadri and A.S.W. De Freitas. 1984. Measuring the elimination of arsenic through the gills of rainbow trout (*Salmo gairdneri*) by using a two-compartment respirometer. Bull. Environ. Contam. Toxicol. 32(6):661-668

Olson, P.A. 1958. Comparative toxicity of Cr (VI) and Cr (III) in salmon. Hanford Biol. Res. Annu. Rep. #HW-53500, 1957:215-218.

Palowski, C.J., J. B. Hunn, and F.J. Dwyer. 1985. Sensitivity of young stripped bass to organic and inorganic contaminants in fresh and saline waters. Trans. Am, Fish Soc. 114:748-753

Pandey, G.N. and Nisha. 1984. Effect of hexavalent chromium on freshwater fish. C.A. Sel-Environ. Pollut. 3(100):2

Pardue, W.J. and T.S. Wood. 1980. Baseline toxicity data for freshwater bryozoa exposed to copper, cadmium, chromium and zinc. J. Tennessee Acad. Sci. 55:27

Patrick, R. et al. 1968. The relative sensitivity of diatoms, snails, and fish to twenty common constituents of industrial wastes. Prog. Fish. Cult. 30:137

Paulauskis, J.E. & R.W. Winner. 1988 Effects of water hardness and humic acid on zinc toxicity to *Daphnia magna* Straus. Aquat. Toxicol. 12: 273-290

Pawlaczyk-Szpilowa, M., M. Moskal and J. Weretelnik, 1972. [The usefulness of biological tests for...]. Acta Hydrobiol. 14(2):115-127

Pearlmutter, N.L. and M.A. Buchheim. 1983. Copper susceptibility of three growth stages of the green alga <u>Haematococcus</u>. PB83-25678. National Technical Information Service, Springfield, Virginia.

Peterson, R.H., et al. 1983. Effects of cadmium on yolk utilization, growth, and survival of Atlantic salmon alevins on newly feeding fry. Arch. Environ. Contam. Toxicol. 12: 37.

Pietilainen, K. 1975. Synergistic and antagonistic effects of lead and cadmium on aquatic primary production. In: Int. Conf. on Heavy Metals in the Environment, Symp. Proc., Vol 2(2), Institute for Environmental Studies, University of Toronto, Ont., Canada: 861-873

Pickering, Q.H. and C. Henderson. 1964. The acute toxicity of some heavy metals to different species of warm water fishes. Proc. 19th Ind. Waste Conf., Purdue University, West Lafayette, IN: 578-591

Pierson, K.B. 1981. Effects of chronic zinc exposure on the growth, sexual maturity, reproduction, and bioaccumulation of the guppy, <u>Poecilia reticulata</u>. Can. J. Fish. Aquat. Sci. 38: 23-31.

Pillard, D.A., P.M. Rocchio, K.M. Cassidy, S.M. Stewart and B.D. Lance. 1987. Hexavalent chromium effects on carbon assimilation in *Selenastrum capricornutum*. Bull. Environ. Contam. Toxicol. 38(4):725-721

Planas, D. and F.P. Healey. 1978. Effects of arsenate on growth and phosphorus metabolism of phytoplankton. J. Phycol. 14(3):337:341

Prasad, P.V.D. and Y.B.K. Chowdary. 1981. Effects of metabolic inhibitors on the calcification of a freshwater green alga, *Gloeotaenium ioitlesbergarianum*. Hansgirg 1. Ann. Bot. 47(4):451-459

Rabe, F.W. and C.W. Sappington. 1970. Biological productivity of the Coeur d'Alene Rivers as related to water quality. Project A-024-Ida. Water Resources Research Institute, Moscow., ID.

Rachlin, J.W. and M. Farran. 1974. Growth responses of the green algae <u>Chlorella</u> <u>vulgaris</u> to selective concentrations of zinc. Water. Res. 8: 575-577.

Rachlin, J.W., T.E. Jensen, and B. Warkentine. 1982. The growth response of the green alga (<u>Chlorella saccharophila</u>) to selected concentrations of the metals Cd, Cu, Pb, and Zn. In: Trace substances in environmental health-XVI. Hemphill, D.D. (Ed.). University of Missouri, Columbia, MO. pp. 145-154.

Rachlin, J.W., T.E. Jensen, and B. Warkentine. 1983. The growth response of the diatom <u>Navicula incerta</u> to selected concentrations of the metals: cadmium, copper, lead and zinc. Bull. Torrey Bot. Club. 110: 217-223.

Rai, U.N. and P. Chandra. 1989. Removal of heavy metals from polluted waters by *Hydrodictyon reticulatum* (Linn.) lagerheim. Sci. Total Environ. 87/88:509-515

Rai, U.N. and P. Chandra. 1992. Accumulations of copper, lead, manganese and iron by field populations of *Hydrodictyon reticulatum* (Linn.) lagerheim. Sci. Total Environ. 116(3):203-211

Reish, D.J. and T.V. Gerlinger. 1984. The effects of cadmium, lead and zinc on survival and reproduction in the Polychaetous annelid *Neanthes arenaceodentata* (F. Nereididae). In: P.A. Hutchings (Ed.), Proc. of the First Int. Polychaete Conf., Sydney, Aust. July 1983. The Linnean Society of New South Wales, Aust. 383-389.

Reuther, R. 1992. Arsenic introduced into a littoral freshwater model ecosystem. Sci. Total. Environ. 115(3):219-237

Reyes, E. 1994. A review of copper, cadmium, and zinc toxicity in aquatic organisms. Staff Report, Standards, Policies and Special Studies Unit, California Regional Water Quality Control Board, Central Valley Region, Sacramento, CA. October 1994.

Richter, J.E. 1982. Memorandum to Charles E. Stephan. U.S. EPA Duluth, MN. June 30.

Sacramento Regional County Sanitation District. 1996. Industrial waste Pretreatment Program: 1995. Annual Report. March

Saiki, M.K., D.T. Castlebury, T.W. May, B.A. Martin and F.M. Bullard. 1995. Copper, cadmium and zinc concentrations in aquatic food chains from the upper Sacramento River (California) and selected tributaries. Arch. Environ. Contam.Toxicol. 29: 484-491.

Sakaguchi, T., et al. 1977. Uptake of copper by <u>Chlorella regularis</u>. Nippon Nog. Kag. Kaishi. 51: 497.

anders, H.O. and O.B. Cope. 1966. Toxicities of several pesticides to two species of cladocerans. Trans. Am. Fish. Soc. 95:165

か

San Francisco Estuary Project. 1992. State of the Estuary: A report on conditions and problems in the San Francisco Bay/Sacramento-San Joaquin Delta Estuary. Report obtained from the San Francisco Bay Regional Water Quality Control Board, Oakland, CA.

Sathya, K. S. and K.P. Blakrishnan, 1987. Physiology of phytoplankton in relation to metal concentration. Water Air Soil Pollution 38(3-4): 283-297

Sauter, S., K.S. Buxton, K.J. Macek and S.R. Petrocelli. 1976. Effects of exposure to heavy metals on selected freshwater fish: toxicity of copper, cadmium, chromium and lead to eggs and fry of seven fish species. EPA-600/3-76-105, Environ. Res. Lab., US EPA Duluth, MN: 74p.

Shabana, E.F., A.F. Dowidar, I.A. Kobbia and S.A. El-Attar, 1986. Studies on the effects of some heavy metals on the biological activities of some phytoplankton species. II. The effects of some metallic ions. Egypt. J. Physiol. Sci. 13(1/2):55-71

Sharma, G.P., R.C. Soboti, A. Chaudhru and K.K. Ahluwalia. 1988. Genotoxicity of two heavy metal compounds- lead acetate and mercuric chloride in the mosquito, *Anopheles stephensi* Liston (Culicidae: Diptera). Cytologia (Tokyo) 53(2):263-267

Shavrina, O.B. and L.D. Gapochka, 1984. Effect of some heavy metals on the growth of the bluegreen algae *Synechocystsis aquatilis* in culture. C.A. Sel.-Environ. Pollut. 3(100):3

Shaw, T.L. and V.M. Brown. 1971. Heavy metals and fertilization of rainbow trout eggs. Nature(London) 230(5291):251

Shiau, S.Y. and S.F. Lin. 1993. Effect of supplemental dietary chromium and vandium on the utilization of different carbohydrates in tilapia, *Oreochromis niloticus* (cross) *O. aureus*. Aquaculture 110 (3/4)"321-330.

Shukla, J.P., U.N. Dwivedi, P. Tewari and M. Prasad. 1985. Deleterious effects of arsenic on the nucleic acids and protein metabolism in the liver of *Heteropneustes fossilis* (Bl.) - a freshwater teleost. Acta Hydrochim. Hydrobiol. 13(5):611-614

Singhal, K.C. 1994. Biochemical and enzymatic alterations due to chronic lead exposure in the freshwater catfish, *Heteropneustes fossilis*. J. Environ. Biol. 15(3): 185-191

Slabbert, J.L. and J.P. Maree. 1986. Evaluation of interactive toxic effects of chemicals in water using a *Tetrahymena pyriformis* toxicity screening test. Water S. A. 12(2):57-62

Slabbert, J.L. and W.S.G. Morgan. 1982. A bioassay technique using *Tetrahymena pyriformis* for the rapid assessment of toxicants in water. Water Res. 16(5):517-523

Sorensen, E.M.B. 1976. Toxicity and accumulation of arsenic in green sunfish, *Lepomis cyanellus*, exposed to arsenate in water. Bull. Environ, Contam. Toxicol. 15(6):756-761

Soundrapandian, S. and K. Venkataraman. 1990. Effect of heavy metal salts on the life history of *Daphnia similis* Claus (Crustacea: Cladocera). Proc. Indian Acad. Sci. Anim. Sci. 99(5): 411-418.

42

Sparks, R.E., W.T. Waller, and J. Cairns. 1972. Effects of shelters on the resistance of dominant and submissive bluegills (<u>Lepomis macrochirus</u>) to a lethal concentration of zinc. J. Fish. Res. Board Can. 29: 1356-1358.

Spehar, R.L. 1976a. Cadmium and zinc toxicity to Jordanella floridiae. EPA-600/3-76-096. National Technical Information Service, Springfield, VA.

Spehar, R.L. 1976b. Cadmium and zinc toxicity to flagfish, Jordanella floridiae. J. Fish. Res. Board Can. 33: 1939-1945.

Spehar, R.L., R.L. Anderson and J.T. Fiandt. 1978. Toxicity and bioaccumulations of cadmium and lead in aquatic invertebrates. Environ. Pollut. 15(3): 195-208

Spehar, R.L., J.T. Fiandt, R.L. Anderson and D.L. Defoe. 1980. Comparative toxicity of arsenic compounds and their accumulation in invertebrates and fish. Arch. Environ. Contam. Toxicol. 9(1):53-63

Spehar R.L. and J.T. Fiandt. 1986. Acute and chronic effects of water quality criteriabased metal mixtures on three aquatic species. Environ. Toxicol. Chem. 5(10):917-931

Spencer, D.F., and R.W. Greene. 1981. Effects of nickel on seven species of freshwater algae. Environ. Pollut. Ser. A Ecol. Biol. 25(4):241-247

Sprague, J.B. 1964. Avoidance of copper-zinc solutions by young salmon in the laboratory. J. Water Pollut. Control Fed. 36:990.

Stauber, J.L. and T.M. Florence. 1987. Mechanism of toxicity of ionic copper and copper complexes to algae. Mar. Biol. 94(4):511-519

Steel, R.G.D. and J.H. Torrie. 1960. Principles and Procedures of Statistics with Special Reference to the Biological Sciences. McGraw-Hill Book Company, New York, N.Y. p.481

Steemann-Nielsen, E. and S. Wium-Anderson. 1970. Copper ions as poisons in sea and in freshwater. Mar. Biol. 6: 93.

Stokes, P.M. 1981. Multiple metal tolerance in copper-tolerant green algae. J. Plant Nutry. 3(1-4):667-678

Sugatt, R.H. 1980. Effects of sodium dichromate exposure on the immune responses of juvenile coho salmon, *Oncorhynchus kisutch* against *Vibro anguillarum*. Arch. Environ. Contam. Toxicol. 9:207

Surber, E.W. and O.L. Meehan. 1931. Lethal concentrations of arsenic for certain aquatic organisms. Trans. Am. Fish. Soc. 61:225-239

Suzuki, K. 1959. The toxic influence of heavy metals salts upon mosquito larvae. Hokkaido Univ. J. Fac. Sci. Ser. 6(14):196-209

SWRCB. 1990. Sacramento River Toxic Chemical Risk Assessment Project. Final Project Report. Section III p22-29. State Water Resources Control Board Division of Water Quality and Regional Water Quality Control Board, Central Valley Region. Report 90-11WQ. October.

Tewari, H., T.S. Gill and J. Pant. 1987. Impact of chronic lead poisoning on the hematological and biochemical profiles of a fish, *Barbus conchonius* (Ham). Bull. Environ. Contam. Toxicol. 38(5):748-752

Thorp, J.H., et al. 1979. Effects of chronic cadmium exposure on crayfish survival, growth, and tolerance to elevated temperatures. Arch. Environ. Contam. Toxicol. 8: 449.

USEPA. 1984A. Ambient water quality criteria for copper. Office of Water Regulations and Standards. EPA-440/5-84-032

USEPA. 1984B. Ambient water quality criteria for cadmium. Office of Water Regulations and Standards. EPA-440/5-84-032

USEPA. 1985A. Ambient water quality criteria for arsenic. Office of Water Regulations and Standards. EPA-440/5-84-033

USEPA. 1985B. Ambient water quality criteria for chromium. Office of Water Regulations and Standards. EPA-440/5-84-029

USEPA. 1985C. Ambient water quality criteria for lead. Office of Water Regulations and Standards. EPA-440/5-84-027

USEPA. 1986. Ambient water quality criteria for nickel. Office of Water Regulations and Standards. EPA-440/5-86-004

USEPA. 1987. Ambient water quality criteria for zinc. Office of Water Regulations and Standards. EPA-440/5-87-003

USEPA. 1991. Methods for aquatic toxicity identification evaluations: Phase I Toxicity Characterization Procedures, Second Edition, Environmental Research Laboratory, Special Publications EPA/600/6-91/003.

USEPA. 1992. Methods for aquatic toxicity identification evaluations: Phase II Toxicity identification procedures for and development, Special Publications EPA/600/R-92/080.

USEPA. 1994. Short-term methods for estimating the chronic toxicity of effluents and receiving waters to freshwater organisms, Third Edition, Special Publication EPA-600/4-91/002.

Van Leeuwen, C.J., W.J. Luttmer, and P.S. Griffioen. 1985. The use of cohorts and populations in chronic toxicity studies with *Dapnia magna*: a cadmium example. Ecotoxicol. Environ Safety 9: 26-39.

Vardia, H.K., P.S. Rao, and V.S. Durve. 1988. Effects of copper, cadmium, and zinc on fish-food organisms, *Daphnia lumholtzi* and *Cypris subglobosa*. Proc. Indian Acad. Sci. (Acad. Sci.) 97(2): 175-180.

Vareille-Morel, C. and C. Chaisemartin, 1982. Natural tolerance and acclimation of different populations of *Austropotamobius pallipes* to heavy metals. Acta. Oecol. Appl. 3(1):105-122

Vasseur, P., P. Pandard, and D. Burnel. 1988. Influence of some experimental factors on metal toxicity to *Selenastrum capricornutum*. Toxic. Assess. 3(3): 331-444.

Vocke, R.W., K.L. Sears, J.J. O'Toole and R.B. Wildman. 1980. Growth responses of selected freshwater algae to trace elements and scrubber ash slurry generated by coal-fired power plants. Water Res. 14(2):141-150

Wangberg, S.A., U. Heyman and H. Blanck. 1991. Long-term and short-term arsenate toxicity to freshwater phytoplankton and periphyton in limnocorrals. Can. J. Fish. Aquat. Sci. 48(2):173-182

Weir, P.A. and C.H. Hine. 1970. Effects of various metals on behavior of conditioned goldfish. Arch. Environ. Health 20(1):45-51

Welsh, P., J. Lipton, R. Hunson, T. Podrabsky, and D. Cacela. 1998. Data report: Acute copper toxicity to salmonids in surface waters in the vicinity of the iron mountain mine. California- Volume I. Hagler Bailly Service Inc.

Westerman, A.G. and W.J. Birge. 1978. Accelerated rate of albinism in channel catfish exposed to metals. Prog. Fish-Cult. 40:143

Wettern, M., D.W. Lorch and A. Weber. 1976. The effect of lead and manganese on the green alga *Pediastrum tetras* in axenic culture I. Accumulation rates and influence on growth. Arch Hydrobiol. 77(3):267-276

Williams, P.L. and D.B. Dusenbery. 1990. Aquatic toxicity testing using the nematode, *Caenorhabditis elegans*. Environ. Toxicol. Chem. 9(10):1285-1290

Willis, M. 1988. Experimental studies of the effects of zinc on *Ancylus fluviatilis* (Muller) (Mollusca: Gastropoda) from the Afon Crafnant, N. Wales. Arch. Hydrobiol. 112(2): 299-316

Wilson, W.B., and L.R. Freeburg. 1980. Toxicity of metals to marine phytoplankton cultures. EPA-600/3-80-025, Ecol. Res. Ser., Ser., U.S. EPA, Environ. Res. lab., Narragansett, RI:110p.

Wilson, D.B., B. Finlayson, and N. Morgan. 1981. Copper, zinc and cadmium concentrations of resident trout related to acid-mine wastes. California Department of Fish and Game 67:176-186

Winner, R.W. 1988. Evaluation of the relative sensitivities of 7-day *Daphnia magna* and *Ceriodaphnia dubia* toxicity tests for cadmium and sodium pentachlorophenate. Environ. Toxicol. Chem. 7: 153-159

Wium-Andersen, S. 1974. The effect of chromium on the photosynthesis and growth of diatoms and green algae. Physiol. Plant. 32: 308-310

Wong, C.K. 1993. Effects of chromium, copper, nickel and zinc on longevity and reproduction of the cladoceran *Moina macrocopa*. Bull. Environ. Contam. Toxicol. 50:633-639

Wurtz, C.B. 1962. Zinc effects on fresh water mollusks. Nautilus. 76: 53-61.

Yongue, W.H. Jr., B.L. Berrent and J. Cairns Jr. 1979. Survival of *Euglena gracilis* exposed to sublethal temperature and hexavalent chromium. J. Protozool. 26(1): 122-125

Young, R.G. and D.J. Lisk. 1972. Effect of copper and silver ions on algae. Jour. Water Pollut. Control Fed. 44: 1643.

Zarafonetis, J.H. and R.E. Hampton. 1974. Some effects of small concentrations of chromium on growth and photosynthesis in algae. Michigan Acad. 6:417

Site Name	Date Sampled
5 Mile Sl	10/5/94
Antioch	7/19/93
Antioch	7/19/93
Antioch	4/27/94
Antioch	11/4/94
Antioch	11/4/94
Duck Slough	5/10/94
Duck Slough	5/10/94
Duck Slough	7/12/94
Duck Slough	7/12/94
Duck Slough	8/9/94
Duck Slough	8/9/94
Duck Slough	9/2/94
Duck Slough	9/2/94
Duck Slough	9/2/94
Duck Slough	1/9/95
French Camp Slough	3/23/94
French Camp Slough	3/23/94
French Camp Slough	9/2/94
French Camp Slough	9/2/94
Grizzly Bay	2/5/95
Grizzly Bay	2/5/95
Martinez	2/5/95
Martinez	2/5/95
Martinez	2/5/95
Middle R. @ Bullfrog	7/7/93
Middle R. @ Bullfrog	7/7/93
Middle R. @ Bullfrog	8/17/93
Middle R. @ Bullfrog	8/17/93
Middle R. @ Bullfrog	10/29/93
Middle R. @ Bullfrog	10/29/93
Middle R. @ Bullfrog	1/11/94
Middle R. @ Bullfrog	1/11/94
Middle R. @ Bullfrog	1/11/94
Middle R. @ Bullfrog	4/27/94
Middle R. @ Bullfrog	4/27/94
Mokelumne River	8/3/93
Mokelumne River	8/3/93
Mokelumne River	9/14/93
Mokelumne River	9/14/93
Mokelumne River	9/14/93
Mokelumne River	10/14/93
Mokelumne River	10/14/93
Mokelumne River	4/12/94
Mokelumne River	4/12/94
Mokelumne River	5/10/94
Mokelumne River	5/10/94

Site Name	Date Sampled
Mokelumne River	7/21/94
Mokelumne River	10/19/94
Mokelumne River	12/13/94
Mokelumne River	12/13/94
Mokelumne River	12/13/94
Mokelumne River	3/22/95
Mokelumne River	3/22/95
Old River @ Tracy Blvd.	5/25/94
Old River @ Tracy Blvd.	5/25/94
Old River @ Tracy Blvd.	6/3/94
Old River @ Tracy Blvd.	6/3/94
Paradise Cut	4/30/94
Paradise Cut	5/10/94
Paradise Cut	5/10/94
Paradise Cut	5/25/94
Paradise Cut	5/25/94
Paradise Cut	6/3/94
Paradise Cut	6/3/94
Paradise Cut	7/12/94
Paradise Cut	7/12/94
Prospect Slough	7/12/94
Prospect Slough	7/12/94
Prospect Slough	8/9/94
Prospect Slough	8/9/94
Prospect Slough	9/2/94
Prospect Slough	9/2/94
Prospect Slough	9/2/94
Prospect Slough	1/10/95
Prospect Slough	1/10/95
Prospect Slough	1/11/95
Prospect Slough	1/12/95
Prospect Slough	1/13/95
Prospect Slough	1/14/95
Prospect Slough	1/15/95
Prospect Slough	1/15/95
Prospect Slough	1/17/95
Prospect Slough	1/18/95
Prospect Slough	1/22/95
Prospect Slough	1/23/95
Prospect Slough	1/25/95
Prospect Slough	1/25/95
Prospect Slough	1/26/95
Prospect Slough	1/26/95
Prospect Slough	1/27/95
Prospect Slough	1/28/95
Prospect Slough	1/28/95
Prospect Slough	1/31/95

Table 1. Sites and Dates of Sampling in the Delta and Lower Sacramento River Basin

Table 1 (	cont).	Sites and Dates	of Sam	pling in	the	Delta and	Lower	Sacramento	River	Basin
-----------	--------	-----------------	--------	----------	-----	-----------	-------	------------	-------	-------

Site Name	Date Sampled
Prospect Slough	2/3/95
Prospect Slough	2/6/95
Prospect Slough	2/10/95
Prospect Slough	2/14/95
Prospect Slough	2/17/95
Prospect Slough	2/28/95
Prospect Slough	3/21/95
S I River @ Pt Antioch	10/29/93
S I River @ Pt Antioch	10/29/93
S I River @ Pt Antioch	10/29/93
S I River @ Pt Antioch	11/29/93
S I River @ Pt Antioch	1/10/94
S I River @ Pt Antioch	1/10/94
Sac River @ G. Landing	1/11/93
Sac River @ G. Landing	1/13/03
Sac River @ G. Landing	1/1/03
Sac River @ G. Landing	11/10/02
Sac River @ G. Landing	11/10/93
Sac River @ G. Landing	11/11/93
Sac River @ G. Landing	1/12/93
Sac River @ G. Landing	1/10/94
Sac River @ G. Landing	1/13/94
Sac River @ G. Landing	1/18/94
Sac River @ G. Landing	1/19/94
Sac River @ G. Landing	1/23/94
Sac River @ G. Landing	1/24/94
Sac River @ G. Landing	1/25/94
Sac River @ G. Landing	1/26/94
Sac River @ G. Landing	1/27/94
Sac River @ G. Landing	1/28/94
Sac River @ G. Landing	1/29/94
Sac River @ G. Landing	1/30/94
Sac River @ G. Landing	1/31/94
Sac River @ G. Landing	2/1/94
Sac River @ G. Landing	2/2/94
Sac River @ G. Landing	2/5/94
Sac River @ G. Landing	2/7/94
Sac River @ G. Landing	2/8/94
Sac River @ G. Landing	2/9/94
Sac River @ G. Landing	2/10/94
Sac River @ G. Landing	2/11/94
Sac River @ G. Landing	2/12/94
Sac River @ G. Landing	2/16/94
Sac River @ G. Landing	2/17/94
Sac River @ G Landing	2/18/94
Sac River @ G Landing	2/19/94
Sac River @ G. Landing	2/20/94
Sac River @ G. Landing	2/20/94
Sac River @ C. Landing	2/21/24
Sac River () C. Landing	2/22/24
Sac River @ C. Landing	2123194
Sac River @ G. Landing	2124124
Sac River @ C. Landing	2122194
Sac River (0) G. Landing	2/2//94

Site Name	Date Sampled
Sac River @ G. Landing	2/28/94
Sac River @ G. Landing	3/1/94
Sac River @ G Landing	3/4/94
Sac River @ G. Landing	3/9/94
Sac River @ G. Landing	3/10/94
Sac River @ G. Landing	3/15/94
Sac River @ G. Landing	3/16/94
Sac River @ G. Landing	5/10/94
Sac River @ G. Landing	10/5/94
Sac River @ G. Landing	1/6/95
Sac River @ G. Landing	1/7/95
Sac River @ G. Landing	1/8/95
Sac River @ G. Landing	1/10/95
Sac River @ G. Landing	1/11/95
Sac River @ G. Landing	1/12/95
Sac River @ G. Landing	1/13/95
Sac River @ G. Landing	1/14/95
Sac River @ G. Landing	1/15/95
Sac River @ G. Landing	1/17/95
Sac River @ G. Landing	1/18/95
Sac River @ G. Landing	1/20/95
Sac River @ G. Landing	1/22/95
Sac River @ G. Landing	1/23/95
Sac River @ G. Landing	1/24/95
Sac River @ G. Landing	1/25/95
Sac River @ G. Landing	1/26/95
Sac River @ G. Landing	1/27/95
Sac River @ G. Landing	1/28/95
Sac River @ G. Landing	1/29/95
Sac River @ G. Landing	1/30/95
Sac River @ G. Landing	1/31/95
Sac River @ G. Landing	2/1/95
Sac River @ G. Landing	2/2/95
Sac River @ G. Landing	2/3/95
Sac River @ G. Landing	2/6/95
Sac River @ G. Landing	2/10/95
Sac River @ G. Landing	2/14/95
Sac River @ G. Landing	2/1//95
Sac River @ C. Landing	2/21/93
Sac River @ C. Landing	2/23/93
Sac River @ C. Landing	2/24/93
Sac River @ G. Landing	2/20/95
Sac River @ G. Landing	3/5/95
Sac River @ G. Landing	3/7/05
Sac River @ G. Landing	3/11/05
Sac River @ G. Landing	3/22/05
Sac R @ Hood	7/19/93
Sac. R. @ Hood	7/19/93
Sac. R. @ Hood	8/3/93
Sac. R. @ Hood	8/3/93
Sac R @ Hood	8/3/93

ý

Site Name	Date Sampled
Sac, R. @ Hood	9/14/93
Sac. R. @ Hood	9/14/93
Sac. R. @ Hood	10/14/93
Sac. R. @ Hood	10/14/93
Sac R @ Hood	10/14/93
Sac R @ Hood	12/13/93
Sac R @ Hood	12/13/93
Sac. R. @ Hood	12/13/93
Sac. R. @ Hood	4/12/94
Sac R @ Hood	4/12/94
Sac R @ Hood	4/12/94
Sac. R. @ Hood	4/12/94
Sac R @ Hood	5/10/94
Sac R @ Hood	5/10/94
Sac R @ Hood	5/10/94
Sac River @ Rio Vista	7/20/93
Sac River @ Rio Vista	7/20/93
Sac River @ Rio Vista	7/20/93
Sac River @ Rio Vista	8/3/03
Sac River @ Rio Vista	8/3/93
Sac River @ Rio Vista	9/14/93
Sac River @ Rio Vista	9/14/93
Sac River @ Rio Vista	9/14/93
Sac River @ Rio Vista	10/14/93
Sac River @ Rio Vista	10/14/93
Sac River @ Rio Vista	12/13/93
Sac River @ Rio Vista	12/13/93
Sac River @ Rio Vista	4/12/94
Sac River @ Rio Vista	4/12/94
Sac River @ Rio Vista	5/10/94
Skag Slough	1/22/95
Skag Slough	1/23/95
Skag Slough	1/28/95
Skag Slough	2/14/95
S L River @ Stockton	10/29/93
S I River @ Stockton	10/29/93
S.J. River @ Stockton	10/29/93
S.J. River @ Stockton	11/29/93
S. J. River @ Stockton	1/10/94
S I River @ Stockton	1/10/94
S I River @ Stockton	1/10/94
S I River @ Stockton	4/27/94
S I River @ Stockton	4/27/94
Sveamore	3/13/05
Ullatis Creek	3/23/94
Ellatis Creek	3/23/04
Ullatis Creek	12/13/94
Illatis Creek	12/13/04
S L River (a) Vernalis	7/7/02
S I River @ Vernalis	בסודוד
B.J. KIVCI (G. Vernans	

.

.

Site Name	Date Sampled
S.J. River @ Vernalis	8/17/93
S.J. River @ Vernalis	8/17/93
S.J. River @ Vernalis	10/29/93
S.J. River @ Vernalis	10/29/93
S.J. River @ Vernalis	1/11/94
S.J. River @ Vernalis	1/11/94
S.J. River @ Vernalis	1/11/94
S.J. River @ Vernalis	4/27/94
S.J. River @ Vernalis	3/22/95
S.J. River @ Vernalis	
S.J. River @ Vernalis	
Victoria island	1/9/95

•

Table 1 (cont). Sites and Dates of Sampling in the Delta and Lower Sacramento River Basin

			Monitoring Progra	m	
	Ambient Monitoring Program	SRCSD Waste Water Treatment Plant	Iron Mountain M Prog	ВРТСР	
Metal Detection Limits (µg/l)			USBR: @ Spring Cr. Dam, Keswick Dam, and Shasta Dam	CVRWQCB	
As	1	0.05	NS	NS	0.03
Cd	0.03	0.01	5-10	0.1	0.002
Cr	1	0.05 - 0.1	NS	NS	0.05
Cu	0.5	0.05	20-40	1	0.04
Ni	1	0.05 - 0.15	NS	NS	0.02
РЬ	0.1	<b>0.1</b> (1997)	NS	NS	0.01
Zn	4	0.2 - 0.5	20-40	3	0.01
Analytical Lab	ToxScan Laboratory	Frontier Geoscience	USBR Keswick Dam Lab	CH2M Hill'; Quality Analytical Labs, Inc. <sup>#</sup>	Moss Landing Mussel Watch
Method EPA methods		Variable - see reports	Graphite Furnace AA	Graphite Furnace AA	Evapo-concentration & AA Spectrophotometer

Table 2. Analytical information for four programs monitoring metals in the Sacramento River Watershed

			<b>Monitoring Program</b>	m		
	Ambient Monitoring Program	SRCSD Waste Water Treatment Plant	Iron Mountain M Progi	ВРТСР		
Sample Method	pumped cross- sectional composite and 24-hour time- composite	24-hour composite	grab	grab	Acid cleaned CPE tubing and peristaltic pump	
Total or total recoverable	l or total verable Total recoverable Total recoverable		Mine samples = Total Sac. River = Total and dissolved	Total	Total recoverable	
Citation	1	2	3	3	4	

.

Table 2 (cont.). Analytical information for four programs monitoring metals in the Sacramento River Watershed

1 = Larry Walker Associates. 1996. Sacramento Coordinated Water Quality Monitoring Program 1995 Annual Report

2 = Sacramento Regional County Sanitation District, 1996

3 = RWQCB IMM Monitoring Reports, 1985-86 through 1992-93

4 = Goetzl, J. and M. Stephenson. 1993. Metals Implementation Project: Metals Monitoring of Central Valley Reservoir Releases: 1991-1992

NS = not sampled

\*= 11/95 to 6/93

# = 7/93 - present

			(	COPPE	R		ZINC CHROMIUM							UM		CAD	HARDNESS_		
	DATE	D	Т	C*	<b>C</b> #	0	D	Т	<b>C*</b>	C#	0.	D	Т	C*#	D	т	C*	C#	
	7/19/93	2.22	4.65	9.2	7.2	10	2.06	9.98	85	96	100	0.78	4.09	145	0.013	0.03	0.86	1.9	78
	. 10/29/93		2.72	37.0	29.0	10		4.99	340	380	100		1.34	550		0.014	2.90	6.2	626
	10/29/93	2.73	1.72	37.0	29.0	10	3.18	1.68	340	380	100	2.62	0.19	550	0.018	0.017	2.90	6.2	626
	11/29/93		2.69	37.0	29.0	10		2.3	340	380	100		1.86	550		0.02	2.90	6.2	616
	1/10/94	3.82	3.68	25.9	20.4	10	2	10.5	236	267	100	0.12	3.35	392	0.04	0.02	2.10	4.6	262
	4/27/94	2.71	4.72	. 16.4	13.0	10	1.46	7.06	151	170	100	0.81	3.27	254	0.013	0.031	1.42	3.1	154
52	4/27/94	2.75	4.85	16.4	13.0	10	1.23	6.48	151	170	100	0.63	2.82	254	0.016	0.029	1.42	3.1	154
	11/4/94	2.19	3.69			10	2.97	7.23				0.71	2.31		0.014	0.012			no data

Table 3. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the San Joaquin River at Antioch During Water Years 1993 and 1994.

D = Dissolved concentration ( $\mu g/l$ ) following 0.45  $\mu m$  filtration

T = Total recoverable concentration ( $\mu$ g/l)

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C^ = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 1-hour average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = dissolved concentrations.

		NIC	KEL			ARS	ENIC			SIL	VER			LEAD	HARDNESS	
DATE	D	Т	C*	C#	D	Т	C†	0.	D	T	C^	0.	D	Т	C*#	
7/19/93	1.47	5.91	127	42						0.01	2.25	10	0.08	0.85	1.9	78
10/29/93		3.21	510	170										0.03	11	626
10/29/93	2.73	1.61	510	170									0.25		11	626
11/29/93		2.97	510	170						0.014	79	10		0.07	11	616
1/10/94	0.98	3.42	355	117						0.004	18	10	0.04	0.41	7.1	262
4/27/94	1.98	5.15	227	75									0.12	0.66	4.0	154
4/27/94	1.43	4.15	227	75									0.13	0.93	4.0	154
11/4/94	2.12	4.2			0.13	0.41	5	10	0.004	0.012		10	0.09	0.36		no data

.

.

Table 3 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the San Joaquin River at Antioch During Water Years 1993 and 1994.

-

.

,

Table 4. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Duck Slough During Water Years 1994 and 1995.

		(	COPPE	R				ZINC			СН	ROMI	UM	CADMIUM				HARDNESS
DATE	D	Т	C*	C#	0	D	Т	C*	C#	0.	D	Т	C*#	D.	Т	C*	C#	
5/10/94	4.9	12	11.2	8.8	10	7.76	26	103	116	100	5.39	18.7	175	0.012	0.069	1.02	2.2	98
7/12/94	4.41	12.6	8.6	6.8	10	. 7.17	32.3	79	89	100	4.78	19.6	136	0.035	0.081	0.81	1.8	72
8/9/94	4.52	12.5	8.2	6.4	10	6.75	27.5	75	85	100	5	22.4	130	•0.011	0.066	0.78	1.7	68
9/2/94	-	13.5	8.4	6.6	10		29.6	77	87	100		23.1	133		0.071	0.79	1.7	70
9/2/94	3.58	14.9	8.4	6.6	10	4.56	30.7	77	87	100	4.08	21.9	133	0.021	0.064	0.7 <del>9</del>	1.7	70
1/9/95	3.39	-	23.5	18.5	10	2.75	-	215	243	100	2.41	-	357	0.021	-	1.93	4.2	234

54

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

- T = Total recoverable concentration ( $\mu$ g/l)

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = dissolved concentrations

.H

		NIC	KEL			ARSI	ENIC			LEAD	HARDNESS		
DATE	D	Т	C*	C#	D	Т	C†	0.	D	Т	C*#		
5/10/94	8.52	24.1	155	51	1.09	2.06	5	10	1.05	3.3	2.5	98	
7/12/94	6.85	28.8	119	39	1.32	1.58	5	10	0.88	4.28	1.8	72	
8/9/94	8	31.4	113	38	2.05	2.4	5	10	1.38	8.98	1.6	68	
9/2/94		35.8	116	38		2.21	5	10		8.56	1.7	70	
9/2/94	5.16	34.3	116	38	2.17	3.98	5	10	1.08	7.39	1.7	70	
1/9/95	6.35	-	323	107	-	-	5	10	0.37	-	6.3	234	

Table 4 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Duck Slough During Water Years 1994 and 1995.

.

.

.

•

• .

•

-

.

,

Table 5. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from French Camp Slough During Water Year 1994.

	COPPER						ZINC					CHROMIUM			CAD	HARDNESS		
DATE	D	Т	C*	C#	0	D	Т	C*	<b>C#</b>	0'	D	Т	C*#	D	Т	C*	C#	
3/23/94	2.83	2.72	5.6	4.4	10	3.59	9.24	52	59	100	0.81	4	91	0.011	0.044	0.56	1.2	44
9/2/94	2.94	6.17	9.6	7.6	10	2.27	13.3	88	100	100	0.99	3.64	151	0.014	0.038	0.89	1.9	82

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

 $T = Total recoverable concentration (\mu g/l)$ 

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = diss. concentrations.

æ

56

		KEL			ARSI	ENIC			LEAD	HARDNESS		
DATE	D	Т	C*	C#	D	Т	Ċ†	o	D	Т	C*#	· · ·
3/23/94	1.29	3.33	78	26	1.33	1.49	5	10	0.41	2.26	1.0	44
9/2/94	0.99	2.15	133	44	2.4	2.71	5	10	0.37	1.58	2.0	82

Table 5 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from French Camp Slough During Water Year 1994.

ŧ'

۰.

- -

.

,

Table 6. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Hood During Water Years 1993 and 1994.

	COPPER						ZINC						UM			HARDNESS		
DATE	D	Т	C*	C#	<b>o</b> .	D	Т	C*	C#	0.	D	T	C*#	D	Т	C*	C#	
7/19/93	1.42	3.6	6.1	4.8	10	1.12	6.46	56	63	100	0.32	2.85	98	nd	0.041	0.60	1.3	48
8/3/93	1.61	3.77	8.0	6.3	10	1.47	5.91	73	83	100	0.36	3.25	127	0.015	0.039	0.76	1.6	. 66
8/3/93		4.18	8.0	6.3	10		7.41	73	83	100		3.27	127		0.037	0.76	1.6	66
9/14/93	2	3.76	7.8	6.1	10	5.02	16	72	81	100	0.36	2.52	124	0.026	0.038	0.74	1.6	64
10/14/93	1.38	2.71	6.1	4.8	10	1.29	8.55	56	63	100	0.22	1.57	98	0.012	0.036	0.60	1.3	48
10/14/93	1.39		6.1	4.8	10	0.95		56	63	100	0.34		98	0.014	·	0.60	1.3	48
12/13/93		4.38	6.7	5.3	10		7.5	62	70	100		3.99	107		0.08	0.65	1.4	54
12/13/93	2.16	4.35	6.7	5.3	10	0.38	7.6	62	70	100	0.19	3.4	107	0.01	0.07	0.65	1.4	54
4/12/94	2.12	2.89	8.4	6.6	10	2.36	4.62	77	87	100	0.4	1.34	133	0.015	0.027	0.79	1.7	70
4/12/94	2.17	2.94	8.4	6.6	10	1.72	3.81	77	87	100	0.34	1.03	133	0.015	0.033	0.79	1.7	70
5/10/94		2.63	6.7	5.3	10		5.14	62	70	100		1.52	107		0.036	0.65	1.4	54
5/10/94	1.84	2.94	6.7	5.3	10	1.33	3.8	62	70	100	0.55	1.36	107	0.016	0.026	0.65	1.4	54

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

· £5

T = Total recoverable concentration ( $\mu g/l$ )

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C^ = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 1-hour average criteria)

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = diss. concentrations.

10

58
		NIC	KEL			LEAD		<u>.</u>	SILV	'ER		HARDNESS
DATE	D	Т	C*	C#	D	Ť	C*#	Ď	Т	<b>C^</b>	Ó•	
7/19/93	0.7	4.19	84	28	0.06	2.85	1.1	0.003	0.009	0.98	10	48
8/3/93	0.84	4.3	111	37	0.05	0.61	1.6	0.004		1.69	10	66
8/3/93		4.81	111	37		0.53	1.6		0.011	1.69	10	66
9/14/93	0.96	3.76	108	36	0.03	0.3	1.5			1.60	10	64
10/14/93	0.63	2.3	84	28	nd	0.31	İ.1			0.98	10	48
10/14/93	0.67		84	28	0.06		1.1			0.98	10	48
12/13/93		4.52	93	31		0.64	1.3	0.002	0.012	1.20	10	54
12/13/93	0.87	4.81	93	31	0.04	0.63	1.3			1.20	10	54
4/12/94	0.92	2.02	116	38	0.07	0.24	1.7			1.87	10	70
4/12/94	0.75	1.64	116	38	0.075	0.24	1.7			1.87	10	70
5/10/94		2.34	93	31		0.29	1.3			1.20	10	54
5/10/94	1	1.83	93	31	0.09	0.34	1.3			1.20	10	54

,

Table 6 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Hood During Water Years 1993 and 1994.

,

Table 7. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Middle River at Bullfrog Landing During Water Years 1993 and 1994.

	COPPER							ZINC			CH	IROMI	UM		CAD	AIUM		HARDNESS
DATE	D	Т	C*	C#	0	D.	Т	C*	C#	o	D	T	C*#	D	Т	C*	C#	
7/7/93.	1.67	2.54	8.8	6.9	10	1.15	6.77	81	92	100	0.45	0.007	139		0.007	0.83	1.8	74
8/17/93	1.73	28.3	6.1	4.8	10,	1.31	6.66	56	63	100	.0.58	26.8	98		0.456	0.60	1.3	48
10/29/93	1.47	1.59	7.5	6.0	10	0.62	1.34	70	79	100	0.24	0.41	120	0.005	0.01	0.72	1.6	62
1/11/94		2.06	10.2	8.0	10		2.2	94	106	100		0.56	160		0.02	0.94	2.0	88
1/11/94	2.01	0.75	10.2	8.0	10	1.2	1.7	94	106	100	0.39	0.24	160	0.02	0.01	0.94	2.0	88
4/27/94	2.07	2.38	13.6	10.8	10	0.16	1.97 ·	125	142	100	0.28	0.68	212	0.007	0.01	1.21	2.6	124

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

 $T = Total recoverable concentration (\mu g/l)$ 

60

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>^</sup> = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 1-hour average criteria)

		NIC	KEL			LEAD			SIL	VER		HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	Ð	Т	<b>C^</b>	0•	
7/7/93	1.04	2.62	122	40	0.1	0.46	1.8	0.005	0.013	2.06	10.00	74
8/17/93	1.22	38.8	84	28	0.22	39.4	1.1					48
10/29/93	0.71	1.07	105	35		0.13	1.5					62
1/11/94		2.16	141	47		0.11	2.2					88
1/11/94	1.52	0.84	141	47	0.06	0.03	2.2					88
4/27/94	1.41	1.98	189	62	0.06	0.16	3.2					124

Table 7 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Middle River at Bullfrog Landing During Water Years 1993 and 1994.

- -

4

.

.

.

\$

			COPPE	R				ZINC			СН	ROMI	UM	CADMIUM			HARDNESS	
DATE	D	Т	C*	C#	0.	D	Т	C*	C#	0	D	Т	C*#	D	Т	C*	C#	
8/3/93			4.7	3.7	10			44	50	100			· 77			0.48	1.1	36
8/3/93	1.62	1.98	4.7	3.7	10	2.49	6.15	44	50	100	0.09	0.66	77	0.013	0.022	0.48	1.1	36
9/14/93		3.19	4.3	3.4	10		4.84	40	45	100		1.08	70		0.031	0.44	1.0	32
9/14/93	1.6	2.8	4.3	3.4	10	3.16	4.12	40	45	100	0.09	1.51	70	0.011	0.026	0.44	1.0	32
10/14/93	1.37	1.77	3.4	2.6	10	1.24	3.37	31	35	100	0.11	0.54	55	0.01	0.017	0.36	0.8	24
4/12/94	1.29	2.21	4.3	3.4	10	0.75	4.2	40	45	100	0.2	1.49	70	0.005	0.013	0.44	1.0	32
5/10/94		2.42	4.1	3.2	10		4.51	38	43	100	÷	0.94	66	:	0.012	0.42	0.9	30
5/10/94		2.05	4.1	3.2	10		2.91	38	43	100		1.06	66		0.006	0.42	<b>0.9</b>	30
7/21/94	1.25	2.01			10	5.65	5.32			<sub>c</sub> 100	0.16	0.72		0.017	0.024			no data
7/21/94	1.14	1.88			10	5.57	6.34			100	0.11	0.57		0.008	0.022			no data
10/19/94		2.15			10	·•	7.29			100		0.73		-	0.019			no data
12/13/94	1.84	3.97			10	4.1	52.8			100	0.72	3.54		0.01	0.02			no data
12/13/94	1.89				10	2				100	0.77			0.01				no data
3/11/95		4.31	3.1	2.5	10		16.1	29	33	100		2.41	52		0.066	0.34	0.7	22
3/11/95		4.79	3.1	2.5	10		6.27	29	33	100		3.86	52		0.033	0.34	0.7	22
3/22/95		4.26	4.7	3.7	10		18.2	44	50	100		2.1	77		0.095	0.48	1.1	36
3/22/95		4.72	4.7	3.7	10		13.3	44	50	100		1.93	77		0.084	0.48	1.1	36

Table 8. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Mokelumne River During Water Years 1993, 1994, and 1995.

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

 $T = Total recoverable concentration (\mu g/l)$ 

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C^ = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 1-hour average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = diss.concentrations.

		NIC	KEL		·	LEAD		<u> </u>	SIL	VER		ARSENIC				HARDNESS
DATE	Ð	Т	C*	C#	Ð	T	C*#	D	Т	<b>C^</b>	0•	D	Т	C†	0•	
 8/3/93			66	22			0.8			0.60	10					36
8/3/93	0.31	0.75	66	22	0.08	0.3	0.8	nđ	0.003	0.60	10					36
9/14/93		1.23	60	20		0.45	0.7									32
9/14/93	0.39	1.11	60	20	0.1	0.5	0.7									32
10/14/93	0.31	0.92	47	16	0.07	0.26	0.5									24
4/12/94	0.55	1.73	60	20	0.1	0.34	0.7									32
5/10/94		1.48	57	19		0.32	0.7						1.27	5	10	30
5/10/94		1.19	57	19		0.38	0.7						1.22	5	10	30
7/21/94	0.44	0.68			0.08	0.3		0.008	0.008		10	0.6	0.5	5	10	no data
7/21/94	0.47	0.63			0.1	0.25						0.45	0.63	5	10	no data
10/19/94		0.83				0.28										no data
12/13/94	1.34	3.34			0.18	0.67										no data
12/13/94	1.33				0.18											no data
3/11/95		2.61	44	14		4.66	0.5									22
3/11/95		5.72	44	14		3.19	0.5									22
3/22/95		2.47	66	22		0.89	0.8									36
3/22/95		1.72	66	22		1.3	0.8									36

Table 8 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Mokelumne River During Water Years 1993, 1994, and 1995.

Table 9. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Old River at Tracy Blvd. During Water Year 1994.

			COPPE	R				ZINC			СН	ROMI	UM		CAD	4IUM		HARDNESS
DATE	D	т	C*	C#	0	D	Т	C*	C#	0.	D	Т	C*#	: D	Т	C*	C#	
5/25/94	1.44	2.43	16.2	12.8	10	1.99	7.18	149	168	100	0.37	2.33	251	0.014	0.02	1.40	3.0	152
6/3/94	1.74	3.84	23.8	18.8	10	1.99	9.26	218	246	100	0.25	3.2	362	0.008	0.023	1.96	4.2	238

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

 $T = Total recoverable concentration (\mu g/l)$ 

64

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

		NIC	KEL			LEAD			_ ARSI	ENIC		HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	D	Т	C†	<b>O•</b>	
5/25/94	3.01	2.82	224	74	0.12	3.06	4.0	1	0.98	5	10	152
6/3/94	1	3.28	327	108	0.05	1.92	6.4	1.58	0.81	5	10	238

Table 9 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Old River at Tracy Blvd. During Water Year 1994.

٠

.

,

1

- -

e

	COPPER							ZINC		,	СН	ROMI	UM		CADN	<u>AIUM</u>		HARDNESS
DATE	D	Т	C*	C#	0.	D	Т	C*	<b>C</b> #	0.	D	Т	C*#	D	Т	C*	C#	
4/30/94	1.19		37	29	10	0.83		340	380	100	0.21		550	0.008		2.9	6.2	432
5/10/94	2.19	3.42	37	29	10	nd	4.86	335	379	100	0.06	2.13	549	0.008	0.018	2.8	6.2	396
5/25/94	1.01		37	29	10	2.07		337	380	100	0.25		550	0.009		2.9	6.2	398
5/25/94	1.81		37	29	10	1.43		337	380	100	0.08		550	nd		2.9	6.2	398
6/3/94	2.41	4.3	36	28	10	2.54	7.3	327	369	100	0.08	nd	536	0.008	0.019	2.8	6.0	384
7/12/94 7/12/94	0.2	4.88	37 37-	29 29	10 10	3.55	8.95	338 338	380 380	100 100	0.2	4.72	550 550	0.007	0.025	2.9 2.9	6.2 6.2	400 400

Table 10. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Paradise Cut During Water Year 1994.

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

T = Total recoverable concentration ( $\mu g/l$ )

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

	<u>KEL</u>			LEAD			ARS	ENIC		HARDNESS		
DATE	D	Т	C*	<b>C</b> #	D	Т	C*#	D	Т	C†	0•	
4/30/94	2.07		510	170	nd		11	1.24		5	10	432
5/10/94	1.83	3.79	504	167	nd	0.33	11	0.24	0.11	5	10	396
5/25/94	2.12		506	167	0.04		11	1.4		5	10	398
5/25/94	2.29		506	167	nd		11	1.34		5	10	398
6/3/94	2.38	4.75	491	162	0.07	0.64	10	1	1.74	5	10	384
7/12/94	2.16	8.59	508	168	0.05	0.6	11	2.27	3.15	5	10	400
7/12/94			508	168			11					400

J.

.

Table 10 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Paradise Cut During Water Year 1994.

.

				COPP	ER				ZINO	C		C	HROM	IIUM	CADMIUM			HARDNESS	
	DATE	D	Т	C*	C#	_0.	D	Т	C*	C#	0	D	Т	C*#	D	T	C*	C#	
	7/12/94	3.52	8.29	9.8	7.7	10	6.83	16.6	90	102	100	3.06	10.8	155	0.017	0.035	0.91	2.0	84.3
	8/9/94	4.1	7.7	8.6	6.8	10	4.03	12.1	79	89	100	3.83	11	136	0.023	0.03	0.81	1.8	72
	9/2/94		8.16	10.0	7.9	10		13.3	92	104	100		9.58	157		0.036	0.92	2.0	86
	9/2/94	4.22	8.49	10.0	7.9	10	3.97	12.2	92	104	100	3.52	9.84	157	0.021	0.031	0.92	2.0	86
	1/10/95		124	9.6	7.6	10		270	88	100	100		242	151		0.568	0.89	1.9	82
	1/10/95		162	9.6	7.6	10		328	88	100	100		271	151		0.52	0.89	1.9	82
	1/11/95		86.9	10.2	8.0	10		172	94	106	100		168	160		0.229	0.94	2.0	88
	1/12/95		34.4	7.5	6.0	10		66.3	70	79	100		57.6	120		0.181	0.72	1.6	62
	1/13/95		17.9	7.1	5.6	10		42.4	66	74	100		32.7	114		0.163	0.69	1.5	58
•	1/14/95		40.3	9.6	7.6	.10		84	88	100	100		- 58	151		0.224	0.89	1.9	82
	1/15/95		29.8	7.3	5.8	10		128	68	77	100		42.3	117		0.203	0.71	1.5	60
	1/15/95		28.9	7.3	5.8	10		128 -	68	77	100		42.5	117		0.197	0.71	1.5	60
	1/17/95		19	6.1	4.8	10		78.9	56	63	100		27.1	98		0.087	0.60	1.3	48
	1/18/95		24.3		no data	10		103		no data	100		32.9	no data		0.17			no data
	1/22/95		13.3	7.8	6.1			26.3	72	81	100		18.7	124		0.092	0.74	1.6	64

Table 11. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Prospect Slough During Water Years 1994 and 1995.

D = Dissolved concentration ( $\mu g/l$ ) following 0.45  $\mu m$  filtration

 $T = Total recoverable concentration (\mu g/l)$ 

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = diss. concentrations.

ō

		1	COPPE	. <u>R</u>		· · · · · · · · · · · · · · · · · · ·		ZINC			C	HROM	IUM	CADMIUM				HARDNESS
DATE	D	Т	C*	C#	0	D	Т	C*	C#	0'	D	Т	C*#	Ď	Т	C*	C#	
1/23/95		14.9	7.3	5.8	10		39.3	68	77	100		17.4	117		0.104	0.71	1.5	60
1/25/95	3.48	9.06	7.8	6.1	10	5.69	28.3	72	81	100	2.51	9.56	124	0.023	0.075	0.74	1.6	64
1/26/95	4.78	15	6.9	5.5	10	8.17	36.3	64	72	100	4.08	21.6	111	0.064	0.107	0.67	1.5	56
1/27/95		12.3	7.3	5.8	10		31.9	68	77	100		19.2	117		0.096	0.71	1.5	60
1/28/95	4.51	12.5	7.3	5.8	10	7.87	32.8	68	77	100	3.69	17.6	117	0.064	0.111	0.71	1.5	60
1/31/95		9.73	8.2	6.4	10		23.3	75	85	100		11.5	130		0.065	0.78	1.7	68
2/3/95		8.69	8.2	6.4	10		19.9	75	85	100		10	130		0.07	0.78	1.7	68
2/6/95		14.7	5.8	4.6	10		29.2	54	61	100		14.3	94		0.082	0.58	1.3	46
2/10/95		7.34	8.0	6.3	10			73	83	100		7.65	127		0.068	0.76	1.6	66
2/14/95		8.22	9.4	7.4	10			87	98	100		10.5	148		0.084	0.87	1.9	80
2/17/95		5.72	15.9	12.5	10			146	165	100		8.08	245		0.036	1.38	3.0	148
2/28/95		8.59	24.3	19.2	10			223	252	100		14.5	370		0.065	1.99	4.3	244
3/21/95		10	6.9	5.5	10		20.5	64	72	100		13.3	111		0.072	0.67	1.5	56

Table 11 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Prospect Slough During Water Years 1994 and 1995.

.

.

69

-

- -

¢

		NIC	KEL _			LEAD		ARSENIC				HARDNESS
DATE	D	Т	<b>C</b> *	C#	D	Т	C*#	D	T	C†	0•	
7/12/94	5.36	15.3	136	45	0.4	1.24	2.1	1,	1.06	5	10	84.3
8/9/94	7.04	15.7	119	39	0.41	1.24	1.8	1.93	1.67	5	10	72
9/2/94		18.3	138	46	i.	2.24	2.1		2.1	5	10	86
9/2/94	6.12	18.5	138	46	0.73	2.06	2.1	2.04	3.24	5	10	86
1/10/95		601	133	44		28.4	2.0		0.6	5	10	82
1/10/95		587	133	44		41.2	2.0			5	10	82
1/11/95		417	141	47		16	2.2		1.46	5	10	88
1/12/95		103	105	35		7.81	1.5		1.5	5	10	62
1/13/95		38	99	33		3.65	1.4		1.63	5	10	58
1/14/95		79.2	1.33	44		13.5	2.0		1.2	5	10	82
1/15/95		53.7	102	34		6.54	1.4		2.48	5	10	60
1/15/95		62.8	102	34		6.15	1.4		2.27	5	10	60
1/17/95		36.6	84	28		2.95	1.1		3.32	5	10	48
1/18/95		45.1		no data		4.82			4.41	5	10	no data
1/22/95		27.3	108	36		2.49	1.5		1.07	5	10 -	64

c

Table 11 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Prospect Slough During Water Years 1994 and 1995.

		NIC	KEL			LEAD			ARS	ENIC		HARDNESS
DATE	D	Ť	C*	C#	D	T	C*#	D	T	C†	0•	
1/23/95		28.8	102	34		3	1.4		1.18	5	10	60
1/25/95	4.39	16.7	108	36	0.38	1.26	1.5	1.43	1.81	5	10	64
1/26/95	7.28	36.6	96	32	0.57	2.53	1.3	1.51	nd	5	10	56
1/27/95		28.3	102	34		2.07	1.4		1.48	5	10	60
1/28/95	6.75	29.3	102	34	0.57	2.11	1.4	1.45	0.99	5	10	60
1/31/95		14.8	113	38		1.45	1.6			5	10	68
2/3/95		13.5	113	38		1.12	1.6			5	10	68
2/6/95		21.3	81	27		1.95	1.1			5	10	46
2/10/95		11.4	111	37		0.76	1.6			5	10	66
2/14/95		15.8	130	43		4.2	2.0			5	10	. 80
2/17/95		13.8	219	72		0.75	3.8			5	10	148
2/28/95		28.3	334	111		1.93	6.5			5	10	244
3/21/95		19.3	96	32		3.45	1.3			5	10	56

Table 11 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Prospect Slough During Water Years 1994 and 1995.

			COPPE	R				ZINC			CH	IROM	IUM		CAD	MIUM		HARDNESS
DATE	D	Т	• C*	C#	0	D	T	C*	C#	0	D	Т	C*#	D	Т	C*	<b>C</b> #	
7/20/93	1.56	3.51	5.6	4.4	10	1.31	6.96	52	59	100	0.41	2.63	91	0.01	0.04	0.56	1.2	44
7/20/93	1.45		5.6	4.4	10	0.7		52	59	100	0.5		91	0.015	-	0.56	1.2	44
8/3/93	2.4	3.17	7.8	6.1	10	2.64	4.55	72	81	100	1.14	2.06	124	0.024	0.031	0.74	1.6	64
9/14/93	1.97	2.98	7.8	6.1	10	1.4	6.08	72	81	100	0.56	2.11	124	0.017	0.035	0.74	1.6	64
9/14/93	1.86		7.8	6.1	10	0.88		72	81	100	0.59	-	124	0.014		0.74	1.6	64
10/14/93	1.91	3.48	6.9	5.5	10	2.64	12.5	64	72	100	0.3	2.36	111	0.025	0.035	0.67	1.5	56
12/13/93	1.58	2.97	9.0	7.1	10	0.71	4.6	83	94	100	0.72	1.56	142	0.01	0.03	0.84	1.8	76
4/12/94	1.88	2.98	9.0	7.1	10	1.06	4.02	83	94	100	0.37	1.77	142	0.019	0.024	0.84	1.8	76
5/10/94	1.9	2.97	7.5	6.0	10	1.75	5:07	<b>70</b> -	79	100	0.52	2.05	120	0.015	0.028	0.72	1.6	62

Table 12 Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Rio Vista During Water Years 1993 and 1994.

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

 $T = Total recoverable concentration (\mu g/l)$ 

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>^</sup> = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 1-hour average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

O' = Site-specific numeric water quality objective (hardness corrected when applicable) for the CVRWQCB Water Quality Control Plan. Objectives = diss. concentrations.

		NIC	KEL			LEAD	<u>.                                    </u>		ARS	ENIC			SILV	ER	<u> </u>	HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	D	Т	C†	0•	D	Т	0^	0•	
7/20/93	1.35	4.97	78	26	0.1	0.62	1.0					nd	0.009	0.84	10	44
7/20/93	1.02		78	26	0.08		1.0					<0.002		0.84	10	44
8/3/93	1.71	2.89	108	36	0.18	0.32	1.5					0.006	0.007	1.60	10	64
9/14/93	1.22	3.24	108	36	0.03	0.21	1.5						0.006	1.60	10	64
9/14/93	1.1		108	36	0.09		1.5					<0.002	nd	1.60	10	64
10/14/93	0.85	3.62	96	32	0.04	0.27	1.3					nd	0.008	1.27	10	56
12/13/93	0.87	2.88	125	41	0.04	0.36	1.9					0.002	0.01	2.15	10	76
4/12/94	1.21	2.99	125	41	0.08	0.26	1.9									76
5/10/94	1.43	3.45	105	35	0.09	0.29	1.5	1.9	2.2	5	10					62

Table 12 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Rio Vista During Water Years 1993 and 1994.

•

.

4

~

- -

Table 13. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Skag SloughDuring Water Year 1995.

		(	COPPE	R				ZINC	2		Cl	HROM	IUM		CAD	MIUM		HARDNESS
DATE	D	Т	C*	C#	oʻ	D	Т	C*	C#	o	D	Т	C*#	D	Т	C*	C#	
1/22/95		11.9	12.9	10.2	10		26.3	119	134	100		22.7	201		0.068	F.15	2.5	116
1/23/95		14.6	13.6	10.8	10		45.6	125	142	100		24.3	212		0.068	1.21	2.6	124
1/28/95		13	11.7	9.3	10		30.3	108.	122	100		20.1	184		0.12	1.06	2.3	104
2/14/95		3.89	19.8	15.6	10			182	205	100		5.74	304		0.026	1.67	3.6	192
3/10/95		5.22	22.3	17.6	10		15.3	204	230	100		4.82	340		0.057	1.85	4.0	220

7

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

T = Total recoverable concentration ( $\mu g/l$ )

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

		NIC	KEL			LEAD			ARSI	ENIC		HARDNESS
DATE	D	Т	C*	C#	D	T	C*#	D	T	C†	0•	
1/22/95		33.9	178	59		2.52	3.0		2.54	5	10	116
1/23/95		41.9	189	62		3.9	3.2		3.08	5	10	124
1/28/95		37.2	162	54		2.19	2.6		1.48	5	10	104
2/14/95		11.1	273	90		0.5	5.1					192
3/10/95		14.1	306	101		4.66	5.9					220

Table 13 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Skag Slough During Water Year 1995.

-

.

.

· · · · · ·

Table 14 Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the San Joaquin River at Stockton During Water Year 1994.

		·	COPPE	R				ZINC			СН	ROMI	UM		CADN	AIUM		HARDNESS
DATE	D	T	C*	C#	0	D	Т	C*	<b>C</b> #	0	D	Т	C*#	D	T	C*	C#	
10/29/93		2.85	8.8	6.9	10		5.55	81	92	100		0.83	139		0.009	0.83	1.8	74
10/29/93	1.98	2.66	8.8	6.9	10	4.5	4.96	81	92	100	0.15	1.16	139	0.006	0.014	0.83	1.8	74
11/29/93		2.66	19.5	15.4	10		8.2	178	202	100		0.98	299		0.03	1.64	3.6	188
1/10/94		2.96	20.9	16.5	10		10.3	191	216	100		0.38	319		0.02	1.75	3.8	204
1/10/94	2.67	2.76	20.9	16.5	10	10	10.8	191	216	100	0.08	0.54	319		0.02	1.75	3.8	204
4/27/94	2.99	4.25	18.0	14.2	10	6.65	13	165	187	100	0.2	0.6	278	0.01	0.021	1.54	3.3	172

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

T = Total recoverable concentration ( $\mu g/I$ ) following  $T = Total recoverable concentration (<math>\mu g/I$ )

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

	_	NIC	KEL			LEAD		HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	
10/29/93		1.66	122	40		1.18	1.8	74
10/29/93	1.29	1.71	122	40	0.23	1.36	1.8	74
11/29/93		1.94	268	89		0.95	5.0	188
1/10/94		2.52	287	95		0.1	5.4	204
1/10/94	2.07	2.3	287	95		0.74	5.4	204
4/27/94	1.84	2.17	249	82	0.16	0.83	4.5	172

.

Table 14 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the San Joaquin River at Stockton During Water Year 1994.

Table 15. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Ulatis Creek During Water Years 1994 and 1995.

		·	COPPE	R				ZINC			CH	ROMI	<u>UM</u>		CAD	MIUM		HARDNESS
DATE	D	Т	C*	<b>C</b> #	0	D	Т	C*	C#	0	<b>D</b> .	Т	C*#	D	Т	C*	C#	
3/23/94	2.98	4.23	29.4	23.2	10	5.55	9.56	268	303	100	1.71	3.87	442	0.018	0.027	2.34	5.1	304
12/13/94	3.89	21.1			10	18.5	57.3			100	0.65	13.1		0.043	0.126			no data

D = Dissolved concentration ( $\mu$ g/l) following 0.45  $\mu$ m filtration

T = Total recoverable concentration ( $\mu g/l$ )

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

		NIC	KEL			LEAD			ARSI	ENIC		HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	D	Т	C†	0.	
3/23/94	3.65	5.69	403	133	0.07	0.46	8.2	1.62	1.78	5	10	304
12/13/94	3.45	16.2			0.2	5.18		1.39	1.22	5	10	no data

Table 15 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from Ulatis Creek During Water Years 1994 and 1995.

Table 16. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the San Joaquin River at Vernalis During Water Years 1993, 1994, and 1995.

				COPPE	R				ZINC			СН	ROMI	UM		CAD	MIUM		HARDNESS
	DATE	D	Т	C*	C#	<b>o</b> .	Ð	Т	C*	C#	0.	D	Т	C*#	D	Т	C*	C#	
	7/7/93	1.63	6.38	15.7	12.4	10	1.52	16.1	144	163	100	0.63	8.38	243		0.015	1.36	3.0	146
	8/17/93	1.5	4.49	14.8	11.6	10	0.96	11.1	136	153	100	0.64	5.7	229		0.011	1.29	2.8	136
	10/29/93	1.09	2.83	14.0	11.1	10	0.47	9.48	129	146	100 -	0.2	2.62	218	0.008	0.02	1.24	2.7	128
	1/11/94	2.47		16.6	13.1	10	0.39		152	172	100	0.17		256			1.43	3.1	156
	1/11/94	1.93	1.51	16.6	13.1	10	0.3	3.5	152	172	100	0.74	1.19	256	0.001	0.01	1.43	3.1	156
00	4/27/94			9.8	7.7	10		0.08	90	102	100			154			0.91	2.0	84
0	4/27/94			9.8	7.7	10		0.24	90	102	100			154			0.91	2.0	84
	4/27/94	1.17	3.58	9.8	7.7	10	0.48	9.24	90	102	100	0.4	4.4	154	0.002	0.014	0.91	2.0	84
	4/27/94	0.68		9.8	7.7	10	0.54		<b>90</b> ·	102	100	0.34		154	· •		0.91	2.0	84
	3/11/95		34.1	12.7	10.0	10		107	117	132	100		69.1	198		0.169	1.14	2.5	114
	3/22/95		2.89	9.8	7.7	10		5.87	90	102	100		2.11	154		0.024	0.91	2.0	84

D = Dissolved concentration ( $\mu g/l$ ) following 0.45  $\mu m$  filtration

 $T = Total recoverable concentration (\mu g/l)$ 

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

Table 16 (cont.).	. Summary of Metal Concentration E	Data and Related Water Qu	ality Objectives for Sampl	es Collected from the San Joa	aquin
River at Vernalis	s During Water Years 1993, 1994, an	d 1995.			

• •

			NIC	KEL			LEAD		HARDNESS
	DATE	D	Т	C*	C#	D	Т	C*#	
	7/7/93	2.23	11.2	217	72		1.43	3.8	146
	8/17/93	1.7	8.9	204	67		1.13	3.5	136
	10/29/93	1.13	4.03	194	64	0.04	0.14	3.3	128
	1/11/94	0.95		229	76			4.1	156
	1/11/94	1.93	2	229	76	0.15	0.06	4.1	156
	4/27/94			136	45			2.1	84
2	4/27/94			136	45			2.1	84
	4/27/94	0.97	5.53	136	45	0.07	0.79	2.1	84
	4/27/94	0.88		136	45	0.09		2.1	84
	3/11/95		128	176	58		17.6	2.9	114
	3/22/95		3.97	136	45		5.43	2.1	84

. .

		(	COPPE	R				ZINC			<u> </u>	IROM	IUM		CAD	<u>MIUM</u>		HARDNESS
DATE	D	Т	C*	<b>C</b> #	o	<b>D</b> .	Т	C*	C#	0 <sup>°</sup>	D	Т	C*#	· D	Т	• C*	C#	
1/6/95	2.99	5.54	10.6	8.3	10	3.2	10.2	97	110	100	1.28	3.71	166	0.028	0.063	0.97	2.1	92
1/7/95	3.39	9.02	8.0	6.3	10	3.75	17.9	73	83	100	1.98	7.2	127	0.028	0.118	0.76	1.6	66
1/8/95	4.91	10.6	7.3	5.8	10	5.59	19.7	68	77	100	2.94	11.4	117	0.038	0.108	0.71	1.5	60
1/10/95	4.9	28.4	6.5	5.1	10	5.99	62.9	60	68	100	3	29	104	0.039	0.474	0.64	1.4	52
1/12/95	3.35	17.4	5.4	4.3	10	2.86	33.1	50	57	100	3.2	19.3	87	0.034	0.184	0.54	1.2	42
1/13/95	3.67	14.2	7.1	5.6	10	6.32	32.5	66	74	100	4.78	21	114	0.035	0.166	0.69	1.5	58
1/14/95	3.94	15.2	5.2	4.1	10	11.2	71.8	48	54	100	4.42	21.3	84	0.018	0.167	0.52	1.1	40
1/15/95	3.62	10.7	5.6	4.4	10	7.93	44.8	52	59	100	3.05	12.2	91	0.031	0.114	0.56	1.2	44
1/17/95	3.6	9.39	5.6	4.4.	. 10	9.4	18.4	52	59	100	3.4	11.6	91	0.002	0.087	0.56	1.2	44
1/18/95	3.68	10.3			10	4.68	46.9			100	3.83	13.3	•	0.033	0.09			no data
1/20/95	4.28	9.68	6.1	4.8	10	4.84	19.5	56	63	100	3.43	12.6	98	0.11	0.089	0.60	1.3	48
1/22/95	3.35	9.98	6.7	5.3	10	4.25	23.3	62	70	100	2.5	12	107	0.025	0.095	0.65	1.4	54
1/23/95	3.42	9.43	6.3	5.0	10	4.41	25.4	58	66	100	2.52	8.57	101	0.024	0.087	0.62	1.3	50
1/24/95	3.09	8.27	6.9	5.5	10			64	72	100	2.68	8.44	111	0.027	0.084	0.67	1.5	56
1/25/95	2.88	7.07	6.7	5.3	10	5.06	20.9	62	70	100	4.43	8.27	107	0.025	0.08	0.65	1.4	54
1/26/95	3.16	9.9	6.3	5.0	10	4.86	24.4	58	66	100	2.07	11	101	0.032	0.111	0.62	1.3	50
1/27/95	3.27	8.82	6.1	4.8	10	6.06	22.3	56	63	100	4.46	10.6	98	0.033	0.08	0.60	1.3	48

 Table 17. Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento

 River at Greene's Landing During Water Year 1995.

			(	COPPE	R				ZINC			Cł	IROM	UM		CAD	MIUM		HARDNESS
	DATE	D	Т	C*	C#	o	D	Т	C*	C#	0.	D	Т	C*#	D	Т	C*	C#	
	1/28/95	2.77	8.11	6.1	4.8	10	5.9	21.7	56	63	100	2.07	9.84	98	0.073	0.082	0.60	1.3	48
	1/29/95	2.89	7.34	5.6	4.4	10	4.34	17.8	52	59	100	2.13	7.75	91	0.034	0.105	0.56	1.2	44
	1/30/95	2.87	6.79	6.1	4.8	10	2.47	14.4	56	63	100	1.75	7.17	98	0.021	0.054	0.60	1.3	48
	1/31/95	1.89	7.02	6.1	4.8	10	3.98	14.6	56	63	100	1.59	6.77	98	0.02	0.104	0.60	1.3	48
	2/1/95		3.53	6.3	5.0	10		12.2	58	66	100		5.02	101		0.07	0.62	1.3	50
	2/2/95		5.9	6.3	5.0	10		13.3	58	66	100		4.88	101		0.042	0.62	1.3	50
1	2/3/95		6.57	6.1	4.8	10		14.3	56	63	100		6.03	98		0.062	0.60	1.3	48
	2/6/95	2.37	6.45	5.8	4.6	10	3.6	14.5	54	61	100	1.68	5.78	94	0.032	0.051	0.58	1.3	46
	2/10/95	2.49	4.95			10	2.41	10.6			100	1.41	4.47		0.012	0.057			no data
	2/14/95		5.07			10							4.65			0.056			no data
	2/17/95		7.3			10							8.79			0.11			no data
	2/21/95		4.99			10							4.16			0.048			no data
	2/23/95		4.78			10							3.93			0.053			no data
	2/24/95		4.08			10							3.9			0.057			no data
	2/28/95		4.14			10							3.97			0.045			no data
	3/3/95		4.75			10							4.44			0.066			no data
	3/5/95		4.94			10							5.02			0.076			no data

Table 17 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Greene's Landing During Water Year 1995.

4

•

- -

٠

•

~ ~ ~

Table 17 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Greene's Landing During Water Year 1995.

			COPPE	R		 		ZINC			C	HROM	IUM		CAD	MIUM		HARDNESS
DATE	D	Т	C*	C#	0'	D	Т	C*	C#	O'	D	Т	C*#	D	Т	C*	C#	
3/7/95		5.73			10							4.94			0.052			no data

D = Dissolved concentration ( $\mu g/l$ ) following 0.45  $\mu m$  filtration

T = Total recoverable concentration ( $\mu g/l$ )

2

C\* = USEPA National Ambient Water Quality Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C# = USEPA Proposed California Toxics Rule Criteria to Protect Freshwater Aquatic Life (expressed as dissolved metal 4-day average criteria)

C<sup>†</sup> = California Proposition 65 Regulatory Level as Drinking Water Level

			NIC	KEL			LEAD			ARSI	ENIC		HARDNESS
	DATE	D	Т	C*	C#	D	Т	C*#	D	Т	C†	0.	
	1/6/95	2.19	6.02	146	48	0.45	1.2	2.3	1.41	1.52	5	10	92
	1/7/95	2.97	10.5	111	37	0.78	3.48	1.6		1.2	5	10	66
	1/8/95	4.51	16	102	34	0.77	3.91	1.4	0.45	0.3	5	10 ·	60
	1/10/95	4.31	3.16	90	30	0.81	11.2	1.2	1.37		5	10	52
	1/12/95	8.5	27.1	75	25	0.53	3.69	1.0	1.19	1.32	5	10	42
	1/13/95	4.78	23.6	99	33	0.65	4.02	1.4	1.14	1.09	5	10	58
ı	1/14/95	6.02	26.9	72	24	0.8	2.66	0.9	0.84	2.45	5	10	40
1	1/15/95	19.1	13.8	78	26	0.48	2.55	1.0	0.91	0.9	5	10	44
	1/17/95	26	24.8	78	26	0.49	1.57	1.0	1.12	0.72	5	10	44
	1/18/95	6.21	23.7			0.52	7.42		1.06	0.61	5	10	no data
	1/20/95	6.33	18	84	28	0.54	2.05	1.1	1.07	1.2	5	10	48
	1/22/95	3.75	16.2	93	31	0.4	1.75	1.3	1.36	1.4	5	10	54
	1/23/95	4.45	13.1	87	29	0.43	3.24	1.2	1.09	1.22	5	10	50
	1/24/95	3.46	11.8	96	32	0.36	1.55	1.3	1.25	1.07	5	10	56
	1/25/95	4.07	12	93	31	0.4	2.11	1.3	1.14	1.52	5	10	54
	1/26/95	4.34	17.4	87	29	0.35	1.83	1.2	1.25	1.59	5	10	50
	1/27/95	4.06	16.2	84	28	0.46	2.28	1.1	1.18	1.08	5	10	48

-

Table 17 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Greene's Landing During Water Year 1995.

\_

4

\*

	,												
			NIC	KEL			LEAD			ARS	ENIC	·	HARDNESS
	DATE	D	Т	. <b>C*</b>	C#	D	Т	C*#	D	Т	<b>C</b> †	0'	
	1/28/95	4.34	15.7	84	28	0.41	2.06	1.1	1	1.24	5	10	48
	1/29/95	3.95	10.8	78	26	0.34	1.63	1.0	1.2	2 1.13	5	10	44
	1/30/95	3.11	11.3	84	28	0.24	1.04	1.1		1.18	5	10	48
	1/31/95	2.99	10.6	. 84	28	0.37	1.04	1.1		1.54	5	10	48
	2/1/95		6.61	87	29	۰ <sup>د</sup> .	1.08	1.2		•			50
	2/2/95		5.92	87	29		0.86	1.2					50
2	2/3/95		8.45	84	28		1.33	1.1					48
	2/6/95	2.44	8.63	81	27	0.25	1.11	1.1					46
	2/10/95	2.15	<b>7.1</b>	• • •	 . ·	0.18	0.63					میں بی	no data
	2/14/95		6.71				0.65					·	no data
	2/17/95		12.3				1.08						no data
	2/21/95		7.04				4.48						no data
	2/23/95		6.31			`	1.56						no data
	2/24/95		4.59				6.94						no data
	2/28/95		5.85				1.16						no data
	3/3/95		5.79				2.86						no data
	3/5/95		6.56				0.96						no data
									,				

Table 17 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Greene's Landing During Water Year 1995.

đ

Table 17 (cont.). Summary of Metal Concentration Data and Related Water Quality Objectives for Samples Collected from the Sacramento River at Greene's Landing During Water Year 1995.

.

4

		NIC	KEL			LEAD			ARS	ENIC		HARDNESS
DATE	D	Т	C*	C#	D	Т	C*#	D	Т	C†	<b>o</b> .	
3/7/95		6.18				I						no data

.

~ ~

Table 18. Number of Dissolved Metal Analyses and Events When Water Quality Objectives or Criteria Were Exceeded for Stations Monitored in the Sacramento/San Joaquin River Delta during Water Years 1993-1995.

STATION	NUMBER OF ANALYSES FOR DISSOLVED METALS	NUMBER OF EVENTS WHEN WATER QUALITY OBJECTIVES/CRITERIA WERE EXCEEDED
Sacramento River @ Antioch	31	0
Duck Slough	34	0
French Camp Slough	14	0
Sacramento River @ Hood	57	0
Middle River @ Bullfrog Landing	28	0
Mokelumne River	25	0 '
Old River @ Tracy Blvd.	14	0
Paradise Cut	42	0
Prospect Slough	42	0
Sacramento River @ Rio Vista	61	0
Skag Slough	0	N/A
San Joaquin River @ Stockton	16	0
Ulatis Creek	7	Ō
San Joaquin River @ Vernalis	35	0
Greene's Landing	143	0
ALL STATIONS COMBINED	549	0

. .

í.

	Ceriodaphnia		Selenas	trum	Pimephales		
Waterway Category	# Events Exhibiting Toxicity (sample size)	Toxicity Related to (number of events):	# Events Exhibiting Toxicity• (sample size)	Toxicity Related to (number of events):	# Events Exhibiting Toxicity (sample size)	Toxicity Related to:	
Main River Inputs into the Delta	2 (29)	diazinon (2) and unknown (1)	0 (26)	N/A	5 (25)	*	
Island Drains	1 (49)	no TIE	0 (45)	N/A	2 (41)	*	
Back-sloughs and Small Upland Drainages	10 (73)	chlorpyrifos (2)†, carbofuran (2)†, and unknown (9)	1 (65)	non-polar organic(1)	7 (62)	*	
Urban Runoff Receiving Water	0 (10)	N/A	0 (9)	N/A	0 (8)	N/A	
Points Along the Pathways of Water Movement Across the Delta	3 (76)	no TIE	0 (68)	N/A	3 (63)	*	
Total Frequency	16 (237)		1 (213)		17 (199)		

~

.

Table 19. Summary of 1993-1994 Toxicity Monitoring Results for the Sacramento/San Joaquin River Delta

= "toxic" defined as significantly reduced cell counts relative to a laboratory control
 = linked to toxicity in fixed-date samples and follow-up samples
 \* = no TIEs conducted due to the chronic nature of the observed toxicity

.

- - -

	Ceri	odaphnia	Selenas	strum	Pimepha	les
Waterway Category	# Events Exhibiting Toxicity (sample size)	Toxicity Related to (number of events):	# Events With Reduced Cell Count• (sample size):	Reduced Cell Count Related to (number of events):	# Events Exhibiting Toxicity (sample size)	Toxicity Related to:
Main River Inputs into the Delta	2 (28)	unknown	6 (20)	unknown	(0) 14	N/A
Island Drains	1 (32)	carbaryl (1)	3 (8)	non-polar organic (1) and unknown (2)	(0) 1	N/A
Back-sloughs and Small Upland Drainages	17 (104)	chlorpyrifos (14)†, diazinon (3), metabolically activated pesticides (2), and unknown (8)	20 (72)	non-polar organic (2) and unknown	(0) 2	N/A
Urban Runoff Receiving Water	4 (7)	diazinon (5)† and chlorpyrifos (4)	1 (5)	no TIE(^)	N/A	N/A
Points Along the Pathways of Water Movement Across the Delta	0(1)	N/A	4 (11)	unknown	N/A	N/A
Total Frequency	24 (172)		29 (116)		(0) 17	

Table 20. Summary of 1994-1995 Toxicity Monitoring Results for the Sacramento/San Joaquin River Delta

(^) = Storm water studies indicate toxicity to algae at Mosher Slough is partially caused by diuron and unknown chemicals

• : cell counts reduced relative to other ambient station sampled on same day

t = linked to toxicity in fixed-date samples and follow-up samples

00

....

 Table 21. Summary of Dissolved Metal Analyses from Samples Collected from 1993 through 1995 and Relationship to Documented

 Effects in the Literature

.

				Documented Effec Concentra	ts in the Literature tions Measured in t	# at Highest Metal his Study
Metal	Average Conc. (ppb)	Range (ppb)	Location of Highest Concentration	Fish	Invertebrates	Algae
Copper	2.64	0.2-9.48	Greene's Landing	Yes	Yes	Yes
Zinc	4.39	0.16-70.2	5-mile	Yes	Yes	Yes
Chromium	1.34	0.06-5.39	Duck Slough	Yes	Yes	Yes
Lead	0.31	0.01-3.87	5-mile	Yes	Yes	No
Cadmium	0.03	0.001-0.55	Greene's Landing	Yes	Yes	No
Nickel	2.72	0.13-26	Greene's Landing	No	Yes	Yes
Arsenic	1.28	0.13-3.03	5-mile	No	Yes	Yes

# = See Tables 22-42 for description of effect, species exposed, and literature reference.

lĜ

Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (µg/L) *	Reference	Where cited
Chlorella rubescens, green algae	lead	IC50	changes in abundance, growth, biochemical process	between 5 and 10	C. E. Calderon Llanten & H. Greppin, 1993. Ref. No. 16488	2
Chlorella pyrenodiosa, green algae	lead	4 d	change in cell number	10.35	J. L. Stauber & T. M. Florence, 1987. Ref. No. 12971	2
Aulosira fertilissima, bluc-green algae	lead acetate	7 d	change in biochemical process	20.7	E.F. Shabana et al., 1986. Ref. No. 3385	2
Anabaena sp., blue green algae	lead nitrate	20 d	change in cell number	21	V. M. Laube et al., 1980. Ref. No. 9477	1, 2
Scenedesmus quadricauda, green algae	lead acetate	14 d	change in chlorophyll content	80	M. Pawlaczyk-Szpilowa et al., 1972. Ref. No. 2741	2
Haematococcus capensis, green algae	lead acetate	7 d	change in cell number	100	T. C. Hutchinson, 1973. Ref. No. 8864	2
Hydrodictyon reticulatum, green algae	lead	7 d	change in biomass	100	U. N. Rai & P. Chandra, 1992. Ref. No. 8987	2
Phytoplankton, mixed freshwater species	lead acetate	4 d	change in biomass	100	K. Pietilainen, 1975. Ref. No. 8184	2
Pediastrum tetras, green algae	lead		change in population size	200	M. Wettern et al., 1976. Ref. No. 10082	2
Chlamydomonas reinhardtii, green algae	lead chloride	1 d	change in chlorophyll content	207	U. Irmer, et al., 1986. Ref. No. 12272	22
Selenastrum capricornutum, green algae	lead nursue	1/d+	changes in cell number: physiology	207	S Capelo et al., 1993. Ref. No. 184 4063	2
Anasystis aeruginosa, bluc-green algae	lead acetate		mortality	250	G. Bringmann & R. Kuhn, 1978. Ref. No. 2463	2
Scenedesmus acuminatus, green algae	lead	6 d	EC50 for change in population size	250	P. M. Stokes, 1981. Ref. No. 9501	2
Scenedesmus obtusiusculus, green algae	lead chloride	7 d	35% growth inhibition	500	T. J. Monahan, 1976	1, 2
Micrasterias thomasiana, green algae	lead chloride	2 hr	histological alteration	621	U. Meindl & G. Roderer, 1990. Ref. No. 3151	2

Table 22. Summary of lead concentrations reported to have adverse effects on sensitive freshwater algal and diatom species

1 - Cited in Lead Criteria Document 1984 (USEPA, 1985A); 2 - Cited in USEPA AQUIRE Database

12

\* Concentration is amount of lead in solution (eg., not as lead acetate); shaded row indicates species used in US EPA Three Species toxicity test protocols

EC50 - median effective concentration; IC50 - mean inhibitory concentration (for growth or a physiological process)

Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (mg/L)	Reference	Where cited
Tetrahymena pyriformis, ciliate	lead chloride	4 min 4	change in oxygen uptake	0.75	J. L. Slabbert & W. S. G. Morgan, 1982. Ref. No. 11048	2
Hyalella azteca, amphipod	lead	70 d	mortality	2.6	U. Borgmann et al, 1993. Rcf. No. 9248	2
Asellus aquaticus, aquatic sowbug	lead nitrate	16 d	LT50	10	L. Migliore & M. De Nicola Giudici, 1990. Ref. No. 10515	2
Lymnaea palustris, marsh snail (freshwater)	lead nitrate	133 d	mortality	12	U. Borgmann et al., 1978. Ref. No. 8314	2
Daphnia magna, water flea	lead acctate	1.7 d	change in biochemical processes	16	<b>R. Berglind</b> et al., 1985. Ref. No. 10906	2
Aeshna cyanea, blue-green dragonfly larvae	lead nitrate	42 d	enzyme alterations	20	W. Meyer et al., 1986. Ref. No. 12306	2
Astacus astacus, European crayfish	lead	14 d	changes in enzymes, histological damage	20	W. Meyer et al., 1991. Ref. No. 376	2
Libuella depressa, dragonfly	lead nitrate	42 d	enzyme alterations	20	W. Meyer et al., 1986. Ref. No. 12306	2
Libuella quadrimaculata, common skimmer dragonfly	lead nitrate	42 d	enzyme alterations	20	W. Meyer et al., 1986. Ref. No. 12306	2
Neanthes arenaceodentata, polychaete	lead chloride	183 d	LOEC for reproductive alterations	20	D. J. Reish & T. V. Gerlinger, 1984. Ref. No. 4007	2
Tubifex tubifex, tubificid worm	lead nitrate	4 d	EC50 for immobilization	42	B. S. Khangarot, 1991. Ref. No. 2918	2
Anodonta grandis, freshwater mussel	lead nitrate	28 d	changes in growth, DNA	50	M. C. Black et al., 1996. Ref. No. 16859	2
Anopheles stephensi, mosquito	lead acetate	1 d	genetic alteration	_60	G. P. Sharma et al., 1988. Ref. No. 5315	2
Caenorhabditis elegans, nematode	lead nitrate	4 d	LC50	60	P. L. Williams & D. B. Dusenbery, 1990. Ref. No. 3437	2
Ceriodaphnia dubia, water flea	lead nitrate	2 d	LC50	248	R. L. Spehar & J. T. Fiandt, 1986. Ref. No. 12093	2

Table 23.	Summary of lead concentrations reported to have adverse effects on sensitive freshwater invertebrate specie	es
-----------	---	----

8

2 - Cited in USEPA AQUIRE database

.

.

\* Concentration is amount of lead in solution (eg., not as lead acetate); shaded row indicates species used in US EPA Three Species toxicity test protocols EC50 - median effective concentration; LC50 - median lethal concentration; LOEC - Lowest observable effect concentration; LT50 - median survival time

88

-

Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (µg/L)	Reference	Where cited
Gasterosteus aculeatus,					J. R. E. Jones, 1938. Ref. No.	
three-spine stickleback	lead nitrate	4.75	LT50	0.2	2657	2
Phoxinus phoxinus,					J. R. E. Jones, 1938. Ref. No.	
minnow	lead nitrate	21 d	mortality	0.5	2657	2
Cyprinus carpio,					H. Nakagawa et al., 1995. Ref.	
common carp	lead nitrate	20 d	enzyme alterations	1.1	No. 16750	2
Heteropneustes fossilis,			changes in enzymes,		K. C. Singhal, 1994. Ref. no	
Indian catfish	lead nitrate	60 d	biochemical processes	6	4448	2
Salmo gairdneri,					P. H. Davies et al., 1976. Ref.	
rainbow trout	lead nitrate	18 min.	physical abnormality	7.2	No. 2103	2
Carassius auratus,					J. R. E. Jones, 1938. Ref. No.	
goldfish	lead nitrate	4.75 d	physiological change	8	2657	2
Pimephales promelas,			States and the second		E.K. Biegert & V. Valkovic.	
fathead minnow	lead nitrate	<u>2.94 d</u>	LT50	10	Ref. No. 5302	2
Salvelinus fontinalis,			• · · · ·		E. S. Adams, 1975. Ref. No.	
brook trout	lead	21 d	impaired locomotion	14.3	15675	2
Salmo salar,			change in hatching		M. Grande & S. Andersen,	
Atlantic salmon	lead nitrate	15.8 d	success	17.2	1983. Ref. No. 10982	2
Brachydanio rerio,			no observable effect on		G. Dave & R. Xiu, 1991. Ref.	
zebrafish	lead acetate	16 d	hatching	20	No. 3680	2
Barbus conchonius,			change in biochemical		H. Tewari et al., 1987. Ref. No.	
rosy barb	lead nitrate	30 d	process	47.4	12599	2
Salvelinus namaycush ,					S. Sauter, et al., 1976. Ref. No.	
lake trout	lead nitrate	115 d	mortality	48	8439	2
Lepomis macrochirus,					S. Sauter, et al., 1976. Ref. No.	
bluegill	lead nitrate	62 d	mortality	70	8439	2
Tilapia aurea,			changes in biochemical,			
tilapia	lead chloride	1 d	blood parameters	100	P. Allen, 1993. Ref. No. 16833	2
Ictalurus punctatus,			,		S. Sauter, et al., 1976. Ref. No.	
channel catfish	lead nitrate	68 d	mortality	75	8439	2

Table 24. Summary of lead concentrations reported to have adverse effects on sensitive freshwater fish species

1 - Cited in Lead Criteria Document 1984 (USEPA, 1985); 2 - Cited in USEPA AQUIRE Database

ø

\* Concentration is amount of lead in solution (eg., not as lead acetate); shaded row indicates species used in US EPA Three Species toxicity test protocols LC50 - median lethal concentration; LT50 - median time for 50% survival

12
Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (mg/L)	Reference	Where cited
Phytoplankton, freshwater species	arsenic acid, sodium salt	109 d	EC50 for change in photosynthetic productivity	1.5	S. A. Wangberg et al., 1991. Ref. No. 9419	2
Scenedesmus obliquus, green algae	arsenic acid, disodium salt	<u>l</u> hr	change in photosynthetic productivity	48	O. Hofslagare et al., 1994. Ref. No. 16250	2
Clorella vulgaris, green algae	arsenic acid, disodium salt	91 d	LOEC for population growth	60	L. E. Den Dooren de Jong, 1965. Ref. No. 2849	2
Chlamydomonas sp ., green algae	arsenic acid, disodium salt	28 d	change in population growth	75	E. R. Christensen & P. A. Zielski, Ref. No. 9773	2
Melosira granulata, diatom	arsenic acid, trisodium salt	20 d	change in population growth	75	D. Planas & F. P. Healey, 1978. Ref. No. 7146	1, 2
Ochromonas vallesiaca, phytoplankton	sodium arsenate		decreased growth	75	D. Planas & F. P. Healey, 1978.	1
Ankistrodesmus falcatus, green algae	arsenic acid, disodium salt	14 d	EC50 for growth	256	Vocke et al., 1980. Ref. No. 5342	1, 2
Spirogvra sp ., green algae	arsenic oxide	<u>1.83 d</u>	physiological change	300	E. W. Surber & O. L. Meehan, 1931. Ref. No. 10297	2
Selenastrum capricornutum, green algae	arsenic acid, misodium salt	4 d	EC50 for population growth	690	J.E.Richter, 1982	1,2
Gloetaenium loitesbergeri, green algae	arsenic acid, sodium salt	1.54 d	physiological change	800	P. V. D. Prasad & Y. B. K. Chowdary, 1981. Ref. No. 15634	2
Nostoc sp ., blue-green algae	arsenic acid, disodium salt	<u>32 d</u>	change in biomass	1000	S. Maeda et al., 1987. Ref. No. 13296	2
Scenesemus quadricauda, green algae	arsenic acid, disodium salt	7 d	change in population growth	2100	G. Bringmann & R. Kuhn, 1980. Ref. No. 5303	2
Chlamydomonas reinhardtii, green algae	arsenic acid, trisodium salt	20 d	change in population growth	2300	D. Planas & F. P. Healey, 1978. Ref. No. 7146	1, 2
Cladophora sp ., green algae	arsenous acid, sodium salt	14 d	100% mortality	2320	B. C. Cowell, 1965.	1
Zygnema sp ., green algae	arsenous acid, sodium salt	14 d	100% mortality	2320	B. C. Cowell, 1965.	1

.

Table 25. Summary of arsenic concentrations reported to have adverse effects on sensitive species of freshwater algae

1 - Cited in Arsenic Criteria Document 1984 (USEPA, 1985B); 2 - Cited in USEPA AQUIRE Database

\* Concentration is amount of arsenic in solution (eg., not as arsenic acid salt); shaded row indicates species used in US EPA Three Species toxicity test protocols

EC50 - median effective concentration; LOEC - lowest observable effect concentration

56

~

.

Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (mg/L)	Reference	Where cited
Daphnia pulex,			EC50 for			_
water flea	arsenic oxide	1d	immobilization	0.5	H. Lilius et al., 1995. Ref. No. 16385	2
Chironomidae,				i	E. W. Surber & O. L. Meehan, 1931.	
midge species	arsenic oxide	2 d	mortality	8	Ref. No. 10297	2
Bosmina longirostris,	arsenic acid,		EC50 for			
water flea	sodium salt	4 d	immobilization	10	A. Novak et al., 1980. Ref. No. 2210	2
Caenis diminuta,					E. W. Surber & O. L. Meehan, 1931.	
mayfly larvae	arsenic oxide	2 d	mortality	16	Ref. No. 10297	_2
Tetrahymena pyriformis,			change in oxygen		J. L. Slabbert & J. P. Maree, 1986. Ref.	
ciliate	arsenic oxide	4.3 min.	uptake	25	No. 12836	2
Paramecium sp .,			change in rate of		E. W. Surber & O. L. Meehan, 1931.	
ciliate	arsenic oxide	2.5 d	growth	80	Ref. No. 10297	2
Gammarus pseudolimnaeus,						
amphipod	arsenic oxide	14 d	mortality	88	R. L. Spehar et al., 1980. Ref. No. 9783	2
Moina macropa,	arsenic acid,		mortality, changes in			
water flea	disodium salt	7 d	growth, reproduction	100	S. Maeda et al., 1990. Ref. No. 3118	2
Belestoma elegans.					M. E. Lanzer-DeSouza & N. M. M.	
water bug	arsenic oxide	1 d	mortality	100	DaSilva, 1988. Ref. No. 13488	2
Hvalella knickerbockeri					F W Surber & O I Meehan 1931	
amphipod	arsenic oxide	2 d · · ·	mortality	800	Ref No 10297	2
Simodophakus sarmulatus	arsenous acid					
Simoaepnaius serruiaius,	arsenous aciu,	acute test	1.050	812	U O Sanders & O B Cone 1966	1
	Soutuin san			012	H. O. Saliders & O. B. Cope, 1900.	
Daphnia magna,	arsenic	14 4	mortality, altered	022	D. J. Contract al. 1090, DC.Ma. 0702	
water fiea	pentoxide	14 0	reproduction	932	R. L. Spenar et al., 1980. Ret. No. 9783	2
Helisoma campanulatum,		1.90		0(1)		
ramshorn snail	arsenic oxide	28 a	monality	961	R. L. Spehar et al., 1980. Ref. No. 9783	2
Lymnaea emarginata,		20.1		0.01		
pond snail	arsenic oxide	28 d	mortality	961	R. L. Spehar et al., 1980. Ref. No. 9783	2
😋 Ceriodaphnia dubia,	arsenic acid,			14651 A Same	R. B-Naddy et al., 1995. Ref. No.	
water flea	-sodium salt	8d	altered reproduction.	146 1020 F	13729	2 C - 2 C

Table 26. Summary of arsenic concentrations reported to have adverse effects on sensitive species of freshwater invertebrates

1 - Cited in Arsenic Criteria Document 1984 (USEPA, 1985B); 2 - Cited in USEPA AQUIRE Database

EC50 - median effective concentration

\* Concentration is amount of arsenic in solution (eg., not as arsenic acid salt); shaded row indicates species used in US EPA Three Species toxicity test protocols

8,

t)

96

e

Species name	Chemical	Duration or test type	Effect/Endpoint	Concentration (µg/L)	Reference	Where cited
Oncorhynchus mykiss, rainbow trout	arsenic acid	1 d	physiological change	25	A. A. Oladimeji, 1984. Ref. No. 10888	2
Morone saxatilis, striped bass larvae	arsenic acid, sodium salt	21 d	mortality	80	R. J. Klauda, 1985. Ref. No. 4233	2
Carassius aratus, goldfish	arsenic acid, monosodium salt	2 d	behavioral change	100	P. A. Weir & C. H. Hine, 1970. Ref. No. 908	2
Lepomis cyanellus, green sunfish	arsenic acid, disodium salt	2 d	LC50	150	E. M. B. Sorensen, 1976. Ref. No. 5549	2
Oncorhynchus kisutch, coho salmon parr	arsenic oxide	183 d	mortality, change in growth & physiology	300	J. W. Nichols et al., 1984. Ref. No. 10236	2
Anabas testudineus, climbing perch	arsenic acid, disodium salt	12 hr	mortality	488	S. Jana & S. S. Sahana, 1989. Ref. No. 2618	2
Clarias batrachus, walking catfish	arsenic acid, disodium salt	13 hr	mortality	488	S. Jana & S. S. Sahana, 1989. Ref. No. 2618	2
Pimephales promelas;	arsenic pentoxide	30/d		530	D. L. DeFoe, 1982. Ref. No. 3687	2
Oncorhynchus mykiss, rainbow trout	arsenic acid, disodium salt	77 d	mortality	1400	S. M. McGreachy & D. G. Dixon, 1990. Ref. No. 273	2
Channa punctatus, snake-head catfish	arsenic acid, disodium salt	28 d	physiological change	1000	K. Ghosh & S. Jana, 1968. Ref. No. 814	2
Colisa fasciata, giant gourami	arsenic oxide	30 d	change in biological process	1500	J. P. Shukla & K. Pandey, 1985. Ref. No. 11412	2
Heteropneustes fossilis, Indian catfish	arsenic oxide	30 d	change in biological process	1500	J. P. Shukla et al., 1985. Ref. No. 11345	2
<i>Jordanella floridae,</i> flagfish ELS	arsenous acid, sodium salt		chronic test	2962	Call et al., 1983; Lima et al., 1984	1
Phoxinus phoxinus, minnow	arsenic acid, disodium salt	65 d	change in biomass of organism	2500	R. Reuther, 1992. Ref. No. 6229	2
Thymallus arcticus, arctic grayling	arsenic acid, disodium salt	4 d	LC50	4760	K. J. Buhl & S. J. Hamilton, 1990. Ref. No. 334	2
Lepomis macrochirus, bluegill larvae	arsenic oxide	1 <u>d</u>	obvious stress on physiology or behavior	5000	V. C. Applegate et al., 1957. Ref No. 638	2

.

Table 27. Summary of arsenic concentrations reported to have adverse effects on sensitive freshwater fishes

.

1 - Cited in Arsenic Criteria Document 1984 (USEPA, 1985B); 2 - Cited in USEPA AQUIRE Database

ELS - early life stage; LC50 - median lethal concentration

\* Concentration is amount of arsenic in solution (eg., not as arsenic acid salt); shaded row indicates species used in US EPA Three Species toxicity test protocols

-

Species Name	Chemical	Chemical Duration (days) or test type Effect/ Endpoint Concentration (ug/L)		Reference	Where Cited	
Microcystis aeroginosa, blue algae	Sodium dichromate (Cr VI)	NR	incipient inhibition	2	Bringmann, 1975. Ref. no. 15144	2
Anabaena orzae, blue green algae	Chromic chloride (Cr III)	7	change in biomass	5.2	Shabana et al., 1986. Ref. no. 3385	2
Aulosira fertilissama, blue green algae	Chromium	7	change in population growth	5.2	Shabana at al., 1986. Ref. no. 3046	2 ·
Chlamydomonis reinhardi, green algae	Potassium dichromate (Cr VI)	NR	reduction in growth	10	Zarafonetis & Hampton, 1974.	]
Selenastrum capricornutum,	Chromiun	0:17	change in photosynthesis	20	Pillard et al., 1987 Ref. no. 12639	2
Thalassiosira guillardi, diatom	Chromium	2	change in population growth	20	Wilson & Freeburg, 1980. Ref. no. 5557	2
Hydrodictyon reticulatum, green algae	Chromium	0.5	change in biomass	100	Rai & Chandra, 1989. Ref. no. 3348	2
Scenedesmus quadricauda, green algae	Chromium oxide (Cr III)	30	change in biochemical processes	100	Angadi & Mathad, 1994. Ref. no. 17433	2
Nitzschia palea, diatom	Chromium	4	change in population growth	150	Wium-Anderson, 1974. Ref. no. 15144	2
Navicula seminuium, diatom	Potassium dichromate (Cr VI)	NR	50% growth reduction	187	Academy of Natural Sciences, 1960	·1
Nitzschia linearis, diatom	Potassium dichromate (Cr VI)	- 5	LC50	208	Patrick et. al., 1968	1
Cyclotella meneghiniana, diatom	Potassium dichromate (Cr Vl)	NR	growth inhibition	500	Cairns and Sheier, 1968	1
Ditylum brightwelli, 	Chromium chloride (Cr III)	5	change in population size	2000	Canterford & Canterford, 1980. Ref. No. 6405	2
Synechocyotis aquatilis. blue-green algae	Chromium	NR	change in population growth	3000	Shavrina & Gapochka, 1984. Ref. No. 11620	2
Chlorella pyrenoidosa, green algae	Chromium	0.17	change in photosynthesis	5000	Wium-Andersen, 1974. Ref. No.15144	2

È.

· P

Table 28. Summary of chromium concentrations reported to have adverse effects on sensitive freshwater algal and diatom species

1 - Cited in Chromium Criteria Document 1984 (USEPA, 1985C); 2 - Cited in USEPA AQUIRE Database; NR = not reported in AQUIRE database

\* Concentration is amount of chromium in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

86

6)

.45

Species Name	Chemical	Duration (days) or test type	Effect/ Endpoint	Concentration (ug/L)	Reference	Where Cited
<i>Euglena gracilis</i> , flagellate euglenoid	Chromium oxide (Cr III)	0.13	mortality	1	Yonge, Berrent, & Cairns, 1979. Ref. no. 15029	2
Daphnia magna, water flea	Chromium (3+) salt	1	LC50	13	Dowden & Bennett, 1965. Ref. no. 915	2
<i>Glenodium halli,</i> dinoflagellate	Chromium	2	change in population growth	20	Wilson & Freeburg, 1980. Ref. no. 5557	2
Tetrahymena pyriformis, ciliate	Chromium nitrate (Cr 111)	0.003	change in oxygen consumption	25	Slabbert & Maree, 1986. Ref. no. 12836	2
Simocephalus vetulas, water flea	Sodium dichromate (Cr VI)	NR	LC50	32.3	Mount, 1982	1
Daphnia pulex, water flea	Sodium dichromate (Cr VI)	NR	LC50	36.3	Mount, 1982	11
Anodonta imbeccillis, mussel	Chromium	4	LC50	39	Keller & Zam, 1991. Ref. no. 108	2
Simocephalus serrulatus, cladoceran	Sodium dichromate (Cr VI)	NR	LC50	40.9	Mount, 1982	1
Ceriodaphnia reticulata, water flea	Chromium	2	LC50	45	Mount & Norberg, 1984. Ref. no. 11181	2
Dugesia dorotocephala, turbellarian	Chromium	0.042	change in behavior	50	Kapu & Schaeffer, 1991. Ref. no. 10582	2
Gymnodium splendons, dinoflagellate	Chromium	2	change in population growth	50	Wilson & Freeburg, 1980. REf. no. 5557	2
Grammararus pseuolimnaeus, amphipod	Potassium dichromate (Cr VI)	NR	LC50	67.1	Call et al., 1983	1
Austropotamobius pallipes, crayfish	Chromium chloride (Cr III)	4	LC50	390	Vareille-Morel & Chaisemartin, 1982. Ref. no. 15732	2
Hyallella azteca, amphipod	Potassium chromate (Cr VI)	NR	LC50	650	Pardue & Wood, 1980. Ref. No. 6703	2
Plumatella emarginata, bryozoan	Chromium	4	LC50	650	Pardue & Wood, 1980. Ref. No. 6703	2

Table 29. Summary of chromium concentrations reported to have adverse effects on sensitive species of freshwater invertebrates

,

.

1 - Cited in Chromium Criteria Document 1984 (USEPA, 1985C); 2 - Cited in USEPA AQUIRE Database; NR = not reported in AQUIRE database \* Concentration is amount of chromium in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

Species Name	Chemical	Duration (days) or test type	Effect/ Endpoint	Concentration (ug/L)	Reference	Where Cited
Carassius aurates giblio, carp	Chromic chloride (Cr III)	9	cytogenetic changes	0.05	Al-Sabtiet al., 1994. Ref. no. 2851	2
Ctenopharyngodon idella, grass carp	Chromium	NR	change in rate of growth	1.5	Mao and Wang, 1990. Ref. no. 9540	2
Heteropneustes fossilis, Indian catfish	Chromium	20	change in rate of growth	10	Pandey and Nisha, 1984. Ref. no 2388	2
Pimephales promeles,	Chromic chloride (Cr III)	÷ 30°	mortality		Gendusa, 1990 Ref. no. 4087	2
Salmo gairdner, rainbow trout	Cromic nitrate (Cr 111)	NR	Chronic value	68.63	Stevens and Chapman, 1984	1
Ictalurus punctatus. Channel catfish	Chromic chloride (Cr III)	30	mortality	154	Gendusa, 1991. Ref. no. 4087	2
Oncorhynchus tshawtscha, Chinook salmon	Chromium potassium salt (Cr IV)	84	mortality	200	Olson, 1958. Ref. no. 14123	2
Salvelinus fontinalis, brook trout	Sodium dichromate (CrVI)	NR	LC50	364.6	Benoit, 1976. Ref. no. 4943	2
Oncorhynchus kisutch, Coho Salmon	Sodium dichromate (CrVI)	14	Immuno-suppression	470	Sugatt, 1980.	1
Carassius auratus, goldfish	Chromium	7	LC50	660	Birge, Black and Westerman, 1979. Ref.no. 4943	2
Micropterus salmoides, largemouth bass	Chromic oxide (Cr III)	8	LC50	1170	Birge et al., 1978. Ref. no. 6199	2
Gasterosteus aculeatus, three spine stickleback	Chromium (3+) salt	10	mortality	1200	Jones, 1939. Ref. no. 2851	2
<i>Tilapia</i> sp., tilapia	Chromic chloride (Cr III)	56	change in rate of growth	1760	Shiau and Lin, 1993. Ref. no. 14617	2
Channa punctatus, snake-head catfish	Chromium	7	LC50	2000	Jana & Bandyopandhyaya, 1988. Ref. no13211	2
Poecilia reticulata, guppy	Chromic potassium sulfate (Cr III)	4	LC50	3330	Pickering and Henderson, 1964. Ref. no. 2033	2

Table 30. Summary of chromium concentrations reported to have adverse effects on sensitive freshwater fish species

1 - Cited in Chromium Criteria Document 1984 (USEPA, 1985C); 2 - Cited in USEPA AQUIRE Database; NR = not reported in AQUIRE database \* Concentration is amount of chromium in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

33

r,

Species Name	Chemical	Duration (days) or test type	Effect/ Endpoint	Concentration (ug/L)	Reference	Where Cited
Anancystis aeruginosa,					Bringmann & Kuhn, 1978.	
blue-green algae	Nickelous chloride	8	unreported mortality	1.2	Ref. no. 2463	2
Microcystis aeruginosa,						
blue-green algae	Nickel chloride	8	incipient inhibition	5	Bringmann & Kuhn, 1978	1
Selenastrum capricornutum,	Nickelous chloride	4	EC50, change in growth	63	Bermingham, Man Coillie & Vasseur, 1986. Ref. no. 12748	2
Clamydomonas reinhardtii			EC30, change in		Welbourn, 1994. Ref. no.	
green algae	Nickelous chloride	7	abundance	6.7	13711	2
Chlorella vulgaris, green algae	Nickelous nitrate	91.3	NOEC, population growth	6.9	Den Dooren Jong, 1965. Ref. no. 2849	2
Anacystis nidulans, blue-green algae	Nickel (2+) salt	0.25	change in photosynthesis	10	Azeez & Banerjee, 1987. Ref. no. 12558	2
Chlorella pyrenoidosa,			change in population		Stauber & Florence, 1987.	-
green algae	Nickel	4	growth	10	Ref. no. 12971	2
Spirulina platensis, blue-green algae	Nickel (2+) salt	0.25	change in photosynthesis	10	Azeez & Banerjee, 1987. Ref. no. 12558	2
Anabaena cylindrica,	Γ		13% reduction in doubling			
blue-green algae	Nickel sulfate	5	time	15.1	Daday et al., 1985	1
Thalassioria guillardii, diatom	Nickel	2	change in population growth	50	Wilson & Freeburg, 1980. Ref. no. 5557	2
Nostoc linckia, blue-green algae	Nickelous chloride	1	change in biochemicalprocesses	.50	Kumar & Kumar, 1985. Ref. no. 11511	2
Scenedesmus acuminata, green algae	Nickel nitrate	12	54% reduction in growth	50	Hutchinson & Stokes, 1975.	1
Navicula pelliculosa, diatom	Nickelous nitrate	7	change in population growth	100	Fezy, Spencer & Greene, 1979. Ref. no. 8347	2
Ankistrodesmus falcatus, green algae	Nickelous nitrate	14	change in biomass	100	Spencer & Greene, 1981. Ref. no. 15439	2
Pediastrum tetras, green algae	Nickelous nitrate	14	change in biomass	100	Spencer & Greene, 1981. Ref. no. 15439	2

.

Table 31. Summary of nickel concentrations reported to have adverse effects on sensitive freshwater algal and diatom species

1 - Cíted in Nickel Criteria Document 1986; 2 - Cited in USEPA AQUIRE Database

.

£

\* Concentration is amount of nickel in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

Species Name	Chemical	Duration (days) or test type	Effect/ Endpoint	Concentration (ug/L)	Reference	Where Cited
Ceriodaphnia dubia, water flea	Nickelous nitrate	ан - <sup>1</sup> 2 - 2	unspecified mortality	3.8	Kszos, Stewart & Taylor, 1992 Ref. no. 5920	2
Culex pipiens, mosquito	Nickelous chloride	7.29	ET50, emergence from larvae to adult	4.5	Suzuki, 1959. Ref. no. 2701	2
Tubifex tubifex, tubificid worm	Nickel sulfate	2	LC50	7	Brkovic-Popovic and Popovic, 1977	1
Asellus aquaticus, aquatic sowbug	Nickelous chloride	27	LC50	10	Migliore & Guidici, 1990. Ref. no. 10515	2
Moina macrocopa, water flea	Nickelous chloride	8.5	LC50	10	Wong, 1993. REf. no. 6973	2
Daphnia magna, water flea	Nickelous chloride	42	mortality	40	Munziger, 1990. Ref. no. 3063	2
Uronema pardnez, protozoan	Nickel chloride	0.833	incipient inhibition	42	Bringmann and Kuhn, 1981	1
Microregma heterostoma, paramecium	Nickel chloride	1.16	incipient inhibition	50	Bringmann & Kuhn, 1959b	1
Biophalaria glabrata, snail	Nickel (2+) salt	<u> </u>	physiological stress observed	100	Harry & Aldrich, 1963. Ref. no. 2853	2
Entosiphon sulcatum, flagellate euglenoid	Nickelous chloride	3	change in population growth	140	Bringmann and Kuhn, 1980. Ref. no. 5303	2
Anocystis imbecillis, mussel	Nickel (2+) salt	4	LC50	190	Keller & Zam, 1991. Ref. no. 108	2
Chilomas paramecium, cryptomonad	Nickelous chloride	2	change in population growth	200	Bringham, Kuhn & Winter, 1980. Ref. no. 5719	2
Juga plicifera, snail	Nickelous chloride	21	LC50	204	Chapmen, 1986. Ref. no. 11982	2
Orconectes limosus, crayfish	Nickelous chloride	30	- LC50	450	Boutet & Chaisemartin, 1973. Ref. no. 5421	2
Daphnia pulicaria, water flea	Nickel	2	LC50	697	Lind, Alto & Chatterton, 1978. Ref. no. 5081	2

¢,

12

Table 32. Summary of nickel concentrations reported to have adverse effects on sensitive species of freshwater invertebrates

1 - Cited in Nickel Criteria Document 1986; 2 - Cited in USEPA AQUIRE Database

\* Concentration is amount of nickel in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

102

ιċ:

L)

Species Name	Chemical	Duration (days) or test type	Effect/ Endpoint	Concentration (ug/L)	Reference	Where Cited
Salmo gairdner, rainbow trout	Nickel chloride	early life stage	Chronic value	<35	Nebeker et al. 1985	1
Lepomis macrochirus, bluegill	Tetracyanonickel	>0.42	acute mortality	75	Broderius, T.C. 1973. Ref. no. 8778	2
Salmo salar, atlantic salmon	Nickelous nitrate	<100	unspecified mortality	104	Grande & Anderson, 1983. Ref. no. 10982	2
Pimephales promeles, fathead minnow	Nickel	30	unspecified mortality	433.5	Lind, Alto & Chatterton, 1978. Ref. no. 5081	.2
Ictalurus punctatus, Channel catfish	Nickel chloride	7	EC50	710	Birge et al., 1981	1
<i>Cyprinus carpio,</i> common carp	Nickel sulfate	10.7	LC50	750	Blaylock & Frank, 1979	1
Gasterosteus aculeatus, three spine stickleback	Nickelous nitrate	10	100% mortality	800	Jones, 1939. Ref. no. 2851	2
Oncorhynchus mykiss, rainbow trout	Nickel (2+) salt	0.021	impaired reproduction	1000	Shaw & Brown, 1971. Ref. no. 9428	2
Tilapia nilotica, Nile tilapia	Nickelous chloride	4	change in behavior	1500	Alkahem, 1994. Ref. no. 16861	2
Micropterus salmoides, largemouth bass	Nickelous chloride	8	LC50	2020	Birge et al., 1978. Ref. no. 6199	2
Carassius auratus. goldfish	Nickelous chloride	7	LC50	2140	Birge, 1978. Ref. no. 5305	2
Ambloplites rupestris, rock bass	Nickel	4	LC50	2480	Lind, Alto, & Chatterton, 1978. Ref. no. 5081	2
Morone saxatilis, stripped bass	Nickelous chloride	4	LC50	3900	Palawski, Hunn & Dwyer, 1985. Ref. no. 11334	2
Poecilia reticulata, guppy	Nickelous chloride	4	LC50	4450	Pickering & Henderson, 1960. Ref. no. 2033	2
Oncorynchus kisutch, Coho salmon	Nickel (2+) salt	14	unspecified mortality	4500	Becker & Wolford, 1980. Ref. no. 478	2

,

•

Table 33. Summary of nickel concentrations reported to have adverse effects on sensitive freshwater fishes

1 - Cited in Nickel Criteria Document 1986; 2 - Cited in USEPA AQUIRE Database

.

٠

\* Concentration is amount of nickel in solution; shaded row indicates species used in US EPA Three Species toxicity test protocols

Table 34.	Summary of copper concentrations r	eported to have adverse effects on 15 freshw	ater fish species
-----------	------------------------------------	--	-------------------

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO3)	Reference	Where Cited
Salmo gairdneri * (fry) rainbow trout		l hr	avoidance	0.1		Folmar, 1976	3
Ictalarus fontinalis channel catfish		· .	Increased albinism	0.5		Westerman & Birge, 1978	3
Oncorhynchus Mykiss steelhead trout		4	Increased susceptibility to Yersinia ruckeri infection	2	30-60	Knittel, 1980	
Thymallus arcticus arctic grayling	copper sulfate	4	LC 50-MOR	2.58	41.3	Buhl & Hamilton, 1990	1
Salvelinus fontinalis brook trout	copper sulfate	ELS	Chronic value	3.873	37.5	Sauter et al., 1976	3
Salmo gairdneri * (fry) rainbow trout	copper spiked ambeint water (pH = 6.0)	` 168 hr	LC 50	5.1	38 +/- 3	Welsh <i>et al.</i> , 1998	
Pimephales promelas fathcad minnow	copper murate	32	матс	6.2	43.9	Spehar & Fiandt, 1986	2
Oncorhynchus tshawytscha chinook salmon	copper chloride	ELS	Chronic value	<7.4		Chapman, 1975, 1982	3
Pimephales notatus bluntnosc minnow	copper sulfate	. LC	Chronic value	8.793	194	Horning & Neiheisel, 1979	3
Oncorhynchus tshawytscha chinook salmon	ambient mixed waste (including Cu)	96 hr	LC 50	13 +/- 3	39-40	Finlayson & Wilson, 1989	
Oncorhynchus Mykiss steelhead trout	ambient mixed waste (including Cu)	96 hr	LC 50	14 +/- 4	39-40	Finlayson & Wilson, 1989	
Oncorhynchus kisutch coho salmon	copper sulfate	4	LC50 MOR	15.1	41.3	Buhl. & Hamilton, 1990	2
Salmo clarki cutthroat trout	copper chloride		LC50 or EC50	15.7	26	Chakoumakos et al., 1979	3
Salmo gairdneri * (fry) rainbow trout	copper spiked ambeint water (pH = 8.0)	168 hr	LC 50	15.9	37 +/- 2	Welsh <i>et al.</i> , 1998	
Ptychocheilus oregonensis northern squawfish	copper chloride		LC50 or EC50	18.	52-56	Andros & Garton, 1980	3
Catostomus commersoni white sucker	copper sulfate	ELS	Chronic value	20.88	45.4	McKim <i>et al.</i> , 1978	3

¢.

-e)

1.

. .

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in AQUIRE database. Cited in Copper Criteria document, (USEPA, 1984a). 2. 3.

ũψ

\$7

\* Salmo gairdneri = Oncorhynchus mykiss Shading Pimephales promelas

\_

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO <sub>3</sub> )	Reference	Where Cited
<i>Daphnia magna</i> water flea		21	LC50	1.4		Dave, 1984	3
<i>Daphnia similis</i> water flea	copper sulfate	4	LC50 MOR	4.1		Soundrapandian & Venkataraman, 1990	2
Asellus aquaticus aquatic sowbug	copper sulfate	1530	REP, MOR	5	300	DcNícola Guidici <i>et al.</i> , 1988	2
<i>Daphnia pulex</i> water flea		2	LC50	5.6		Cairus, 1978	3
Moina macrocopa water flea	copper sulfate	2	LC50 MOR	5.9		Hatakeyama & Sugaya, 1989	2
Insect community	copper	14	POP	6	88g/m <sup>3</sup>	Clementes et al., 1989	2
Gammarus pseudolimnaeus amphipod	copper sulfate	LC	Chronic Value	6.066	45	Arthur & Leonard, 1970	3
Ceriodaphnia dubia water flea	copper	7	NOEL REP	6.3	94.1	Belanger et al., 1989	2
Daphnia pulicaria water flea			LC50 or EC50	7.24	48	Lind et al., manuscript	3
<i>Daphnia lumholzi</i> water flea	copper	4	LC50 MOR	9.4	200	Vardia <i>et al.</i> , 1988	2
Corbicula manilensis Asíatic clam		70	ILC	<10		Harrison et al., 1981, 1984	3
Proasselus coxalis isopod	copper sulfate	21.3	LT50 MOR	10		DcNicola Guidici et al., 1987	2
Clistornia magnifica caddisfly	copper chloride	LC	Chronic Value	10.39	26	Nebeker et al., 1984b	3
Compeloma decisum snail	copper sulfate	LC	Chronic Value	10.88	35-55	Arthur & Leonard, 1970	3
Physa integra snail	copper sulfate	LC	Chronic Value	10.88	35-55	Arthur & Leonard, 1970	3

-

٠

.

in a factor concentrations reported to have adverse effects on 15 freshwater invertebrate species T 11- 75 0

ŧ

.

1.

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in AQUIRE database. Cited in Copper Criteria document, (USEPA, 1984a). 2 3

Ceriodaphnia dubia Shading .

.

-

- -

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µ/L)	Hardness (mg/L as CaCO3)	Reference	Where Cited
Chlorella pyrenoidosa green algae			lag in growth	1		Steeman-Nielsen & Wium-Anderse 1970	n, 3
Mixed periphyton algae		2.5	photosynthesis	2.5		Leland & Carter, 1984	4
Algae mixed culture			significant reduction in photosynthesis	5		Elder & Horne, 1978	3
Nitzchia palea diatom			complete growth inhibition	5		Steeman-Nielsen & Wium-Anderse 1970	, 3
Scenedesmus quadricuada grccn algac			metabolism	5		Peterson et al., 1984	4
Chlamydomonas reinhardtii green algae	copper sulfate	3	NOEC-LOEC	5.9	76	Garvey et al., 1991	2
<i>Chlorella</i> sp green algae			photosynthesis inhibited	6.3	·.	Gachter et al., 1973	3
Phytoplankton mixed species		5.2	reduced rate of primary production	10		Cote, 1983	• 3
<i>Selenastrum capricornutum</i> green algae	copper sulfate	3	EC50 GRO	10		Vasseur et al., 1988	1
Uroglena sp crysophyte	copper sulfate	14.35	PGR	19.7	102	Moore & Winner, 1989	<b>2</b> .
Chlorella regularis green algae			lag in growth	20		Sakaguchi et al., 1977	3
Haematococcus sp green algae		4	inhibited growth	.50		Pcarlmutter & Buchheim, 1983	3
Chlorella vulgaris grccn algac		4	IC50	62		Ferard et al., 1983	2
Anabeana strain 7120 algac			lag in growth	64		Laube et al., 1980	2
Anubcana nidulans algac			growth inhibition	100		Young & Lisk, 1972	2

ud ta hava advarca affaata on 15 frash Table 36

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in AQUIRE database.

2. 3. 4.

Cited in Copper Criteria document (USEPA, 1984a). Cited in Table II-10 (Lillebo *et al.*, 1988).

Shading Sclenastrum capricornutum

106

i ÉÌ

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO3)	Reference .	Where Cited
Salmo gairdneri rainbow trout	zinc sulfate	10 minutes	Avoidance	5.6	13-15	Sprague, 1964b	2
<i>Jordanella floridea</i> flagfish	zinc sulfate	LC	Chronic Value	36.41	44	Spchar, 1976a,b	2
Salmo salar Atlantic salmon (parr)	zinc sulfate	4 hours	EC50 avoidance	49.88	18	Sprague, 1964b	2
Oncorhynchus tshawytscha chinook salmon	zinc sulfate		acute toxicity	84	21	Finlayson & Verrue, 1982	2
Salmo clarki cutthroat trout (fingerling)	zinc sulfate		acute toxicity	90		Rabe & Sappington, 1970	2
Morone saxatilis striped bass (larvae)			acute mortality	100	38	Hughes, 1973	4
Pimephales promelas fathcad minnow	zinc sulfate	LC	Chronic value	106.3	46	Benoit & Holcombe, 1978	2
<i>Thymallus arcticus</i> arctic grayling		4	LC50 MOR	112	41.3	Buhl & Hamilton, 1990	3
Salmo trutta brown trout	zinc chloride	48 hr	LC 50	164	102	Marr et al., 1995	
Oncorhynchus mykiss steelhead trout	acid mine waste	96 hr	LC 50	167	52	Finlayson and Wilson, 1989	
Poecilia reticulata guppy	zinc sulfate	LC	Chronic value	<173	30	Picrson, 1981	2
Oncorhynchus tshawytscha chinook salmon	acid mine waste	96 hr	LC 50	178	52	Finlayson and Wilson, 1989	
Salmo trutta brown trout	zinc chloride	48 hr	LC 50	164	102	Mart <i>et al.</i> , 1995	
<i>Lepomis macrochirus</i> blucgill (fry)	zinc sulfate	3	lethal	235	51	Cairns & Sparks, 1971: Sparks et al., 1972b	2
Oncorhynchus kisutch coho salmon (fry)	zinc sulfate	1	decreased white blood cells	500	3-10	McLeay, 1975	2

.

٠

Table 37 Summary of zine concentrations reported to have adverse effects on 14 freshwater fish species

٠

1.

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in Zinc Criteria document, (USEPA, 1987). Cited in AQUIRE database. Cited in Table II-12 (Lillebo et al., 1988).

2. 3. 4.

Salmo gairdneri = Oncorhynchus mykiss; Shading = Pimephales promelas

Summary of zinc concentrations reported to have adverse effects on 15 freshwater invertebrate species Table 38.

Species Name	Chemical	Duration or Test Type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO3)	Reference	Where Cited
Asellus aquaticus aquatic sowbug	zinc sulfate	. 18	LT50 MOR	10	240	Migliore & DeNicola Guidici, 1990	3
Daphnia magna water fica	zinc sulfate	. 50	REP	25	51.9	Paulauskis & Winner, 1988	3
Ceriodaphnia reticulata water flea	zinc chloride		acute toxicity	32	45	Carlson & Roush, 1985	2
Tanytarsus dissimilis midge (embryo-3rd instar)	zinc chloride	10	LC50	36.8	46.8	Anderson et al., 1980	2
<i>Corbicula</i> sp. clam	zinc sulfate	5-30	GRO, ENZ	34-1130		Farris <i>et.al.</i> , 1989	3
Ceriodaphnia dubia water Nca	zinc	7	NOECLOEC	<b>&lt;25-25</b>	46	UCD Aquatic Toxicology Lab(unpublished results)	
Tropocyclops prasinus copepod	zinc chloride	2	EC50 motility	52	10	Lalande & Pinel-Alloul, 1986	2
Ancylus fluviatilis river limpet	zinc sulfate	100	LC50 MOR	80		Willis, 1988	3
Zooplankton (mixed species)	zinc chloride	3 weeks	reduced crustacean density and diversity	100		Marshall et al., 1981	2
Daphnia pulex wäter fica			acute toxicity	. 117	. 45	Mount & Norberg, 1984	· .·· 2 .
Anodonta imbecilis mussel	zinc sulfate	4	LC50 MOR	268		Keller & Zam, 1991	3
Physa heterostropha snail (young)	zinc sulfate		acute toxicity	303	20	Wurtz, 1962	2
Daphnia lumholzi water fica	zinc	4	LC50 MOR	437.5		Vardia et al., 1988	3
Aedes aegypti mosquito (pupa)	zinc sulfate	3	20% mortality	500	4	Abbasi et al.,1985	2
Biomphalaria glabrata snail	zinc chloride	33	REP	500	61-61.8	Munzinger & Guarducci, 1988	3

ŧ۶.

ą.

Duration given in days unless otherwise noted. Cited in Zinc Criteria document SEPA, 1987). Cited in AQUIRE database. t.

2. 3.

Ceriodaphnia dubia Shading

\*)

**4**)

**R**)

Table 39.	Summary of a	zinc	concentrations	reported	to I	have adverse	effects on	10 freshw	ater algal sp	ecies

.

Species Name	Chemical	Duration or Test Type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO3)	Reference	Where Cited
Ankistrodesmus falcatus grccn algae	zinc chloride	I	PGR	5-30		Wong & Chau, 1990	3
Navicula pelliculosa diatom	zinc chloride	1	PGR	5-30		Wong & Chau, 1990	3
Scenedesmus quadricauda grccn algac	zinc chloride	1	PGR	5-30		Wong & Chau, 1990	3
Selenastrum capricornutum grccn algac	zinc chloride	7	incipient growth inhibition	30		Bartlett et al., 1974	2
Chlamydomonas variabilis green algae		6	30% reduction in division rate	503		Balcs et al., 1983	2
Algac mixed species	zinc sulfate	5-30	BMS	540		Genter et al., 1988	3
Navicula seminulum diatom	zinc chloride	5	EC50 growth	1320		Acad. of Nat. Sci., 1960	2
Chlorella vulgaris green algae	zinc sulfate	4	EC50 growth	2400		Rachlin & Farran, 1974	2
Chlorella saccarophila green algae	zinc chloride	4	EC50	7100		Rachlin <i>et al.</i> , 1982	2
Navicula incerta diatom	zinc chloride	4	EC50	10000		Rachlin et al., 1983	2

4

٠

Duration given in days unless otherwise noted. Cited in Zinc Criteria document (USEPA, 1987). Cited in AQUIRE database.

1. 2. 3.

Selenastrum capricornutum Shading

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO <sub>3</sub> )	Reference	Where Cited
Salmo guirdneri rainbow trout		18 months	reduced survival	0.2	112	Birgc <i>et al.</i> , 1981	2
Ictalurus punctatus catfish	cadmium chloride		increased albinism	0.5		Westerman & Birge, 1978	2
Morone saxatilis striped bass	cadmium chloride		LC50 or EC50	1	34.5	Hughes, 1973	2
Oncorhynchus tshawytscha Chinook salmon (juvenile)			acute mortality	1.1	20-22	Finlayson & Verrue, 1982	3
Salmo trutta brown trout			acute montality	1.4	39.48	Spchar & Carlson, 1984	3
Salvelinus fontinalis brook trout			acute montality	<1.5	42	Carrol et al., 1979	3
Oncorhynchus mykiss steelhead trout (fry)	acid mine waste	96 hr	LC 50	1.6	52	Finlayson and Wilson, 1989	
Oncorhynchus tshawytscha Chinook salmon (fry)	acid mine waste	96 hr	LC 50	1.9	52	Finlayson and Wilson, 1989	
Oncorhynchus kisutch coho salmon (juvenile)	cadmium chloride	9	LC50	2.0	22	Chapman & Stevens, 1978	2
<i>Salmo sulur</i> Atlantic salmon	cadmium chloride	70	reduced growth	2.0	13	Peterson, 1983	2
<i>Jordanella floridea</i> flagfish	cadmium chloride	LC	Chronic value	4.4161	44-51	Carlson et al., 1982	2
Catostomus commersoni white sucker	cadmium chloride	ELS	Chronic value	7.099	44	Eaton et al., 1978	2
Salvelinus namaycush lake trout	cadmium chloride	ELS	Chronic value	7.357	44	Eaton et al., 1978	2
Esox lucius nothern pike	cadmium chloride	ELS	Chronic value	7.361	44	Eaton <i>et al.</i> , 1978	2
Micropterus dolomieui smallmouth bass	cadmium chloride	ELS	Chronic value	7.390	44	Eaton <i>et al.</i> , 1978	2

<sup>+</sup> Table 40 concentrations reported to have adverse effects on 15 freshwater fish species Sur m. of and

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in Cadmium Criteria document, (USEPA, 1984b). Cited in Table 11-7 (Lillebo et al., 1988).

ŕ

÷

2. 3.

• Salmo gairdneri = Oncorhynchus mykiss Shading Pimephales promelas

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO <sub>3</sub> )	Reference	Where Cited
Daphnia magna	cadmium chloride	LC	Chronic value	0.1523	53	Chapman et al., manuscript	3
Ceriodaphnia reticulata water flea	cadmium chloride	7	LOEC REP	0.2	240	Elnabarawy <i>et al.</i> , 1986	2
Moina macrocopa water flea	cadmium chloride	20	reduced survival	0.2	80-84	Hatakeyama & Yasuno, 1981b	3
Acanthocyclops viridis copepod	cadmium chloride	3	LC50	0.5		Braginsky & Scherban, 1978	3
Hyalella azteca scud	cadmium	42	LC50*MOR	0.53	130	Borgmann <i>et al.</i> , 1991	2
Ceriodaphnia dubia water flea	cadmium sulfate	7	GRO, REP	1	90	Winner, 1988	2
Daphnia pulex water flea	cadmium chloride	140	reduced reproduction	1	57	Bertram & Hart, 1979	3
Polypedilum nubifer midge	cadmium chloride	8	DVP	1		Hatakcyama, 1987	2
Gammarus fasciatus scud	cadmium	42	MOR	1.49		Borgmann et al., 1989	2
Astacus astacus European crayfish	cadmium	14-70	ENZ, HIS	2		Mcycr et al., 1991	2
<i>Ephemerella</i> sp mayfly	cadmium chloride	28	LC50	<3	44-48	Spchar <i>et al.</i> , 1978	3
Aplexa hypnorum snail	cadmium chloride	LC	Chronic value	3.460	45.3	Holcombe et al., 1984	3
Tanytarsus dissimilis midge	cadmium chloride	10	LC50	3.8	47	Anderson et al., 1980	3
Daphnia galvata menilotae cladoceran	cadmium chloride	22 weeks reduced biomass 4.0 Ma		Marshall, 1978a	3		
Cambarus latimus ¢rayfish	cadmium chloride	5 months	significant mortality	5	11.1	Thorp <i>et al.</i> , 1979	3

•

· · ·

Table 41. Summary of cadmium concentrations reported to have adverse effects on 15 freshwater invertebrate species

. .

1. Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage.

2. Cited in AQUIRE database.

3. Cited in Cadmium Criteria document (USEPA, 1984b).

Shading

.

Ceriodaphnia dubia

Table 42.	Summary of cadmium	concentrations r	reported to have	adverse effects	on 15	freshwater algal species
	-					~ I

Species Name	Chemical	Duration or test type <sup>1</sup>	Effect/ Endpoint	Concentration (µg/L)	Hardness (mg/L as CaCO3)	Reference	Where Cited
Asterionella formosa diatom			factor of 10 growth rate decrease	2		Conway, 1978	3
Algae mixed species	cadmium chloride		significant reduction in population	5		Gicsy et al., 1979	2
Scenedesmus quadricauda green algae	cadmium chloride		reduction in cell count	6.1		Klass <i>et al.</i> , 1974	3
Chlamydomonas reinhardtii green algae	cadmium chloride	6.7	PGR	7.5-40		Lawrence et al., 1989	2
Selenastrum capricornutum green algae	cadmium chloride	4.0	PGR	8		Thompson et al., 1987	2
Chara vulgaris	cadmium sulfate	14	IC50 GRO	9.5		Heumann, 1987	2
Scenedesmus hijugatus	cadmium sulfate	1-12	physiological	10 .		Sathya & Balakrishnan, 1987	2
Chlorella vulgaris green algae			reduction in growth	50		Hutchinson & Stokes, 1975	3
<i>Scenedesmus dimorphus</i> green algae	cadmium nitrate	2	EC50*IMM	63		Ghosh et al., 1990	2
Scenedesmus subspicatus green algae	cadmium chloride	3	EC50 BMS	100		Kuhn & Pattard, 1990	2
Algae	cadmium	14	BMS	100	-	Kerrison et al., 1988	2
Chlorella saccharophila green algae	cadmium chloride	. 4	EC50	105		Rachlin <i>et al.</i> , 1984	3 1
Anabeana flos-aquae	cadmium chloride	4	EC50	120		Rachlin et al., 1984	3
Chlorella pyrenoidosa	cadmium chloride		reduction in growth	250		Hart & Scalfe, 1977	3
Navicula incerta diatom	cadmium chloride		EC50	310		Rachlin <i>et al.</i> , 1982	3

Duration given in days unless otherwise noted. Test Types: LC-Life Cycle, ELS-Early Life Stage. Cited in AQUIRE database. Cited in Cadmium Criteria document (USEPA, 1984b). 2. 3.

14

ίŧ,

Shading Selenastrum capricornutum

	Сор	per	Zir	nc	Chror	nium	Le	ad	Cadn	nium	Nic	kel	Ars	enic
Year and Method	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.	Total (kg)	Daily Avg.
1994 Average						·				, , ,				
Concentration Method	20,900	174	50,700	423	14,700	123	3,240	27	698	6	19,800	165		
Model	16,500	141†	37,900	323†	10,500	89†	2,290	20†	*	*	13,700	117†		
1995														
Average Concentration Method	144,000	1360^	394,000	3720^	155,000	1,460^	54,400	513^	1,660	16^	#######	1,900^	20,800	196^
Model	*	*	*	*	*	*	*	*	*	*	*	*	*	*
% Increase	% Increase 872 (1)		1040	(1)	1476	(1)	2,37	6 (1)	237	(2)	1,467	/ (1)	N/	A

Table 43. Comparison of Metal Load Estimates in the Sacramento River at Greene's Landing from January Through April During a Dry Year (1994) and Wet Year (1995).

£.

1

(1) = % increase from 1994 model calculation to 1995 average concentration method

(2) = % increase from 1994 average concentration method to 1995 average concentration method

\* = Model could not be applied due to insignificant relationship between total metal concentrations and flow

 $\dagger$  = Daily average based on 117 days when flows were recorded

^ = Daily average based on 106 days when flows were recorded

The number of significant figures for load estimates was set at three due to uncertainties in flow measurements and regression analyses.

Table 44. Comparison of Metal Loads to the Delta Contributed by Sources Which Drain Into the Yolo Bypass and Sacramento River During High Flows From January Through April 1995

METAL	CONTRIBUTION	BYPASS	RIVER	TOTAL
Copper	Total (kg)	296,000	144,000	440,000
	Daily Average	2,850*	1,360†	4,210
	Percent	67	33	100
Zinc	Total (kg)	727,000	394,000	1,120,000
	Daily Average	6,990*	3,720†	10,700
	Percent	65	35	100
Chromium	Total (kg)	472,000	155,000	627,000
	Daily Average	4,540*	1,460†	6,000
	Percent	74	26	100
Lead	Total (kg)	64,700	54,400	119,000
	Daily Average	622*	513†	1,140
	Percent	54	46	100
Cadmium	Total (kg)	1,550	1,660	3,210
	Daily Average	15*	16†	31
	Percent	48	52	100
Nickel	Total (kg)	911,000	201,000	1,110,000
	Daily Average	8,760*	1,900†	10,700
	Percent	82	18	100
Arsenic	Total (kg)	22,400	20,800	43,200
	Daily Average	215*	196†	410
	Percent	52	48	100

\* = Yolo Bypass daily average based on 104 days when USGS gage station #11453000 was functional

† = Sacramento River daily average based on 106 days when flows were recorded

The number of significant figures for load estimates was set at three due to uncertainties in flow measurements and regression analyses.

II (

Table 45. Comparison of Metal Load Estimates in the Sacramento River at River Mile 44 from January Through April of a Dry Year (1994) and Wet Year (1995) Based on Metal Analyses Conducted for the Sacramento Coordinated Water Quality Monitoring Program's Ambient Monitoring Program

.

.

	Copp	)er	Zir	10	Chror	nium	Lea	ad	Cadn	nium	Nicl	kel	Arse	enic
	Total	Daily	Total	Daily	Total	Daily	Total	Daily	Total	Daily	Total	Daily	Total	Daily
Year and Method	(k <u>g</u> )	Avg.	(kg)	Avg.	(kg)	Avg.	(kg)	Avg.	(kg)	Avg.	(kg)	Avg.	(kg)	Avg.
1994						-*				<b></b>				
Average Concentration	12 000	100	20 000	241	5 5 9 0	47	1 6 4 0	14	172	1	0.440	70	7 000	(5
Method**	12,000	100	20,000	241	5,580	47.	1,040	14	125	1	9,440	19	/,800	03
% of BPTCP estimates														
(same method)	57		57		38		51		18	3	48		N/.	A
Model (estimated by														
regression)	12,600	108†	30,700	262†	7,020	60†	1,680	14†	193	2†	10,300	88†	6,680	57†
% of BPTCP estimates														
(same method)	76		81		67		73		N/	A	75		N/.	<u>A</u>
1995														
Average Concentration		~~~			46 700		10.000	100	000	0	100.000			
Method**	95,100	897	198,000	1,860	46,700	441	19,300	182	998	9	102,000	966	21,300	201
% of BPTCP estimates														
(same method)	66		50		30	)	36		60	)	51		102	2
Model (estimated by											-	-	-	
regression)	116,000	1090^	190,000	1790^	58,800	555^	20,600	194^	1,830	17^	149,000	1400^	28,100	265^
% of BPTCP estimates														
(same method)	N/A	4	N/.	A	N/	A	N/.	Α	N/	A	Ŋ/A	4	N/.	4
Minimum % Increase in load from WY94 to WY95	792		619	)	83	7	1,18	80	81	1	1,08	0	42(	)

t = Daily average based on 117 days when flows were recorded

.

^ = Daily average based on 106 days when flows were recorded

The number of significant figures for load estimates was set at three due to uncertainties in flow measurements and regression analyses.

\*\* = values reported as non-detectable were set at zero for the purposes of obtaining an average concentration.

Note: AMP model estimates were provided by Klauss Suverkropp of Larry Walker Associates

\_ \_ \_

		Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.
		Cu	Cu	Zn	Zn	Cr	Cr	Pb	Pb	Cd	Cd	Ni	Ni	As	As
		(μg/l)	(μg/l)	(μg/l)	(µg/l)	· (μg/l)	(μg/l)	(µg/l)	<u>(μg/l)</u>	(µg/l)	μg/l)	(µg/l)	(µg/l)	(µg/l)	(µg/l)
1993		:													
(norma	ıl)														
	Mean	5.56	1.83	9.61	1.94	4.65	0.60	2.81	0.11	0.06	0.02	6.90	1.37		
	SD	5.85	0.58	6.56	1.10	6.07	0.36	8.88	0.07	0.10	0.01	8.83	0.85		
	Max.	28.3	2.91	26.8	5.02	26.8	1.42	39.4	0.26	0.456	0.03	38.8	4.15		
	Min.	1.98	0.32	4.12	0.7	0.007	0.09	0.2	0.03	0.007	0.009	0.75	0.31		
	n =	19	19	19	19	19	19	19	16	19	14	19	19		
1994			· .												
(critica dry)	illy						۰ ۲								
• •	Mean	4.54	2.45	10.03	3.40	3.71	1.00	0.97	0.24	0.09	0.04	5.39	1.97	1.72	1.38
	SD	. 3.11	1.32	8.21	2.79	4.79	1.20	1.42	0.26	0.14	0.08	6.94	1.71	0.91	0.61
	Max.	14.9	9.48	39	18.5	23.1	5.39	8.98	1.38	0.74	0.55	35.8	8.52	3.98	2.4
	Min.	0.75	0.2	0.08	0.16	0.19	0.06	0.01	0.01	0.006	0.001	0.52	0.13	0.11	0.24
	n =	111	86	116	85	110	86	112	78 ່	113	79	111	8 <u>6</u>	25	24
1995											-				
(wet)															
	Mean	21.20	3.48	57.61	7.74	33.76	2.45	5.82	0.55	0.13	0.03	63.50	5.02	1.49	1.19
	SD	31.77	0.95	75.23	11.20	63.37	1.18	8.03	0.59	0.13	0.02	141.17	4.50	0.83	0.49
	Max.	162	5.4	333	70.2	312	4.78	41.2	3.87	0.568	0.11	653	26	4.41	3.03
	Min.	1.15	1.84	3.2	1.98	0.73	0.39	0.28	0.09	0.012	0.002	0.83	1.33	0.3	0.13
	<u>n</u> =	113	39	97	39	113	39	113	38	113	38	113	39	43	26

•

١ε.

Table 46. Total Recoverable and Dissolved (0.45 µm) Metal Concentrations (µg/l) in Samples Collected from All Stations Monitored during water years 1993, 1994, and 1995.

	<u> </u>														
		Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.	Total	Dis.
		Cu	Cu	Zn	Zn	Cr	Cr	Pb	Pb	Cd	Cđ	Ni	Ni	As	As
		(µg/l)	(μg/l)	(µg/l)	(μg/l)	(µg/l)	(μg/l)	(µg/l)	(µg/l)	<u>(μg/l)</u>	μġ/l)	(µg/l)	(µg/l)	(µg/l)	_(μg/l)
1993															
	Mean	3.92	2.91	6.20	2.10	1.54	0.29	0.29	0.08	0.05	0.03	1.85	0.75		
	SD	0.41		0.14		0.88		0.12		0.01		0.36			
	Max.	4.21	2.91	6.3	2.1	2.16	0.29	0.37	0.08	0.05	0.03	2.1	0.75		
	Min.	3.63	2.91	6.1	2.1	0.92	0.29	0.2	0.08	0.04	0.03	1.59	0.75		
	n=	2	1	2	1	2	1	2	1	2	1	2	1		
1994															
	Mean	5.08	2.93	12.35	4.53	3.57	1.15	0.79	0.25	0.17	0.05	4.83	1.87		
	SD	3.05	1.70	9.01	3.29	3.30	0.81	0.50	0.15	0.19	0.12	4.36	1.05		
	Max.	14.29	9.48	39	18.5	14.9	3.78	2.15	0.53	0.74	0.55	19.5	4.62		
	Min.	1.29	1.32	0.11	1.4	0.26	0.31	0.01	0.01	0.01	0.01	0.52	0.64		
	n=	46	30	49	30	46	30	.48	29	48	27	46	30		
1995															
	Mean	8.64	3.44	23.68	5.63	9.34	2.76	3.27	0.51	0.10	0.03	12.10	5.51	1.25	1.09
	SD	5.40	0.82	17.16	3.93	6.17	1.03	4.39	0.22	0.08	0.02	6.95	5.20	0.58	0.22
	Max.	28.4	5.05	71.8	22.4	29	4.78	28.7	0.99	0.474	0.11	28.3	26	2.97	1.41
	Min.	2.76	1.89	3.98	1.98	1.67	1.28	0.39	0.18	0.027	0.002	2.71	2.15	0.3	0.45
	n=	47	27	37	27	47	27	47	27	47	27	47	27	24	20

Table 47. Total Recoverable and Dissolved (0.45  $\mu$ m) Metal Concentrations ( $\mu$ g/l) in Samples Collected at Greene's Landing from January Through March of 1993, 1994, and 1995.

- -

ŧ

٠

.

.

	Cu	Zn	Cr	Pb	Cd	Ni	As
	n=36	n=36	n=31	n=33	n=38	n=37	n=1
Total vs. Diss.	r2 = 0.24	r2 = 0.19	r2 = 0.22	r2 = 0.15	$r^2 = 0.13$	r2 = 0.26	
	n=56	n=63	n=54	n=58	n=58	n=56	
Total vs. Flow	r2 = 0.56*	$\tau 2 = 0.52^{+1}$	r2 = 0.64*	r2=0.58*	r2 = 0.027	r2 = 0.6*	
	P<.001	P≤.001	P<.001	P≤001		P<.001	
	n=47	n=46	n=41	n=43	n=45	n=46	
Diss. vs. Flow	r2 = 0.3*	r2 = 0.24	r2=0.34*	r2 = 0.12	r2 = 0.11	r2 = 0.37*	
	P<.05		<u>P≤.05</u>			P<.02	
	n=30	n=32	n=29	n=29	n=30	n=29	
Total vs. TSS	$r2 = 0.7^*$	<u>1</u> 2=0.64*	r2=0.72*	r2=0.61*	r2 = 0.023		
	P<:001 ∓		P≤.001*	P<.001.		₩4 P<.001	
	n=31	n=32	n=27	n=27	n=30	n=29	
Diss. vs TSS	$r_2 = 0.1$	$r_2 = 0.065$	$r^2 = 0.047$	$r^2 = 0.25$	$r^2 = 0.015$	$r^2 = 0.14$	

Table 48. Bay Protection Toxic Cleanup Program: Summary of regression coefficients for total recoverable and dissolved (0.45 µm) metals, flow, and TSS for the Sacramento River at Greene's Landing during water year 1994.

\* = significant relationship

	Cu	Zn	Cr	Pb	Cd	Ni	As
Total vs. Diss.	n= 26 r2 = 0.59* P<:002	n= 26 r2 =0.022	n=26 $r^{2}=0.37$	$n \equiv 26$ $r^2 = 0.41$ * P < .05	n=31 r2=0.029	n=29 r2 = 0.099	n=17 r2 =0.004
Total vs. Flow	n=51 $r^2 = 0.12$	n=39 r2=0.06	n=51 $r^{2}=0.18$	n=49 r2 = 0.0054	n=50 r2=0.077	n=52 r2 =0.23	n=24 r2=0.042
Diss. vs. Flow	n= 28 r2 = 0.0026	n=27 r2=0.011	n=27 r2 =0.14	n=26 r2 = 0.0000069	n= 33 r2 =0.016	n=29 r2=0.051	n=19 r2 =0.00082
Total vs. TSS	n=31 r2 =0.85* P<.001	n= 30 +r2 =0.52* P< 005	n=31 r2=0.78* P<001	n= 29 r2 =0.16	n=30 r2=0.92* P<001	n=31 r2 =0.081	n=21 r2=0.0013
Diss. vs TSS	n=23 r2 = 0.43* P<.05	n=22 r2=0.000051	n=22 $r^2 = 0.12$	n=21 r2=0.47* P<.05	n= 28 r2 = 0.0087	n=23 r2 = 0.0042	n= 16 r2 =0.012

1

.

Table 49. Bay Protection Toxic Cleanup Program: Summary of regression coefficients for total recoverable and dissolved (0.45 µm) metals, flow, and TSS for the Sacramento River at Greene's Landing during water year 1995.

1

\* = significant relationship

**.** -

	Cu	Zn	Cr	Pb	Cd	Ni	As
	n= 62	n=60	n=56	n=59 s	n=67	n=69	n=18
Total vs. Diss.	r2 = 0.32*	r2 =0.11	r2 = 0.55*	r2 = 0.46*	r2 =0.12	* r2 = 0.29*	r2 = 0.014
	P<.02		P<:001	P<.001		P≤.02	
	n= 107	- n= 102	n=105	n= 107	n=108	n=108	n=25
Total vs. Flow	r2 = 0.26*	r2=0.24*	r2=0.38*	r2 = 0.15	r2 =0.018	r2 = 0.45*	r2 =0.063
	P<.01	P<.02	P<:001			P<.001	
	n=75	n= 73	n= 68	n=69	n= 78	n=75 ∔	n=20
Diss. vs. Flow	r2 = 0.11	r2 = 0.078	r2=0.58*	r2=052*	r2 = 0.039	• <b>1</b> =0.28*	$r^2 = 0.14$
			P<:001	P≪01		P< 02	
	– n= 61	n=62	n=60	n=58	n=60	n=60	n=21
Total vs. TSS	r2=0.83*	r2=0.6*	r2=0.81*	r2 =0.22	r2 =0.039	$r2 = 0.3^{*}$	r2 = 0.0013
	P<001	- P<001	P<:001			P<02 ↓	
	n=54	n=54	n= 49	n=48	n= 58	n=52	n=16
Diss. vs TSS	r2 =0.17	r2 =0.023	r2 =0.28	2=256	r2 =0.069	r2 =0.087	r2 =0.012
		·		<sup>4</sup> P≤1001 <sup>4</sup> a**			

Table 50. Bay Protection Toxic Cleanup Program: Summary of regression coefficients for total recoverable and dissolved (0.45 µm) metals, flow, and TSS for the Sacramento River at Greene's Landing for the combined water years 1994 and 1995.

\* = significant relationship

4

120

	Cu	Zn	Cr	Pb	Cd	Ni	As
	n=22	n= 22	n= 31	n=22	n= 22	n=14	n=22
	r2 =0.19	r2 =0.0012	N/A: all values	r2 =0.0053	r2 = 0.036	r2 =0.62*	r2 =0.70*
Total vs. Diss.			< detection limit			.01 <p<.02< th=""><th>P&lt;.001</th></p<.02<>	P<.001
	n=22	n=22	n=22	n=22	n=22	n=14	n=22
	$r^2 = 0.35$	r2 = 0.2072	r2 = 0.011	r2 =0.12	r2 = 0.076	r2 =0.68*	r2 = 0.14
Total vs. Flow						.005< P<.01	
	n=22	n=22	n=22	n=22	n=22	n=14	n=22
	r2 =0.024	r2 =0.15	N/A: all values	r2 =0.056	r2 =0.10	r2 =0.51	r2 = 0.23
Diss. vs. Flow			< detection limit				
	n=22	n=22	n=22	n=22	n=22	_n=14	n=22
	r2 =0.84	r2 =0.323	r2 =0.17	r2 =0.20	r2 = 46	r2 = 0.74	r2 =0.0132
Total vs. TSS	P<.001				.02 <p<.05< th=""><th>.002<p<.005< th=""><th></th></p<.005<></th></p<.05<>	.002 <p<.005< th=""><th></th></p<.005<>	
	n=22	n=22	n=22	n=22	n=22	r.=14	n=22
	r2 =.096	r2 =0.015	N/A: all values	r2 =0.012	r2 =0.056	r2 =0.45	r2 =0.075
Diss. vs TSS			< detection limit				

Table 51. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved metals, flow, and TSS for the Sacramento River at River Mile 44 for WY94. Sampling dates ranged from 10/4/93 - 9/13/94.

.

.

\* = significant relationship

	Cu	Zn	Cr	Pb	Cd	Ni	As
	n=24	n=24	n=24	n= 24	n= 24	n = 12	n=22
	r2 =0.20	r2 =0.33	r2 =0.00013	r2 =0.0085	N/A: all values	r2 =0.053	r2 =0.32
Total vs. Diss.					< detection limit		
	n=21	n=21	n=21	n=21	n=21	n=10	n=19
	r2 =0.13	r2 =0.080	r2 = 0.069	r2 = 0.22	r2 = 0.071	r2 = 0.00034	r2 = 0.16
Total vs. Flow					-		
	n=21	n=21	n=21	n=21	n=21	n=10	n=21
	r2 = 0.059	r2 = 0.0002	r2 = 0.0032	r2 = 0.021	N/A: all values	r2 = 0.0035	$-r2 = 0.51^{*}$
Diss. vs. Flow					< detection limit		.01 <p<02< th=""></p<02<>
		n=24	n=24	n=24.	n=24	n=12	n=22
	r2 = 0.72*.	r × r2 = 0.54*	r2=0.47*	r2(≡ 0.69*	r2=0.74*	$r^2 = 0.431$	$r^2 = 0.001032$
Total vs. TSS	P<.001	-:005< P<:01			- P<001		
	n=24	n=24	n=24	n=24	n=24	n=12	n=24
-	r2 = 0.096	r2 = 0.628	r2 = 0.019	r2 = 0.0003005	r2 = 5X10(-16)	r2 =0.067	r2 =0.085
Diss. vs TSS			<u> </u>				<u> </u>

Table 52. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved metals, flow, and TSS for the Sacramento River at River Mile 44 for WY95. Sampling dates ranged from 10/25/94 - 9/25/95.

\* = significant relationship

	Cu	Zn	Cr	Pb	Cd	Ni	As
	n=46	n=46	n=46	n=46	n=46	n= 26	n= 44
	r2 = 0.088	r2 =0.060	r2 = 0.010	r2 = 0.042	r2 = 0.0034	r2 = 0.20	r2=0.52*
Total vs. Diss.							P<.001
	n=43	n=43	n=43	n=43	n=43	n=24	n=41
	r2 =0.27	r2 = 0.015	r2 = 0.27	r2 =0.31*	r2 =0.072	r2 =0.24	r2=0.36
Total vs. Flow				.02 <p<.05< th=""><th></th><th></th><th>.02<p<.05< th=""></p<.05<></th></p<.05<>			.02 <p<.05< th=""></p<.05<>
	n=43	n=43	n=43	n=43	n=43	n=24	n=43
	r2 =0.0053	r2 =0.024	r2 =0.032	r2 =0.0048	r2 =0.031	r2 =0.11	r2=0.56*
Diss. vs. Flow							P<.001
	n= 46	n= 46	n= 46	n=46	n= 46	n= 26	n=44
	r2 =0.75*	r2 =0.15	r2=0.50*	r2=0.66*	r2=0.61	r2 =0.46	r2 =0.062
Total vs. T <u>SS</u>	- P<.001		P<.001	P<:001	P<:001	.001 <p<.002< th=""><th></th></p<.002<>	
	n= 46	n=46	n= 46	n= 46	n= 46	n= 26	n= 46
	r2 =0.024	r2=0.39	r2 =0.000031	r2 =0.024	r2 =0.018	r2 =0.15	r2 =0.19
Diss. vs TSS		.005 <p<.01< th=""><th></th><th></th><th></th><th></th><th></th></p<.01<>					

Table 53. Ambient Monitoring Program: Summary of regression coefficients for total recoverable and dissolved metals, flow, and TSS for the Sacramento River at River Mile 44 fof WY94-WY95. Sampling dates ranged from 10/4/93 - 9/25/95.

•

.

\* = significant relationship

•

•

Table 54. Bay Protection Toxic Cleanup Program: Summary of total recoverable metals regressed against other metals for samples collected from the Sacramento River at Greene's Landing during the critically dry Water Year 1994 (upper right) and wet Water Year 1995 (lower left).

	Total Cu	Total Zn	Total Cr	Total Pb	Total Cd	Total Ni
		n=54	n=54	n=54	n=54	n=54
l'otal Cu		P≤.005	$P_{2} = 0.38^{+}$	$r2 = 0.78^{-1}$ P<.001	P>0.50	r2 = 0.84* P< 001
Total Zn	n=37 r2 =0.69*		n=54 r2 =0.84*	n=56 r2=0.80*	n=56 r2 =0.10	n=54 $r^2 = 0.84 *$
	P<.001	a the second	P≤.001	P< 001	0.50>P>0.20	P<.001
	n=48	n=37		n=54	n=43	n= 54
I otal Cr	r2 ≡0.83* P<.001	P≤.001		r2 =0.51* P<.001	$P_{P>0.50}$	r2 =0.9/* P<:001
	n=48	n=37	n=48		n=56	n=54
Total Pb	r2 =0.14 0.50>P>0.20	r2 =0.28 0.10>P>0.05	r2 =0.14 0.50>P>0.20		r2 =0.027 P>0.50	r2=0.83* P<001
	n=48	n=37	n=48	n=48		n=54
Total Cd	r2 =0.82*	r2 =0.61*	r2 =0.54*	r2 = 0.12		r2 = 0.072
L	P<.001	P<.001	P<.001	0.50>P>0.20		P>0.50
	n=48	n=37.	n=48	n=48	n=48	
	r2 =0.45*	r2=0.41*	r2 =0.51*	r2 =.026	r2 = 0.18	Martin - See
Total Ni	P<.002	P<.02	P<.001	P>0.50	0.50>P>0.20	States of the second second

1

\* = significant relationship

Table 54 (cont). Bay Protection Toxic Cleanup Program: Summary of total recoverable metals regressed against other metals for samples collected from the Sacramento River at Greene's Landing during the combined 1994 and 1995 Water Years.

.

.

	Total Cu	Total Zn	Total Cr	Total Pb	Total Cd	Total Ni
Total Cu		n=94 r2 =0.77* P<:001	n=102 r2 = 0.85* P<.001	n=102 r2=0.22* P<.05	n=102 r2 =0.012 P>0.50	n=102 r2=0.59* P<:001
Total Zn			n=94 r2 =0.79* P<.001	n= 94 r2 = 0.34* P<.001	n=94 r2 =0.00002 P>0.50	n=94 r2 = 0.57 * P<.001
Total Cr				n=102 r2 =0:25* P<.02	n= 102 r2 = 0.00058 P>0.50	n= 102 r2 =0.69* P<.001
Total Pb					n=104 r2 =0.00079 P>0.50	n=102 r2 = 0.80* P<.001
Total Cd						n=102 r2 =0.01 P>0.50
Total Ni						

\* = significant relationship



Figure 1. Map of the Sacramento-San Joaquin River Delta and its major tributaries. Numbers refer to stations sampled during the Delta studies and are described in Appendix A. Sample dates are identified in Table 1.





Figure 3. Regression of flow versus total copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 4. Regression of flow versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 5. Regression of flow versus total recoverable chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.

16-


Figure 6. Regression of flow versus total recoverable lead concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.

-



Figure 7. Regression of flow versus total recoverable nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 8. Regression of TSS versus total recoverable copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 9. Regression of TSS versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 10. Regression of TSS versus total recoverable chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.







Figure 12. Regression of TSS versus total recoverable nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 13. Regression of flow versus dissolved (0.45 µm) copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 14. Regression of flow versus dissolved (0.45 µm) chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 15. Regression of flow versus dissolved (0.45 µm) nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 16. Regression of TSS versus dissolved (0.45 µm) copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 17. Regression of TSS versus dissolved (0.45  $\mu$ m) chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 18. Regression of TSS versus dissolved (0.45 µm) nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 19. Regression of total recoverable chromium versus dissolved (0.45  $\mu$ m) chromium in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 20. Regression of total recoverable lead versus dissolved (0.45 µm) lead in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 21. Regression of flow versus total recoverable copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 22. Regression of flow versus dissolved (0.45 µm) copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 23. Regression of flow versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 24. Regression of flow versus dissolved (0.45 µm) zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.

]49



Figure 25. Regression of flow versus total recoverable chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 26. Regression of flow versus dissolved (0.45 µm) chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.







Figure 28. Regression of flow versus dissolved (0.45  $\mu$ m) lead concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 29. Regression of flow versus total recoverable cadmium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 30. Regression of flow versus dissolved (0.45  $\mu$ m) cadmium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.

.







Figure 32. Regression of flow versus dissolved (0.45 µm) nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 33. Regression of flow versus total recoverable arsenic concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.

• ·



Figure 34. Regression of flow versus dissolved (0.45 µm) arsenic concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 35. Regression of TSS versus total recoverable copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 36. Regression of TSS versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 37. Regression of TSS versus total recoverable cadmium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 38. Regression of flow versus total recoverable cadmium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.

163

**-** -



Figure 39. Regression of total recoverable copper versus dissolved (0.45 µm) copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 40. Regression of total recoverable lead versus dissolved (0.45  $\mu$ m) lead concentration in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 41. Regression of TSS versus total recoverable copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.


¢

Figure 42. Regression of TSS versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.

÷



Figure 43. Regression of TSS versus total recoverable chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 44. Regression of TSS versus total recoverable nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Flow (cfs)

Figure 45. Regression of flow versus total recoverable copper concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 46. Regression of flow versus total recoverable zinc concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.







Figure 48. Regression of flow versus total recoverable nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 49. Regression of TSS versus dissolved (0.45  $\mu$ m) chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 50. Regression of TSS versus dissolved (0.45  $\mu$ m) lead concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 51. Regression of TSS versus dissolved (0.45 µm) nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 52. Regression of flow versus dissolved (0.45 µm) chromium concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 53. Regression of flow versus dissolved (0.45 µm) lead concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 54. Regression of flow versus dissolved (0.45  $\mu$ m) nickel concentration in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 55. Regression of total recoverable copper and total recoverable zinc concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 56. Regression of total recoverable copper and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.







Figure 58. Regression of total recoverable copper and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.

**] 8**3



Figure 59. Regression of total recoverable zinc and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 60. Regression of total recoverable zinc and total recoverable lead concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.







Figure 62. Regression of total recoverable chromium and total recoverable lead concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.



Figure 63. Regression of total recoverable chromium and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.

1.88



Figure 64. Regression of total recoverable lead and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1994.







.

Figure 66. Flow and total suspended solids (TSS) pattern in the Sacramento River at Greene's Landing during low flow conditions from January through March of 1994.



Figure 67. Regression of total recoverable copper and total recoverable zinc concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 68. Regression of total recoverable copper and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 69. Regression of total recoverable copper and total recoverable cadmium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



e

Figure 70. Regression of total recoverable copper and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 71. Regression of total recoverable zinc and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



¢

1

Figure 72. Regression of total recoverable zinc and total recoverable cadmium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 73. Regression of total recoverable zinc and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.

Q.



ł

Figure 74. Regression of total recoverable chromium and total recoverable cadmium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



Figure 75. Regression of total recoverable chromium and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995.



۰.

Figure 76. Regression of flow versus total suspended solids in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995. Solid arrow represents a first flush event with very high suspended solids which was preceeded by a low flow period (open arrows).





ø


¢

,

Figure 78. Flow and total suspended solids (TSS) pattern in the Sacramento River at Greene's Landing during high flow conditions from January through March of 1995.

203

į.

.



Figure 79. Regression of flow versus total suspended solids in water samples collected from the Sacramento River at Greene's Landing during Water Year 1995 without first flush and pre-first flush values.

(1



Figure 80. Regression of total recoverable copper and total recoverable zinc concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 81. Regression of total recoverable copper and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.

)



Figure 82. Regression of total recoverable copper and total recoverable lead concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.





Figure 83. Regression of total recoverable copper and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 84. Regression of total recoverable zinc and total recoverable chromium concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 85. Regression of total recoverable zinc and total recoverable lead concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 86. Regression of total recoverable zinc and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.







Figure 88. Regression of total recoverable chromium and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.

ł



Figure 89. Regression of total recoverable lead and total recoverable nickel concentrations in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.



Figure 90. Regression of flow versus total suspended solids in water samples collected from the Sacramento River at Greene's Landing during Water Years 1994 and 1995.

#### **APPENDIX A:**

.

ł

1

## List of Site Locations

Site numbers correspond to numbers in Figure 1.

<u>Sacramento River @ Greene's Landing (site 1)</u>: Sacramento River sampled from end of the U.S. Bureau of Reclamation water quality pier off Randall Island Road. Site is about three miles downstream of Hood. Samples collected at outgoing tide.

<u>Sacramento River (a) Hood (site 2)</u>: Sacramento River samples collected by boat from mid channel off steps on east bank of River upstream of Hood. Samples collected at outgoing tide.

<u>Mokelumne River (site 3)</u>: Samples collected from shore approximately one mile downstream of confluence of Cosumnes River off New Hope Road. Samples collected at outgoing tide.

<u>Ulatis Creek (site 4)</u>: Samples collected from mid channel under bridge at Brown Road. Ulatis Creek discharges into Cache Slough.

<u>Skag Slough (site 5)</u>: Sampled from middle of Liberty Island Road bridge. Skag Slough is the secondary channel draining the Yolo Bypass. Samples collected at outgoing tide.

**Prospect Slough (site 6)**: Sampled by boat at junction of Prospect Slough and Toe drain. Prospect Slough is the main channel draining the Yolo Bypass. Samples collected at outgoing tide.

**Duck Slough (site 7)**: Samples collected from middle of drain off discharge pump platform. Drain discharges into Miners Slough at Five Points Marina.

Sacramento River @ Rio Vista (site 8): Sacramento River samples collected at low tide in mid channel by boat about one mile downstream of HWY 12 bridge.

San Joaquin River @ Vernalis (site 9): San Joaquin River samples collected off middle of Airport Way Bridge (County Road J3).

<u>**Paradise Cut (site 10)</u>**: Samples collected from middle of south channel off Paradise Road bridge.</u>

Old River @ Tracy Blvd (site 11): Samples collected in mid channel off Tracy Blvd. bridge.

**French Camp Slough (site 12)**: Samples collected from mid channel off Manthey Road bridge. Slough is discharged into the San Joaquin River about one mile upstream of Highway 4 Bridge.

San Joaquin River @ City of Stockton (site 13): San Joaquin River samples collected by boat off entrance to McLeod Lake.

<u>Middle River @ Bullfrog (site 14)</u>: Middle River samples collected on an incoming tide at mid channel off Bacon Island Road Bridge.

San Joaquin River @ Point Antioch (site 15): San Joaquin River samples collected from boat in mid channel at low tide off Point Beenar. Site is about five miles upstream of confluence of Sacramento River.

<u>Chipps Island</u>: Sacramento River samples collected from boat in mid channel off Chipps Island at lower low tide.

Grizzly Bay: Sample collected by boat at lower low tide in mid Bay off pilings.

<u>Martinez</u>: Samples collected by boat at lower low tide in mid channel about two miles downstream of HWY 680 bridge.

### **APPENDIX B:**

# **Raw Metal Analysis Data**

.

1

1

.

		1		Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
1/11/93		GL 22	Greene's Landing	4 2 1		61		2.16		037
1/13/93	<u> </u>	GL 23	Greene's Landing	-1.2.1	2 91	- 0.1	21	2.10	0.29	0.57
1/14/93		GL 24	Greene's Landing	3.63		63		0.92	0.27	02
3/23/93	1030	3	Sac R denth 1	9.92		26.8		111		1 53
3/23/93	1030	1	Sac R - surface 1	85		243		7 28		13
3/23/93	1030	2	Sac R - surface 2		2 34		2.63	/.20	1.01	
3/23/03	1030	4	Sac R - denth 2		2.21		3.63		1 42	
4/13/03	1700	36	Sac River @ Delta		0.32		1 34		1 13	
7/7/03	1510	135	Middle R @ Bullfrog I dg	2 54	0.52	677	1,54	0.007	1.15	0.46
7/7/03	1510	136	Middle R. @ Bullfrog Ldg.		1.67	0.77	1 15	0.007	0.45	0.40
7/7/03	1750	140	S I River @ Vernalis	638	1.07	16.1	1.15	838	0.45	1 43
7/7/35	1750	150	S.J. River @ Vernalis	0.56	1.62	10.1	1.52	0.00	0.62	1.45
7/10/02	1020	150	S.J. River @ Antioch	1.65	1.05	0.00	1.54	1.00	0.05	0.05
7/19/93	1038	151	S.J. River @ Antioch	4.05	2.22	9.98	2.06	4.09	0.70	0.03
7/19/95	1030	152	S.J. River @ Hood	26		6.46	2.00	205	0.78	2.05
7/19/93	1300	155	Sac, River @ Hood	3.0	1 42	0.40	1.12	2.85	0.22	2.85
7/19/93	1300	134	Sac, River @ Hood	2.51	1.42	6.06	1.12	2.62	0.32	0.00
7/20/93			Sac R. @ Rio Vista	3.51	1.50	0.90	1 21	2.03	0.41	0.62
7/20/93			Sac R. @ Rio Vista		1.30		1.31		0.41	
1/20/93	1011	F3	Sac R. @ RIO VIsta	1.00	1.45		0.7	0.00	0.5	
8/3/93	1311	193	Mokelumne River	1.98		0.15	0.40	0.00	0.00	0.3
8/3/93	1311	194 E 11	Mokelumne River		1.62		2.49		0.09	
8/3/93		[F-1]	Sac R. @ Rio Vista		2.4	1.55	2.64	0.00	1.14	
8/3/93		F-12	Sac R. @ Rio Vista	3.17		4.55		2.06		0.32
8/3/93	ļ	F-10/QC	Sac. River @ Hood	3.77		5.91.		3.25	0.0/	0.61
8/3/93	,x	F-8	Sac. River @ Hood		1.61		1.47		0.36	
8/3/93		F-9	Sac. River @ Hood	4.18		7.41		3.27		0.53
8/17/93	1200	207	Middle R. @ Bullfrog Ldg.	28.3		6.66		26.8		39.4
8/17/93	1200	208	Middle R. @ Bullfrog Ldg.		1.73		1.31		0.58	
8/17/93	1450	221	S.J. River @ Vernalis	4.49		11.1		5.7		1.13
8/17/93	1450	222	S.J. River @ Vernalis		1.5		0.96		0.64	
9/14/93	1200	246	Mokelumne River	3.19		4.84		1.08		0.45
9/14/93	1200	247	Mokelumne River	2.8		4.12		1.51		0.5
9/14/93	1200	248	Mokelumne River		1.6		3.16		0.09	
9/14/93		13 CF	Sac R. @ Rio Vista	2.98		6.08		2.11		0.21
9/14/93		14 CF	Sac R. @ Rio Vista		1.97		1.4		0.56	
9/14/93		15 CF	Sac R. @ Rio Vista		1.86		0.88		0.59	
9/14/93		16 CF	Sac. River @ Hood	3.76		16		2.52		0.3
9/14/93		17 CF	Sac. River @ Hood		2		5.02		0.36	
10/4/93	2030	269	Sac. River @ Freeport		2.26		3.84		0.99	
10/4/93	2030	270	Sac. River @ Freeport	1.69		1.26		1.08		0.45
10/4/93	1100	272	Sac. River @ Freeport	2.34		4.67		1.04		0.18
10/4/93		271		2.24		3.25		1.14		0.18
10/4/93		273		2.7		2.99		1.14		0.22
10/14/93	1251	298	Mokelumne River	1.77		3.37		0.54		0.26
10/14/93	1251	299	Mokelumne River		1.37		1.24		0.11	
10/14/93		18 CF	Sac R. @ Rio Vista	3.48	-	12.5		2.36	_	0.27
10/14/93		19 CF	Sac R. @ Rio Vista		1.91		2.64		0.3	
10/14/93	ļ	20 CF	Sac. River @ Hood	2.71	,	8.55		1.57		0.31
10/14/93	t	21 CF	Sac. River @ Hood		1.38	[	1.29		0.22	
10/14/93	<u> </u>	22 CF	Sac. River @ Hood		1.39		0.95	1	0.34	
10/29/93	1030	312	Middle R. @ Bullfrog Ldg.	1.59		1.34	<u> </u>	0.41		0.13
10/29/93	1030	313	Middle R. @ Bullfrog Ldg.		1.47		0.62		0.24	
10/29/93	<u> </u>	23 CF	S.J. River @ Antioch	2.72		4.99		1.34		0.03
10/29/93	<u> </u>	24 CF/OC	S.J. River @ Antioch	1.72	l	1.68		0.19		
10/29/93	1	25 CF/QC	S.J. River @ Antioch	<u>-</u>	2.73		3.18	<u> </u>	2.62	

D		Ct. 1: #		Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
10/29/93		26 CF	S.J. River @ Stockton	2.85		5.55		0.83		1.18
10/29/93		27 CF	S.J. River @ Stockton	2.66		4.96		1.16		1.36
10/29/93		28 CF	S.J. River @ Stockton		1.98		4.5		0.15	
10/29/9?		323	S.J. River @ Vernalis	2.83		9.48		2.62		0.14
10/29/93		324	S.J. River @ Vernalis		1.09		0.47		0.2	
11/10/93		29 CF	Greene's Landing	5.16		5.5		1.19		0.28
11/10/93		30 CF A	Greene's Landing		1.62		1.6		0.63	
11/10/93	·	30 CF B	Greene's Landing		1.81		1.4		0.71	
11/11/93		31 CF	Greene's Landing	2.18		5.3		1.1		0.26
11/11/93		32 CF	Greene's Landing		1.43		1.4		0.3	
11/11/93		33 CF	Greene's Landing	2.44		4.9		0.83		0.52
11/11/93		34 CF	Greene's Landing		2.04		6		0.38	
11/11/93		35 CF	Greene's Landing	2.94		6		1.15		0.62
11/11/93		36CF	Greene's Landing		1.77		4.4		0.33	
11/12/93		37 CF A	Greene's Landing	3.45		7.8		1.13		0.58
11/12/93		37 CF B	Greene's Landing	2.62		6.4		1.21		0.51
11/12/93		38 CF	Greene's Landing	3.09		9.9		1.16		0.54
11/12/93		39 CF	Greene's Landing		1.72		2.1		0.32	
11/29/93		40 CF	S.J. River @ Antioch	2.69		2.3		1.86		0.07
11/29/93		41 CF	S.J. River @ Stockton	2.66		8.2		0.98		0.95
12/13/93	[	42 CF	Sac R. @ Rio Vista	2.97	1.50	4.6		1.56		0.36
12/13/93	<u> </u>	43 CF	Sac R. @ Rio Vista		1.58		0.71	2.00	0.72	
12/13/93		44 CF	Sac. River @ Hood	4.38		1.5		3.99		0.64
12/13/93		44 CF	Sac. River @ Hood	4.35	0.16	/.6	0.00	3.4	0.10	0.63
12/13/93		45 CF	Sac. River @ Hood		2.16		0.38		0.19	
1/10/94		GL 21	Greene's Landing	2 (0	1.40	10.5	4.5	2.25	0.32	0.41
1/10/94	ļ	40 CF	S.J. River @ Antioch	3.08	2 02	10.5	<u> </u>	3.33	0.12	0.41
1/10/94		47 CF	S.J. River @ Antioch	2.06	3.82	10.2	2	0.20	0.12	01
1/10/94		40 CF	S.J. River @ Stockton	2.90		10.5		0.30		0.1
1/10/94		48 CF	S.J. River @ Stockton	2.70	267	10.0	10	0.34	0.00	0.74
1/10/94	014	49 CF	S.J. River @ Slockton	2.04	2.07		10	0.56	0.08	0.11
1/11/94	914	410	Middle R. @ Builfog Ldg.	2.00		2.2		0.30		0.11
1/11/94	914	411	Middle R. @ Bullfrog Ldg.	0.75	2.01	1.7	1.2	0.24	0.30	0.05
1/11/94	914	412	Middle R. @ Builfog Ldg.		2.01		1.2		0.39	
1/11/94	914	423	S.J. River @ Vernalis		2.47		0.39		0.17	
1/11/94	914	420	S.J. River @ Vernalia	1.51	1.95	2.5	0.5	1 10	0.74	0.06
1/11/94	714	421	Groopo's Londing	1.51	1.01		82	1.19	2 40	0.00
1/13/94	<u> </u>	65 4	Greene's Landing	6 11	4.01	101	0.2	18	2.49	1 23
1/13/94		65 B	Greene's Landing	6.64		11.2		4.0		1.25
1/18/04		25	Greene's Landing	1 20		37		0.26		0.02
1/10/04		23	Greene's Landing	2.96		103		0.20	·	0.02
1/23/94		27	Greene's Landing	2.90	1 32	10.5	18	0.00	0.48	0.10
1/24/94	+	26	Greene's Landing	271	1.52	133	1.0	1 45	0.40	0.67
1/24/94	·····	120	Greene's Landing	2.71	1.33	10.0	1.4		0.37	
1/25/94		128	Greene's Landing	2.01		9.5		1.45		0.56
1/26/94		30	Greene's Landing	3 53		12.5		2.54		1.14
1/26/94		31	Greene's Landing	1 2.25	1.79	1	8.5		0.72	t
1/27/94	<u> </u>	33	Greene's Landing	<u> </u>	2.11	1	3.9		0.81	<u> </u>
1/28/94	+	32	Greene's Landing	632		18		4.61		1.08
1/28/94	+	35	Greene's Landing	7.74		136		5.43		0.93
1/28/94		36	Greene's Landing		36	1.5.0	48	- 0.40	1.54	
1/20/04	900	40	Greene's Landing		3 18	+	26	<u> </u>	1.24	+
1/30/94	200	38	Greene's Landing	6.21	1 2.10	13.4	+- <u></u>	3.95		0.87
1/30/94	1000	42	Greene's Landing		3.27	1	4.2	1	1.32	+

-	· 1			Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
1/31/94		41	Greene's Landing	5.31		20.3		3.31		0.78
2/1/94		44	Greene's Landing	3.43		11.2		1.87		0.31
2/1/94		48	Greene's Landing		4.94		3		0.94	
2/2/94		43	Greene's Landing	4.09		4.3		2.14		0.51
2/5/94	1700	55	Greene's Landing		1.92		5.6		0.86	
2/7/94		50	Greene's Landing	nd		0.14		nd		nd
2/7/94		53	Greene's Landing		1.84		2.5		0.48	
2/8/94		51	Greene's Landing	0.16		0.16		nd		nd
2/8/94		52	Greene's Landing	3.04		11.8		1.64		0.51
2/9/94		54	Greene's Landing	5.76		16.8		4.25		1.58
2/10/94		56	Greene's Landing	13.34		39		14.85		2.15
2/10/94	930	58	Greene's Landing		5.33		7.3		2.58	
2/11/94	1000	61	Greene's Landing		6.12		18.5		2.64	
2/11/94	1600	62	Greene's Landing	nd		nd		nd		nd
2/12/94		60	Greene's Landing	10.16		28.8		7.91		1.63
2/16/94	700	63	Greene's Landing	6.67		24.7		5.31		0.88
2/16/94	700	64	Greene's Landing							
2/17/94		67	Greene's Landing	4.05		19.8		2.78		1.07
2/17/94		68	Greene's Landing		2.23		4.6		1.07	
2/18/94	1200	70	Greene's Landing	1	1.94		3.2		0.67	
2/19/94		69	Greene's Landing	4.09		11.9		3.02		0.87
2/19/94	1400	72	Greene's Landing	1	2.26		2.9		0.86	
2/19/94	1400	71 A	Greene's Landing	5.05		17.3		4.28		0.8
2/19/94	1400	71 B	Greene's Landing	6.63		13.6		3.96		0.95
2/20/94	1550	74	Greene's Landing		2.11		3		0.98	
2/21/94		73	Greene's Landing	7.12		21.8		5.64		1.16
2/21/94	1600	76	Greene's Landing		3.05		6.4		1.5	
2/22/94		75	Greene's Landing	14.29		22.5		6.65		1.39
2/22/94		77	Greene's Landing	10.74	1	28.8		10.24		1.84
2/22/94	1600	79	Greene's Landing		3.14		4.5		1.49	
2/23/94		81	Greene's Landing	12.05	1	33.4		14.9		2.02
2/23/94	1700	82	Greene's Landing		3.01		3.7		0.31	
2/24/94	1	83	Greene's Landing	7.16		19.7	,	6.68		1.04
2/24/94	1700	84	Greene's Landing	1	9.48		8.4		3.78	
2/25/94		85	Greene's Landing	5.94		14.6		4.5		0.82
2/25/94	1800	86	Greene's Landing		2.56		3.8		1.81	
2/27/94		87	Greene's Landing	6.74	1	20.3		5.73		1.28
2/28/94		89	Greene's Landing	4.86		11.7		4.02		0.71
2/28/94	1200	90	Greene's Landing		2.29		3.8		1.19	
3/1/94		91	Greene's Landing	4.24		10.1		2.76		0.73
3/1/94		93	Greene's Landing		3.03		3.4		0.87	
3/4/94		95	Greene's Landing	4.61		11.2		3.1		0.61
3/4/94	1200	96	Greene's Landing		2.32		2.3		0.6	
3/9/94	1130	100	Greene's Landing	T		0.23				0.01
3/9/94	1130	101	Greene's Landing			0.02				
3/9/94	1130	102	Greene's Landing	1		1.62				0.02
3/9/94	1130	103	Greene's Landing		1	1.88		1		0.01
3/9/94	1130	104	Greene's Landing	1.99		2.8		0.87		0.34
3/9/94	1130	107	Greene's Landing	2.4		2.9		0.97		0.41
3/9/94	1130	105a	Greene's Landing	2.44	†# <u>.</u>	3.4		0.94		0.43
3/9/94	1130	105b	Greene's Landing	2.39	12	3.1		0.91		0.33
3/9/94	1130	106a	Greene's Landing	2.44	1	34		0.91	ļ	0.43
3/9/94	1130	106b	Greene's Landing	2.34	1	3.2		0.86	<u> </u>	0.32
3/10/94	1.00	108	Greene's Landing	3.46	1	8.2		2.04		0.42
3/10/94	1800	109	Greene's Landing	+	1.79		2	+	0.48	

				Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
3/15/94		110	Greene's Landing			0.11				
3/15/94	L	111	Greene's Landing	2.75		3.8		0.9		0.5
3/15/94		112	Greene's Landing	1.44		4.4		0.44		0.26
3/15/94		113	Greene's Landing	3.97		4.9		1.24		0.58
3/15/9-7		113	Greene's Landing	4.2		4.6		1.34		0.76
3/15/94	1800	115	Greene's Landing		1.5		1.7		0.33	
3/16/94		114	Greene's Landing	3		12.3		1.36		0.46
3/16/94	1100	116	Greene's Landing	0.14		0.03				0.01
3/16/94		117	Greene's Landing			0.43				L
3/16/94		118	Greene's Landing		0.26		0.58		0.02	
3/16/94		119	Greene's Landing	3.26		3.2		0.95		0.51
3/16/94		120	Greene's Landing	2.66		3		0.88		0.49
3/16/94	 	121	Greene's Landing		2.4		2.9		0.86	
3/16/94		122	Greene's Landing		2.59		2.8		0.85	
3/23/94		aa33	French Camp Slough	2.72		9.24		4		2.26
3/23/94	ļ	aa34	French Camp Slough		2.83		3.59		0.81	
3/23/94		aa31	Ulatis Creek	4.23		9.56		3.87		0.46
3/23/94		aa32	Ulatis Creek		2.98		5.55		1.71	
4/12/94	1400	414	Mokelumne River	2.21		4.2		1.49		0.34
4/12/94	1400	475	Mokelumne River		1.29		0.75		0.2	
4/12/94	1200	104CF	Sac R. @ Rio Vista	2.98	1.00	4.02		1.77	0.00	0.26
4/12/94	1200	105CF	Sac R. @ Rio Vista		1.88		1.06		0.37	0.04
4/12/94	900	100CF	Sac. River @ Hood	2.89		4.62		1.34		0.24
4/12/94	900	TUTCF	Sac. River @ Hood	2.94	- 10	3.81	0.04	1.03		0.24
4/12/94	900	102CF	Sac. River @ Hood		2.12		2.30		0.4	
4/12/94	1900	103CF	Sac. River @ Hood	20	2.17	1.07	1.72	0.00	0.34	0.16
4/2/194	1300	497	Middle R. @ Builfrog Ldg.	2.38	2.07	1.97	0.16	0.08	0.20	0.10
4/2/194	1300	498	Middle R. @ Builfrog Ldg.	472	2.07	7.04	0.10	2 27	0.28	0.66
4/2/194	1900	100CF	S.J. River @ Antioch	4.12		7.00		3.27		0.00
4/2/194	900	10/CF	S.J. River @ Antioch	4.83	2.71	0.48	1.46	2.82	0.01	0.95
4/2/194	1900	108CF	S.J. River @ Antioch		2.71		1.40		0.61	
4/2/194	900	109 CI	S.J. River @ Antioch	1 25	2.75	12	1.23	0.6	0.05	0.83
4/2/194	900	11UCF	S.J. River @ Stockton	4.2.5	2.00	15	6.65	0.0	0.2	0.05
4/2/194	1900	111CF	S.J. River @ Stockton		2.99	0.00	0.05		0.2	
4/2/194	1930	400	S.J. River @ Vernalis	<u> </u>		0.06	<u> </u>			
4/2/194	020	401	S.J. River @ Vernalis	2.50		0.24		A A		0.70
4/2/194	020	402	S.J. River @ Vernalis	5.56	117	9.24	0.48	4.4	04	0.72
4/2/194	930	403	S.J. River @ Vernalis		0.69		0.40		0.4	
4/2/194	930	484	Deredice Cut		0.08	<b>\</b>	0.54		0.34	<b> </b>
5/10/04			Duck Slough	12	1.19	26	0.85	187	0.21	22
5/10/94		1227	Duck Slough	12	10		776	10.7	5 30	<u> </u>
5/10/94	020		Creana's Landing	+	4.9		2 20		0.45	
5/10/94	930	0L 201	Greene's Landing	071	1.95	214	2.59	5.95	0.45	1 41
5/10/94		1 <u>g1200</u>	Greene's Landing	0./1	1.05	21.4	230	5.85	0.45	1.41
5/10/94	1200	1541	Mokolumna Diver	2 42	1.95	4.51	2.55	0.04	0.45	032
5/10/94	1200	541/0	Mokelumne River	2.42	· · · · · · · · · · · · · · · · · · ·	2.01		1.06		0.32
5/10/04	1200	1293	Paradise Cut	3 47		4.86		2 13		033
5/10/94	+	445	Paradise Cut	<u></u>	219	4.00	nd		0.06	10.35
5/10/94			Sac P @ Rio Vista	207		5.07	110	2.05	0.00	0.29
5/10/04	+	115cf	Sac R @ Rio Vista	2.31	10	5.07	1 75	2.05	0.52	+
5/10/04	+	112cf	Sac River @ Hood	2.63	+	5 14	1.15	1 52		0.29
5/10/04	+	11201	Sac River @ Hood	2.05	+	3.14	<u> </u>	1.52	<u> </u>	034
5/10/04	+	112cf	Sac River @ Hood	+	1.84	- 5.0	1 33	+	0.55	+ 0.5 -
5/25/04		2210	Old River @ Tracy Rlvd		1.04	+	1 99		+0.37	+
15125194		14410	Toto Kitor @ Tracy Divu.	1	1 1 1 7 7		1	1	1	J

,

				Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
5/25/94		aa9	Old River @ Tracy Blvd.	2.43		7.18		2.33		3.06
5/25/94		aa35	Paradise Cut		1.01		2.07		0.25	
5/25/94		aa8	Paradise Cut		1.81		1.43		0.08	
6/3/94		aa11	Old River @ Tracy Blvd.	3.84		9.26		3.2		- 1.92
6/3/94		aa12	Old River @ Tracy Blvd.		1.74		1.99		0.25	
6/3/94		aal4	Paradise Cut	4.3		7.3		nd		0.64
6/3/94		aa15	Paradise Cut		2.41		2.54		0.08	
7/12/94		aa21	Duck Slough	12.6		32.3		19.6		4.28
7/12/94		aa22	Duck Slough		4.41		7.17		4.78	
7/12/94		aa19	Paradise Cut	4.88		8.95		4.72		0.6
7/12/94		aa20	Paradise Cut		0.2		3.55		0.2	
7/12/94		aa23	Prospect Slough	8.29	1	16.6		10.8		1.24
7/12/94		aa24	Prospect Slough		3.52		6.83	·	3.06	
7/21/94		aa25a	Mokelumne River		1.25		5.65		0.16	
7/21/94		aa25b/QA	Mokelumne River		1.14		5.57		0.11	
7/21/94		aa26a	Mokelumne River	2.01		5.32		0.72		0.3
7/21/94		aa26b/QA	Mokelumne River	1.88		6.34		0.57		0.25
8/9/94		bp 27	Duck Slough	12.5		27.5		22.4		8.98
8/9/94		bp 28	Duck Slough		4.52		6.75		5	
8/9/94		bp 29	Prospect Slough	7.7		12.1		11		1.24
8/9/94	1	bp 30	Prospect Slough	1	4.1		4.03		3.83	
9/2/94		bp1	Duck Slough	13.5		29.6		23.1		8.56
9/2/94		bp1/OA	Duck Slough	14.9		30.7		21.9		7.39
9/2/94		bp2	Duck Slough	+	3.58		4.56		4.08	
9/2/94		bp5	French Camp Slough	6.17		13.3		3.64		1.58
9/2/94		bp6	French Camp Slough		2.94		2.27		0.99	
9/2/94		bp3	Prospect Slough	8.16		13.3		9.58		2.24
9/2/94		bp3/OA	Prospect Slough	8.49		12.2		9.84		2.06
9/2/94		bp4	Prospect Slough		4.22		3.97		3.52	
10/5/94		bp36	5 mile		5.12		70.2		1.01	
10/5/94		bp96	Greene's Landing	4.99				4.16		4.48
10/19/94		aa36	Mokelumne River	2.15		7.29		0.73		0.28
11/4/94		aa27	S.J. River @ Antioch	3.69		7.23		2.31		0.36
11/4/94		aa28	S.J. River @ Antioch		2.19		2.97		0.71	
12/13/94	1245	400	Mokelumne River	3.97		52.8		3.54		0.67
12/13/94	1245	401	Mokelumne River		1.84		4.1		0.72	
12/13/94	1245	402	Mokelumne River		1.89		2		0.77	
12/13/94		122	Ulatis Creek	211		573		131		518
17/13/94		2230	Ulatis Creek	+	3.89		185		0.65	
1/6/95	1500	hp44	Greene's Landing	5 54		10.2		3.71		1.2
1/6/95	1500	bp45	Greene's Landing		299		32		1.28	
1/7/95	1.500	bp46	Greene's Landing	9.02	4.77	179		72		3.48
1/7/95		bp47	Greene's Landing		3.93		3.75		1.98	
1/8/95	1330	bn48	Greene's Landing	106	3.33	197		114		3 91
1/8/95	1330	bp49	Greene's Landing	10.0	4 91		. 5 50		2.94	
1/0/05	1330	bp53	Duck Slough	+	3 30		275		2.24	
1/10/05		bp52	Greene's Landing	28.4	- 3.37	62.9	4.15	20		112
1/10/05		bn53	Greene's Landing	0,4	40	02.9	5 00	47	7	11.6
1/10/05		lbn54	Prospect Slough	124		270		242	<u> </u>	28.4
1/10/05		bn54/04	Prospect Slough	162		328		271		412
1/11/05	1420	10057/QA	Greene's Landing	272	<u> </u>	600		26.9		6.65
1/11/05	1430	bp55	Greene's Landing	-1.5	5.05	07.7	502	20.0	215	-0.05
1/11/05	1620	10050	Prospect Slough	86.0		172		160	5.45	16
1/12/05	1400	10p39	Greene's Londing	17 4		221	·	100		2 40
1/12/95	1400	bp62/0 4	Greene's Landing	20	+	222		19.5		6 20
11112175	11400	TOPOSIQA	La conce a canality	1 40	1	1 22.2	1	1 17	1	1 0.40

				Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
1/12/95	1400	bp63	Greene's Landing		3.35		2.86		3.2	
1/12/95	1400	bp64/QA	Greene's Landing		4.9		4.11		3.04	
1/12/95	1030	bp60	Prospect Slough	34.4		66.3		57.6		7.81
1/13/95	1500	bp65	Greene's Landing	14.2		32.5		21		4.02
1/13/95	1500	bp66	Greene's Landing		3.67		6.32		4.78	
1/13/95	1000	bp67	Prospect Slough	17.9		42.4		32.7		3.65
1/14/95	1300	bp69	Greene's Landing	15.2		71.8		21.3		2.66
1/14/95	1300	bp70	Greene's Landing		3.94		11.2		4.42	
1/14/95	1000	bp68	Prospect Slough	40.3		84		58		13.5
1/15/95	1400	bp71	Greene's Landing	10.7		44.8		12.2		2.55
1/15/95	1400	bp72	Greene's Landing	10.9		48.2		13.3		28.7
1/15/95	1400	bp77	Greene's Landing		3.62		7.93		3.05	
1/15/95	1000	bp74	Prospect Slough	29.8		128		42.3	·····	6.54
1/15/95	1000	bp75	Prospect Slough	28.9		128		42.5		6.15
1/17/95	1400	bp78	Greene's Landing	9.39		18.4		11.6		1.57
1/17/95	1400	bp79	Greene's Landing		3.6		9.4		3.4	
1/17/95	1000	bp80	Prospect Slough	19		78.9		27.1		2.95
1/18/95	1400	bp82	Greene's Landing	10.3		46.9		13.3		7.42
1/18/95	1400	bp83	Greene's Landing		3.68		4.68		3.83	
1/18/95	1100	bp81	Prospect Slough	24.3		103		32.9		4.82
1/20/95	1600	bp86	Greene's Landing	9.68		19.5		12.6		2.05
1/20/95	1600	bp87	Greene's Landing		4.28		4.84		3.43	
1/22/95	1430	bp90	Greene's Landing	9.98		23.3		12		1.75
1/22/95	1430	bp91	Greene's Landing		3.35		4.25		2.5	
1/22/95	1200	bp89	Prospect Slough	13.3		26.3		18.7		2.49
1/22/95	1100	bp88	Skag Slough	11.9		26.3		22.7		2.52
1/23/95	1500	cf500	Greene's Landing	9.43		25.4		8.57		3.24
1/23/95	1500	cf501	Greene's Landing		3.42		4.41		2.52	
1/23/95	1200	cf502	Prospect Slough	14.9		39.3		17.4		3
1/23/95	1000	cf503	Skag Slough	14.6		45.6		24.3		3.9
1/24/95	1600	cf504	Greene's Landing	8.27		11.3		8.44	-	1.55
1/24/95	1600	cf505	Greene's Landing		3.09		22.4		2.68	
1/25/95	1500	cf506	Greene's Landing	7.07		20.9		8.27		2.11
1/25/95	1500	cf507	Greene's Landing		2.88		5.06		4.43	
1/25/95	1000	cf508	Prospect Slough	9.06		28.3		9.56		1.26
1/25/95	1000	cf509	Prospect Slough	·	3.48		5.69		2.51	
1/26/95	1400	cf512	Greene's Landing	9.9		24,4		11		1.83
1/26/95	1500	cf513	Greene's Landing		3.16		4.86		2.07	
1/26/95	1600	cf510	Prospect Slough	15		36.3		21.6		2.53
1/26/95	1600	cf511	Prospect Slough		4.78		8.17		4.08	
1/27/95	1000	cf514	Greene's Landing	8.82		22.3		10.6		2.28
1/27/95	1000	cf515	Greene's Landing		3.27		6.06		4.46	
1/27/95	1530	cf516	Prospect Slough	12.3		31.9		19.2		2.07
1/28/95	1500	cf517	Greene's Landing	8.11		21.7		9.84		2.06
1/28/95	1500	cf518	Greene's Landing		2.77		5.9		2.07	
1/28/95	1200	cf519	Prospect Slough	12.5		32.8	<u> </u>	17.6		2.11
1/28/95	1200	cf520	Prospect Slough	1	4.51		7.87	+	3.69	
1/28/95	11000	cf521	Skag Slough	13		30.3		20.1		2,19
1/29/95	1100	bp92	Greene's Landing	7.34	<u> </u>	17.8		7.75		1.63
1/29/95	1100	bp93	Greene's Landing		2.89		4.34	t	2.13	
1/29/95	+	bn94	Greene's Landing	<u> </u>	3		4.58	<u> </u>	2.17	
1/30/95	1700	cf600	Greene's Landing	679	†	144		7.17		1.04
1/30/95	1700	cf601	Greene's Landing	+	2.87	+	2.47	+	1.75	1
1/31/95	1600	cf602	Greene's Landing	7.02	<u> </u>	14.6		6.77		1.04
1/31/95	1600	ct'603	Greene's Landing	0.02	1	0.599	<u> </u>	0.09		nd

×

				Total	Dis	Total	Dis	Total	Dis	Total
Date	Hour	Station #	Station Name	Cu	Cu	Zn	Zn	Cr	Cr	Pb
1/31/95	1600	cf604	Greene's Landing	7.04	-	16.7		6.27		1.31
1/31/95	1600	cf605/QA	Greene's Landing	7.36		12		6.41		1.99
1/31/95	1600	cf607	Greene's Landing	0.18		1.81		0.2		nd
1/31/95	1600	cf610	Greene's Landing		1.89		3.98		1.59	
1/31/95	1600	cf611	Greene's Landing	2.76		3.98		1.67		0.39
1/31/95	1200	cf606	Prospect Slough	9.73		23.3		11.5		1.45
2/1/95	1300	cf608	Greene's Landing	3.53		12.2		5.02		1.08
2/1/95	1600	cf609	Greene's Landing							
2/2/95	1600	cf612	Greene's Landing	5.9		13.3		4.88		0.86
2/3/95	1400	cf613	Greene's Landing	6.57		14.3		6.03		1.33
2/3/95	1000	cf614	Prospect Slough	8.69		19.9		10		1.12
2/5/95	1500	cf615	Chipps Island	7.96		16.2		7		1.18
2/5/95	1500	cf625	Chipps Island		3.13		4.37		1.7	
2/5/95	1300	cf616	Grizzly Bay	6.58		13.4		5.94		0.95
2/5/95	1300	cf623	Grizzly Bay		3.29		4.84		2.26	
2/5/95	1600	cf617	Martinez	7.15		17.9		6.69		1.01
2/5/95	1000	cf624a	Martinez	3.09		4.21		1.86		0.36
2/5/95	1000	cf624b/OA	Martinez	3.77		3.2		2.05		0.64
2/6/95	1600	cf619	Greene's Landing	6.45		14.5		5.78		1.11
2/6/95	1600	cf622	Greene's Landing		2.37		3.6		1.68	
2/6/95	1400	cf618	Prospect Slough	14.7		29.2		14.3		1.95
2/10/95	1600	cf701a	Greene's Landing	495		10.6		4 4 7		0.63
2/10/95	1600	cf701b/0A	Greene's Landing	54		8 38		3.95		1.04
2/10/95	1600	cf702a	Greene's Landing		2 49	0.50	2 4 1	5.75	1 4 1	
2/10/95	1300	cf702b/0A	Greene's Landing		2.42		1 98		1 37	
2/10/05	1400	cf700	Prospect Slough	734	2.34		1.70	7.65	1.57	0.76
2/10/95	1600	of703	Greene's Londing	5.07	· · · · · ·		····	1.65		0.70
2/14/95	1200	cf704	Brospect Slough	8 22				10.5		4.2
2/14/95	1000	lof705	Flog Slough	2.00				574		4.2
2/14/95	1250	of706	Greene's Londing	73				9 70		1.09
2/17/95	1330	101700	Brospost Slough	572			~- <u>.</u>	0.79	<u> </u>	0.75
2/1//93	1100	101/07	Groopo'a Londing	3.72				0.00		4 49
2/21/95	1400	10090	Greene's Landing	4.99				4.10		4.40
2/21/95	930	101/08	Greene's Landing	3.31				2.02		1.55
2/23/95	1000		Greene's Landing	4.78				3.93		1.30
2/24/95	900		Greene's Landing	4.08				2.9		0.94
2/28/95	2030	CT/12	Greene's Landing	4.14				3.97		1.10
2/28/95	1800	CI/15	Prospect Slough	0.39				14.5		2.95
3/3/95	1530	CT/14	Greene's Landing	4.75				4.44		2.80
3/3/95	1600		Greene's Landing	4.94				5.02		0.90
3/ 1/95	1220		Greene's Landing	3.73		100		4.94		200
3/10/95	1330	bp102	Cottonwood Creek	89.8		189		170		20.9
3/10/95	1330	16p102	Cottonwood Creek	95		151		130		18.9
3/10/95		bp114	East Yolo Bypass	121		333		303		33.3
3/10/95	1115	bp106	Little Cow Cr. @ Dersch Br.	11.6		36.7		8.47		6.65
3/10/95	1115	bp106	Little Cow Cr. @ Dersch Br.	13.2	· .	29.3		6.3		1.14
3/10/95	1240	[bp108	Putah Creek @ Mace Blvd.	76.9	ļ	253		98.4		28
3/10/95	1430	bp105	Sac R. @ Bend Bdg	28.8		68.8		39.6		7.68
3/10/95	2000	bp100	Sac R. @ Colusa Bdg	58.1		129		94.8		12.1
3/10/95	1000	bp97	Sac R. @ Cypress Bdg	8.23		18.7		2.03		0.83
3/10/95	1830	bp98	Sac R. @ Old Ferry	46.8		97.2		75.7		10.2
3/10/95	1550	bp99	Sac R. @ Road a-8	70.4		157		150		15.7
3/10/95	1700	bp107	Sac R. @ Road a-9	56.6		134		99.6		12.9
3/10/95	800	bp103	Sac R. @ Shasta Dam	1.23		4.6		1.44		2.68
3/10/95	1230	bp104	Sac R. @ Balls Ferry Bdg	10.7		29.6		6.5		4.32
3/10/95	2230	bp101	Sacramento Slough	73.2		173		122		17.5

Date	Hour	Station #	Station Name	Total Cu	Dis Cu	Total Zn	Dis Zn	Total Cr	Dis Cr	Total Pb
3/10/95	1	bp112	Skag Slough	5.22		15.3		4.82		4.66
3/10/95		bp113	West Yolo bypass	43		144		90		15.6
3/11/95	1530	bp110	American River @ Sac State	1.15		3.87		1.28		0.44
3/11/95	1200	bp109	Cache Creek 102	130		311		312		30
3/11/95	1200	bp109	Cache Creek 102	151		266		270		31.2
3/11/95	1630	bplll	Feather River @ Hwy 99	4.54		6.29		3.14		0.72
3/11/95	1300	CF 800	Greene's Landing	8.6		19.8		13.8		3.04
3/11/95	1500	CF 801	Mokelumne River	4.31		16.1		2.41		4.66
3/11/95	1500	CF 801	Mokelumne River	4.79		6.27		3.86		3.19
3/11/95	1600	CF 802	S.J. River @ Vernalis	34.1		107		69.1		17.6
3/13/95	1100	CF 803	Sutter Bypass	12		24.8		17.6		4.88
3/13/95		bp117	Sycamore		5.4		18.4		0.39	
3/14/95		bp115	Greene's Landing	6.92		11		8.87		2.86
3/21/95	1800	CF 807	Prospect Slough	10		20.5		13.3		3.45
3/22/95	1700	CF 808	Greene's Landing	3.54		7.92		6.4		2.96
3/22/95	1700	CF 811	Greene's Landing	4.79		6.27		3.86		3.19
3/22/95	1000	CF 809	Mokelumne River	4.26		18.2		2.1		0.89
3/22/95	1000	CF 809	Mokelumne River	4.72		13.3		1.93		1.3
3/22/95	1400	CF 810	S.J. River @ Vernalis	2.89		5.87		2.11		5.43

)

}

Date Hom Station Value FD Cu Cu Nu   1/11/93 GL 22 Greene's Landing 0.04 2.1   1/13/93 GL 23 Greene's Landing 0.02 0.02	0.75		
1/11/93 OL 22 Olcene's Landing 0.04 2.1	0.75	1	
	1 0.75		
1112/25 OL 25 Olectic's Landing 0.00 0.05			
1/14/95 OL 24 Oleene's Landing 0.05 1.55		l	· · · · · · · · · · · · · · · · · · ·
3/23/93 1030 3 Sac R depin 1 0.12 17.4	·		
3/23/93 1030 1 Sac Rsurface 1 0.099 11.0	1.65		
3/23/93 1030 2 Sac R- surface 2 0.21 0.009	1.05	<u> </u>	
3/23/93 1030 4 Sac. R depth 2 0.26 0.02	2.15		
4/13/93 1700 36 Sac. River @ Delta 0.02	4.15	1	
7/7/93 1510 135 Middle R. @ Bullfrog Ldg. 0.007 2.62			
7/7/93 1510 136 Middle R. @ Bullfrog Ldg. 0.1	1.04		· · · · · · · · · · · · · · · · · · ·
7/7/93 1750 149 S.J. River @ Vernalis 0.015 11.7			
7/7/93   1750   150   S.J. River @ Vernalis	2.23		
7/19/93 1038 151 S.J. River @ Antioch 0.03 5.9			
7/19/93 1038 152 S.J. River @ Antioch 0.08 0.013	1.47		
7/19/93 1300 153 Sac. River @ Hood 0.041 4.19			
7/19/93 1300 154 Sac. River @ Hood 0.06 nd	0.7		
7/20/93 F1 Sac R. @ Rio Vista 0.04 4.9			
7/20/93 F2 Sac R. @ Rio Vista 0.1 0.01	1.35		
7/20/93 F3 Sac R. @ Rio Vista 0.08 0.015	1.02	1	
8/3/93 1311 193 Mokelumne River 0.022 0.75	5		
8/3/93 1311 194 Mokelumne River 0.08 0.013	0.31	1	
8/3/93 F-11 Sac R. @ Rio Vista 0.18 0.024	1.71	1	
8/3/93 F-12 Sac R. @ Rio Vista 0.031 2.80		1	
8/3/93 F-10/OC Sac River @ Hood 0.039 4.3			
8/3/93 F-8 Sac River @ Hood 0.05 0.015	0.84		
8/3/93 F-9 Sac River @ Hood 0.037 48			
$\frac{375793}{8/17/93}$ 1200 207 Middle R @ Bullfrog L dg 0.456 38 5			
$\frac{8/17/93}{1200}$ 1200 208 Middle R @ Bullfrog Ldg 0.22	1 22		·
8/17/03 1450 221 S L Piver @ Vernalis 0.011 80			
8/17/02 1450 222 S. J. River @ Vernalis 0.011 0.7	17		
$\frac{0/11/95}{0/14/02}$ 1200 246 Makalumna Divar	1.7		
9/14/95 1200 240 Mokelumine River 0.051 1.2.	<u> </u>		
9/14/93 1200 247 Mokelumie River 0.1	0.20		
9/14/93 1200 248 Mokelumne River 0.1 0.011	0.39	<u> </u>	
9/14/93 13 CF Sac R. @ Rio Vista 0.035 3.24			
9/14/93 14 CF Sac R. @ Rio Vista 0.03 0.017	1.22		
9/14/93   15 CF   Sac R. @ Rio Vista   0.09   0.014	1.1	ļ	
9/14/93 16 CF Sac. River @ Hood 0.038 3.76		ļ	
9/14/93 17 CF Sac. River @ Hood 0.03 0.026	0.96		
10/4/93 2030 269 Sac. River @ Freeport 0.13 0.029	1.62	ļ	
10/4/93 2030 270 Sac. River @ Freeport 0.015 0.54	,		
10/4/93 1100 272 Sac. River @ Freeport 0.044 1.7			
10/4/93 271 0.022 1.5		1	
10/4/93 273 0.036 1.8			
10/14/93 1251 298 Mokelumne River 0.017 0.92	2		
10/14/93 1251 299 Mokelumne River 0.07 0.01	0.31		
10/14/93 18 CF Sac R. @ Rio Vista 0.035 3.6	2		
10/14/93 19 CF Sac R. @ Rio Vista 0.04 0.025	0.85	1	
10/14/93 20 CF Sac. River @ Hood 0.036 2.3		1	
10/14/93 21 CF Sac. River @ Hood nd 0.012	0.63		1.
10/14/93 22 CF Sac. River @ Hood 0.06 0.014	0.67	1	1
10/29/93 1030 312 Middle R. @ Bullfrog Ldg 0.01 1.0	7	+	
10/29/93 1030 313 Middle R @ Bullfrog Ldg 0.005	0.71	+	1
10/29/93 23 CF S I River @ Antioch 0.014 3.2		+	
10/29/93 24 CF/OC SI River @ Antioch 0.017 1.6		+	
10/29/93 25 CF/OC S L River @ Antioch 0.25 0.018	2 73		

Dete		C4-41	Citation N	Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
10/29/93		26 CF	S.J. River @ Stockton		0.009		1.66			
10/29/93		27 CF	S.J. River @ Stockton		0.014		1.71			
10/29/93	l	28 CF	S.J. River @ Stockton	0.23		0.006		1.29		
10/29/93		323	S.J. River @ Vernalis		0.02		4.03			
10/29/93		324	S.J. River @ Vernalis	0.04		0.008		1.13		
11/10/93		29 CF	Greene's Landing		0.04		2.43			
11/10/93		30 CF A	Greene's Landing	0.13		0.15		0.87		
11/10/93		30 CF B	Greene's Landing	0.16		0.14		0.86		
11/11/93		31 CF	Greene's Landing		0.05		1.79			
11/11/93		32 CF	Greene's Landing	0.17		0.1		0.76		
11/11/93		33 CF	Greene's Landing		0.06		1.54			
11/11/93		34 CF	Greene's Landing	0.72		0.35		3.36		
11/11/93		35 CF	Greene's Landing		0.05		2.22			
11/11/93		36CF	Greene's Landing	0.2		0.04		0.9		
11/12/93		37 CF A	Greene's Landing		0.05		2.65			
11/12/93		37 CF B	Greene's Landing		0.05		2.35			
11/12/93		38 CF	Greene's Landing		0.15		2.17			
11/12/93		39 CF	Greene's Landing	0.13		0.04		0.13		
11/29/93		40 CF	S.J. River @ Antioch		0.02		2.97			
11/29/93		41 CF	S.J. River @ Stockton		0.03		1.94			
12/13/93		42 CF	Sac R. @ Rio Vista		0.03		2.88		· · · · · ·	
12/13/93		43 CF	Sac R. @ Rio Vista	0.04	0.02	0.01	2.00	0.87		
12/13/93		44 CF	Sac River @ Hood	- 0.0 .	0.08	0.01	4 52			
12/13/93		44 CF	Sac River @ Hood		0.07		4.81			
12/13/93		45 CF	Sac River @ Hood	0.04	0.07	0.01	1.01	0.87		
1/10/94		GL 21	Greene's Landing	0.01		nd		0.64		
1/10/94		46 CF	S L River @ Antioch	0.01	0.02		3 4 2	0.04		
1/10/94		47 CF	S I River @ Antioch	0.04	0.02	0.04	5.44	0.98		<u> </u>
1/10/94		48 CF	S I River @ Stockton	0.01	0.02	0.01	2 52	0.20		
1/10/94		48 CF	S I River @ Stockton		0.02		23			
1/10/94	<u> </u>	49 CF	S I River @ Stockton		0.02		4.5	2 07		
1/11/94	914	410	Middle R @ Bullfrog I dg		0.02		2 16	2.07		
1/11/94	914	411	Middle R @ Bullfrog I dg		0.01		0.84			
1/11/94	914	412	Middle R. @ Bullfrog Ldg.	0.06	0.01	0.02	0.04	1.52		
1/11/04	014	425	S L River @ Vernalis	0.00		0.02		0.05		
1/11/04	01/	425	S I River @ Vernalis	0.15		0.001		1 03		
1/11/04	01/	420	S.J. River @ Vernalis	0.15	0.01	0.001	2	1.95		┟┥
1/13/04	714	66	Greene's Landing	0.47	0.01	0.03	<u>ت</u>	36		<u>  </u>
1/13/94		65 A	Greene's Landing	0.47	0.00	0.05	672	5.0		<b>├</b> ────┤
1/13/94		65 D	Greene's Landing		0.09		65			
1/15/94	<u>.</u>	25	Greene's Landing		0.09		0.5			┨────┤
1/10/94		23	Greene's Landing		0.01	<u> </u>	1 20			<u>├</u>
1/19/94		24	Greene's Landing	0.06	0.05	0.02	1.59	0.76		
1/25/94		27	Greene's Landing	0.00	0.00	0.02	2.62	0.70		<u> </u>
1/24/94	·	20	Greene's Landing	0.07	0.08	nd	2.05	0 47		
1/24/94		27	Greene's Landing	0.07	0.04	110	2.24	0.07	<u> </u>	+
1/26/04		20	Greene's Landing		0.04		2.24	<u> </u>	ļ	+
1/20/94	+	21	Greene's Landing	0.22	0.05	0.01	5./1	117	<u>}</u>	<u> </u>
1/20/94		122	Greene's Landing	0.23				1.17		<del> </del>
11/2/194		22	Greene's Landing	0.22	1 0.00	0.01	675	1.21	ļ	<u> </u>
11/28/94		32	Greene's Landing	<u> </u>	0.09		0.33	<u> </u>	<b> </b>	<b></b>
1/28/94	<del> </del>	33	Greene's Landing	0.00	0.1	0.00	1.59			<u> </u>
1/28/94	000	130	Greene's Landing	0.26	<u> </u>	0.02		2.3	ļ	<b></b>
1/29/94	1900	40	Greene's Landing	0.22	- 0.02	0.01	6 22	1.89	<b> </b>	
1/30/94	1.000	38	Greene's Landing	0.00	0.06		5.33	1 2 00		l
1/30/94	1000	42	Greene's Landing	0.25		0.01		2.09	1	

÷

				Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
1/31/94		41	Greene's Landing		0.06		4.18			
2/1/94		44	Greene's Landing		0.02		2.56			
2/1/94		48	Greene's Landing	0.14		0.01		1.61		
2/2/94		43	Greene's Landing		0.05		2.97			
2/5/94	1700	55	Greene's Landing	0.39		0.01		1.36		
2/7/94	ļ	50	Greene's Landing		nd		nd			
2/7/94		53	Greene's Landing	0.12		nd		0.87		
2/8/94		51	Greene's Landing		nd		nd			
2/8/94		52	Greene's Landing		0.04		2.2			
2/9/94		54	Greene's Landing		0.09		5.77			
2/10/94		56	Greene's Landing		0.19		19.5		····	
2/10/94	930	58	Greene's Landing	0.46		0.04		3.79		
2/11/94	1000	61	Greene's Landing	0.46	<u> </u>	0.03	<u> </u>	4.01		
2/11/94	1600	62	Greene's Landing		nd		nd			
2/12/94	1200	60	Greene's Landing		0.12		10.8			
2/16/94	700	63	Greene's Landing		0.07		7.09			
2/16/94	1/00	64	Greene's Landing							
2/17/94		67	Greene's Landing		0.06		4			
2/17/94	1000	68	Greene's Landing	0.21		0.02		1.89		
2/18/94	1200	70	Greene's Landing	0.2	0.05	0.02		1.39		
2/19/94	1100	69	Greene's Landing	0.10	0.05	0.00	4.52			
2/19/94	1400	72	Greene's Landing	0.18	0.05	0.02	- <u></u>	1.85		
2/19/94	1400	/ I A	Greene's Landing		0.07		5.91			
2/19/94	1400	/1 B	Greene's Landing	0.10	0.07	0.00	5.55	1.00		
2/20/94	1550	74	Greene's Landing	0.18	- 01	0.03	0.41	1.98	- <del></del>	
2/21/94	1(00	73	Greene's Landing	0.26	0.1	0.00	8.41			
2/21/94	1600	/0	Greene's Landing	0.35	···-	0.02		3.4		·
2/22/94		75	Greene's Landing		0.1		9.4			<b>  </b>
2/22/94	1600	70	Greene's Landing	0.24	0.15	0.01	15.7			
2/22/94	1000	01	Greene's Landing	0.54	0.12	0.01	10	- 2		·
2/23/94	1700	01	Greene's Landing		0.13	0.02	19	2.02		
2/25/94	1700	02	Greene's Landing		0.02	0.03	1 62	2.02		·
2/24/94	1700	0.0	Greene's Landing	0.52	0.05	0.02	4.02	4.62		
2/24/94	1700	04	Greene's Landing	0.52	0.07	0.05	7.4	4.02		
2/25/94	1900	85	Greene's Landing	0.2	0.07	0.07	1.4	2.21		
2/25/94	1800	00	Greene's Landing	0.5	- 01	0.02	0.25	2.31		
2/2/194		0/	Greene's Landing		0.1		9.25			<b>├</b> ────
2/20/94	1200	00	Greene's Landing	0.25	0.00	0.07	3.09	1.02		
2/28/94	1200	90	Greene's Landing	0.25	0.05	0.03	2 72	1.92		
3/1/94		91	Greene's Landing	0.16	0.05	0.02	3.73	1.50	·	
3/1/94		95	Greene's Landing	0.10	0.04	0.02	4.07	1.39		
2/4/94	1200	95	Greene's Landing	0.1	0.00	0.02	4.07	112		
2/0/04	11200	90	Greene's Landing	0.1		0.03		1.15		
2/0/04	1130	100	Greene's Landing				<u> </u>			
3/0/04	1120	102	Greene's Landing		0.01	<u> </u>				┥────┤
3/9/94	1130	102	Greene's Landing		0.01	·				
3/0/04	1120	105	Greene's Londing		0.24		1 12			
3/0/04	1130	107	Greene's Landing		0.50		0.06			<u> </u>
3/0/04	1120	105	Greene's Landing		0.41		1 1			╆┉┉┉┥
3/0/04	1130	1056	Greene's Landing	<b> </b>	0.42		0.00	<u> </u>		┥────┥
2/0/04	1130	1050	Greene's Landing	<u> </u>	0.43		1.05			<u> </u>
3/9/94	1130	1066	Greene's Landing	<b> </b>	0.42		1.05	<u> </u>		┥────┤
2/10/04	1130	1000	Greene's Landing		0.42		2.40		ļ	┥────┤
2/10/94	1000	108	Greene's Landing	0.00	0.04	0.01	3.49	1.25		┿╾────┤
10/94	11800	109	Joreene's Landing	1 0.08	1	0.01	1	1.20		1

				Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
3/15/94		110	Greene's Landing		0.01					
3/15/94		111	Greene's Landing		0.52		1.03			
3/15/94		112	Greene's Landing		0.26		0.52			
3/15/94		113	Greene's Landing		0.68		1.6			
3/15/94		113	Greene's Landing		0.74		1.54			
3/15/94	1800	115	Greene's Landing	0.06		0.02		0.94		
3/16/94		114	Greene's Landing		0.06		2.4			
3/16/94	1100	116	Greene's Landing							
3/16/94	L	117	Greene's Landing		0.01		0.32			
3/16/94		118	Greene's Landing	0.02		0.01				
3/16/94		119	Greene's Landing		0.54		0.99			
3/16/94		120	Greene's Landing		0.54		1.03			
3/16/94		121	Greene's Landing	0.53		0.55		0.92		
3/16/94	L	122	Greene's Landing	0.36		0.41		0.84		
3/23/94		aa33	French Camp Slough		0.044		3.33		1.49	
3/23/94	L	aa34	French Camp Slough	0.41		0.011		1.29		1.33
3/23/94	L	aa31	Ulatis Creek		0.027		5.69		1.78	
3/23/94		aa32	Ulatis Creek	0.07		0.018		3.65		1.62
4/12/94	1400	474	Mokelumne River		0.013		1.73			
4/12/94	1400	475	Mokelumne River	0.1		0.005		0.55		
4/12/94	1200	104CF	Sac R. @ Rio Vista		0.024		2.99			
4/12/94	1200	105CF	Sac R. @ Rio Vista	0.08		0.019		1.21		
4/12/94	900	100CF	Sac. River @ Hood		0.027		2:02			
4/12/94	900	101CF	Sac. River @ Hood		0.033		1.64			
4/12/94	900	102CF	Sac. River @ Hood	0.07		0.015		0.92		
4/12/94	900	103CF	Sac. River @ Hood	0.075		0.015		0.75		
4/27/94	1300	497	Middle R. @ Bullfrog Ldg.		0.01	0.00	1.98			
4/27/94	1300	498	Middle R. @ Bullfrog Ldg.	0.06	0.001	0.007		1.41		
4/2/194	900	106CF	S.J. River @ Antioch		0.031		5.15			
4/27/94	900	107CF	S.J. River @ Antioch		0.029	0.010	4.15			
4/27/94	900	108CF	S.J. River @ Antioch	0.12		0.013		1.98		
4/27/94	900	109 cf	S.J. River @ Antioch	0.13	0.001	0.016		1.43		
4/27/94	900	110CF	S.J. River @ Stockton	0.16	0.021	0.01	2.17	1.04		
4/2/194	900		S.J. River @ Stockton	0.16		0.01		1.84		
4/2/194	930	480	S.J. River @ Vernalis							
4/2/194	930	481	S.J. River @ Vernalis		0.014		5.53			
4/2/194	930	482	S.J. River @ Vernalis	0.07	0.014		5.55	0.07		
4/2/194	930	483	S.J. River @ Vernalis	0.07		0.000		0.97		
4/2//94	930	484	S.J. River @ Vernalis	0.09	1	0.002		0.88		1.24
4/30/94		aal	Paradise Cut	nd	0.000	0.008	24.1	2.07	2.06	1.24
5/10/94		aab	Duck Slough	1.05	0.069	0.010	24.1	0.50	2.00	1.00
5/10/94	020	aa /	Duck Slough	1.05	ļ	0.012		8.52		1.09
5/10/94	930	GL 201	Greene's Landing	0.1	0.104	0.032	0.07	1.23	0.02	0.71
5/10/94		<u>g1200</u>	Greene's Landing		0.104	0.020	9.27	1.00	0.83	0.71
5/10/94	1200	g1201	Greene's Landing	0.1	0.012	0.032	1.40	1.23	1.27	0.71
5/10/94	1200	541	Mokelumne River		0.012		1.48	<u> </u>	1.27	
5/10/94	1200	1541/QA	Depending Cut	+	0.000		1.19		0.11	
5/10/94		1223	Paradise Cut		10.018	0.000	3.79	1 02	0.11	0.24
5/10/94		11405	Paradise Cut	na	0.020	0.008	2 45	1.83	22	0.24
5/10/94		114CI	Sac K. W KIO VISIA	0.00	0.028	0.015	5.45	1 42	2.2	10
5/10/94			Sac K. @ KIO Vista	0.09	0.027	10.015	1 2 24	1.43	1 72	1.9
5/10/94		1112ct	Sac. River @ Hood		0.036		1.02		1.72	
5/10/94		112ct/QA	Sac. Kiver @ Hood	1000	0.026	0.016	1.83	1	1.01	1 94
5/10/94			Old Diver @ Hood	+0.09	+	0.010	<b> </b>	2 01	<b> </b>	1.84
13/23/94		aaru	JOIG KIVER @ IFACY BIVD.	0.12		0.014	1	1 3.01	1	1.1

)

}

				Dis	Total	Dis	Total	<b>D:</b> N	Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
5/25/94		aa9	Old River @ Tracy Blvd.		0.02		2.82		0.98	
5/25/94		aa35	Paradise Cut	0.04		0.009		2.12		1.4
5/25/94		aa8	Paradise Cut	nd		nd		2.29		1.34
6/3/94		aall	Old River @ Tracy Blvd.		0.023		3.28		0.81	
6/3/94		aa12	Old River @ Tracy Blvd.	0.05		0.008		1		1.58
6/3/94		aal4	Paradise Cut		0.019		4.75		1.74	
6/3/94		aal5	Paradise Cut	0.07		0.008		2.38		1
7/12/94		aa21	Duck Slough		0.081		28.8		1.58	
7/12/94		aa22	Duck Slough	0.88		0.035		6.85		1.32
7/12/94		aa19	Paradise Cut		0.025		8.59		3.15	
7/12/94		aa20	Paradise Cut	0.05		0.007		2.16		2.27
7/12/94		aa23	Prospect Slough		0.035		15.3		1.06	
7/12/94		aa24	Prospect Slough	0.4		0.017		5.36		1
7/21/94		aa25a	Mokelumne River	0.08		0.017		0.44		0.6
7/21/94		aa25b/QA	Mokelumne River	0.1		0.008		0.47		0.45
7/21/94		aa26a	Mokelumne River		0.024	-	0.68		0.5	
7/21/94		aa26b/QA	Mokelumne River		0.022		0.63		0.63	
8/9/94		bp 27	Duck Slough		0.066		31.4		2.4	
8/9/94		bp 28	Duck Slough	1.38		0.011		8	·	2.05
8/9/94		bp 29	Prospect Slough		0.03		15.7		1.67	
8/9/94		bp 30	Prospect Slough	0.41		0.023		7.04		1.93
9/2/94		bpl	Duck Slough		0.071		35.8		2.21	
9/2/94		bp1/OA	Duck Slough		0.064		34.3		3.98	
9/2/94		bp2	Duck Slough	1.08		0.021		5.16		2.17
9/2/94		bp5	French Camp Slough		0.038		2.15		2.71	
9/2/94		hn6	French Camp Slough	0.37		0.014		0.99		24
9/2/94		bn3	Prospect Slough		0.036		183		21	
9/2/94		bp3/OA	Prospect Slough		0.031		18.5		3.24	<u>├</u>
9/2/94		hn4	Prospect Slough	0.73	0.001	0.021		6.12		2.04
10/5/94		bp36	5 mile	3.87	· ·	0.081		5 29		3.03
10/5/94		bp96	Greene's Landing	5.07	0.048	0.001	7.04	5.25		
10/19/94		2236	Mokelumpe River		0.019		0.83			
11/4/94		aa27	S L River @ Antioch		0.012		4 2		0.41	
11/4/94		aa28	S I River @ Antioch	0.09	4.012	0.014	1.4	212	011	013
12/13/94	1245	400	Mokelumne River	0.07	0.02	0.011	3 34			
12/13/94	1245	401	Mokelumne River	0.18	0.02	0.01	2.24	1 34		<u>+</u>
12/13/94	1245	402	Mokelumne River	0.18	<u> </u>	0.01		1 33		
12/13/04		102	Illatis Creek	0.10	0.126	0.01	16.2	1.55	1 22	<u>  </u>
12/13/04		2230	Ulatis Creek	0.2	0.120	0.043	10.2	3.45	1.22	1 30
1/6/05	1500	bp44	Greene's Londing	0.2	0.063	0.045	6.02	5.45	1.52	
1/6/05	1500	bp45	Greene's Landing	0.45	0.005	0.028	0.02	210	1.54	1 41
1/7/05	1500	bp45	Greene's Londing	0.45	0110	0.020	10.5	2.17	1 2	
1/7/05		bp40	Greene's Landing	0.79	0.118	0.028	10.5	2.07	1.2	{
1/8/05	1330	bp48	Greene's Landing	0.78	0.108	0.028	16	2.91	03	
1/8/05	1330	bp40	Greene's Landing	0.77	0.106	0.038	10	4.51	0.5	0.45
1/0/05	1220	bp53	Duck Slough	0.77	<u> </u>	0.038		6 25		0.40
1/9/93		bp55	Creana's Landing	0.57	0 474	0.021	216	0.35		
1/10/95		10µ02	Greene's Landing	0.01	0.4/4	0.020	3.10	4 21	<u> </u>	1 27
1/10/95		lbp54	Prospect Slough	0.81	0.569	0.039	601	4.51	06	1.2/
1/10/95		10p54	Prospect Slough		0.508		507	<u> </u>	0.0	<del>  </del>
1/11/05	1420	10004/QA	Greene's Londing		0.52		201		2.07	<u> </u>
1/11/95	1450	10055	Greene's Landing	0.00	0.329	0.045	28.3	2.07	2.97	1.
1/11/95	1430	10050	Dreene's Landing	0.99	0 220	0.045	417	3.91	1.46	0.88
1/11/95	1030		Crospect Slougn		0.229		41/	<u> </u>	1.40	<b> </b>
1/12/95	1400		Greene's Landing	ļ	0.184	ļ	2/.1	<u> </u>	1.32	<u> </u>
1/12/95	1400	opo2/QA	Greene's Landing		0.19		23.7	1	i	

....

				Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
1/12/95	1400	bp63	Greene's Landing	0.53		0.034		8.5		1.19
1/12/95	1400	bp64/QA	Greene's Landing	0.99		0.04		4.85		
1/12/95	1030	bp60	Prospect Slough		0.181		103		1.5	
1/13/95	1500	bp65	Greene's Landing		0.166		23.6		1.09	
1/13/95	1500	bp66	Greene's Landing	0.65		0.035		4.78		1.14
1/13/95	1000	bp67	Prospect Slough		0.163		38		1.63	
1/14/95	1300	bp69	Greene's Landing		0.167		26.9		2.45	
1/14/95	1300	bp70	Greene's Landing	0.8		0.018		6.02		0.84
1/14/95	1000	bp68	Prospect Slough		0.224		79.2		1.2	
1/15/95	1400	bp71	Greene's Landing		0.114		13.8		0.9	
1/15/95	1400	bp72	Greene's Landing		0.124		14.9	10.1	0.31	0.01
1/15/95	1400	bp77	Greene's Landing	0.48	0.000	0.031		19.1		0.91
1/15/95	1000	bp74	Prospect Slough		0.203		53.7		2.48	
1/15/95	1000	bp75	Prospect Slough		0.197		62.8		2.27	
1/1//95	1400	bp /8	Greene's Landing		0.087		24.8		0.72	
1/17/95	1400	bp79	Greene's Landing	0.49	0.00-	0.002	~~~~~	26		1.12
1/1//95	1000	bp80	Prospect Slough		0.087		36.6		3.32	
1/18/95	1400	bp82	Greene's Landing	0.50	0.09		23.7		0.61	
1/18/95	1400	bp83	Greene's Landing	0.52	0.13	0.033		6.21		1.06
1/18/95	1100	bp81	Prospect Slough		0.17		45.1		4.41	
1/20/95	1600	6086	Greene's Landing	0.54	0.089	0.11	18	6 22	1.2	1.07
1/20/95	1600	bp87	Greene's Landing	0.54	0.005	0.11	160	0.33	1.4	1.07
1/22/95	1430	bp90	Greene's Landing	0.4	0.095	0.025	10.2	275	1.4	1.26
1/22/95	1430	10091	Greene's Landing	0.4	0.000	0.025		3.75	1.07	1.30
1/22/95	11200	10089	Prospect Slough		0.092		27.3		1.07	
1/22/95	1100	0088	Skag Slougn		0.008		33.9		2.54	
1/23/95	1500	C1500	Greene's Landing	0.42	0.087	0.004	.13.1		1.22	1.00
1/23/95	1500	CT501	Greene's Landing	0.43	0.104	0.024	- 20.0	4.45	1.10	1.09
1/23/95	1200	c1502	Prospect Slough		0.104		20.0		1.18	
1/23/95	1600	c1505	Grappo's Londing		0.008		41.9		3.08	
1/24/95	1600	c1304	Greene's Landing	0.26	0.064	0.027	11.0	3.46	1.07	1 25
1/24/93	1500	c1505	Greene's Landing	0.30	0.08	0.027	12	3.40	1.52	1.25
1/25/95	1500	cf507	Greene's Landing	0.4	0.06	0.025	12	4.07	1.52	1 14
1/25/95	1000	cf508	Brospect Slough	0.4	0.075	0.025	16.7	4.07	1 8 1	1.14
1/25/95	1000	cf509	Prospect Slough	0.38	0.075	0.023	10.7	4 39	1.01	1 43
1/26/95	1400	cf512	Greene's Landing	0.50	0111	0.025	174	4.57	1 59	1.45
1/26/95	1500	cf513	Greene's Landing	035	0.111	0.032		4 34	1.57	1 25
1/26/95	1600	cf510	Prospect Slough	- <u></u>	0.107	0.002	36.6		nd	
1/26/95	1600	cf511	Prospect Slough	0.57	0.107	0.064	50.0	7.28		1 51
1/27/95	1000	cf514	Greene's Landing	0.57	0.08	0.004	162		1.08	
1/27/95	1000	cf515	Greene's Landing	0.46	0.00	0.033	10.2	4.06		1.18
1/27/95	1530	cf516	Prospect Slough		0.096	0.000	28.3	1.00	1.48	
1/28/05	1500	cf517	Greene's Landing	<u> </u>	0.082		157	<u> </u>	1.24	
1/28/95	1500	cf518	Greene's Landing	041	10.002	0.073		4.34		1
1/28/95	1200	cf519	Prospect Slough	0.41	0.111	0.01.	293		0.99	<u> </u>
1/28/95	1200	cf520	Prospect Slough	0.57	+	0.064		6.75		1.45
1/28/05	1000	cf521	Skag Slough		0.12	0.001	37.2	+	1.48	1
1/20/95	1100	hn92	Greene's Landing		0.105		10.8		1.13	<u>†</u>
1/29/95	$\frac{1100}{1100}$	bn93	Greene's Landing	0.34	+	0.034		3.95	+	1.22
1/20/05	+	bp94	Greene's Landing	0.41	†	0.039		3.72		0.94
1/30/95	1700	cf600	Greene's Landing		0.054	10.000	11.3	+	1.18	+
1/30/95	1700	cf601	Greene's Landing	0.24	+	0.021	<u>-</u>	3.11	1	
1/31/95	1600	cf602	Greene's Landing		0.104		10.6	+	1.54	<u>+</u>
1/31/95	1600	cf603	Greene's Landing		nd	+	0.18	1	nd	+
1		- t	1							

		0		Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cd	Cd	Ni	Dis Ni	As	Dis As
1/31/95	1600	cf604	Greene's Landing		0.057		10.6		1.54	
1/31/95	1600	cf605/QA	Greene's Landing	<u></u>	0.05		10			
1/31/95	1600	cf607	Greene's Landing		0.008		0.91			
1/31/95	1600	cf610	Greene's Landing	0.37		0.02		2.99		
1/31/95	1600	cf611	Greene's Landing		0.027		2.71			
1/31/95	1200	cf606	Prospect Slough		0.065		14.8			
2/1/95	1300	cf608	Greene's Landing		0.07		6.61			
2/1/95	1600	cf609	Greene's Landing							
2/2/95	1600	cf612	Greene's Landing		0.042		5.92			
2/3/95	1400	cf613	Greene's Landing		0.062		8.45			
2/3/95	1000	cf614	Prospect Slough		0.07		13.5			
2/5/95	1500	cf615	Chipps Island		0.065		11.5			
2/5/95	1500	cf625	Chipps Island	0.43		0.039		2.67		
2/5/95	1300	cf616	Grizzly Bay		0.045		9.64			
2/5/95	1300	cf623	Grizzly Bay	0.31		0.024		3.27		
2/5/95	1600	cf617	Martinez		0.056		10.9			
2/5/95	1000	cf624a	Martinez		0.035		3.12			
2/5/95	1000	cf624b/QA	Martinez		0.03		3.88			
2/6/95	1600	cf619	Greene's Landing		0.051		8.63			
2/6/95	1600	cf622	Greene's Landing	0.25		0.032		2.44		
2/6/95	1400	cf618	Prospect Slough		0.082		21.3			
2/10/95	1600	cf701a	Greene's Landing		0.057		7.1			
2/10/95	1600	cf701b/QA	Greene's Landing		0.04		6.33			
2/10/95	1600	cf702a	Greene's Landing	0.18	· ·	0.012		2.23		
2/10/95	1300	cf702b/QA	Greene's Landing	0.29		0.02		2.15		
2/10/95	1400	cf700	Prospect Slough		0.068		11.4			
2/14/95	1600	cf703	Greene's Landing		0.056		6.71.			
2/14/95	1300	cf704	Prospect Slough		0.084		15.8			
2/14/95	1000	cf705	Skag Slough		0.026		11.1			
2/17/95	1350	cf706	Greene's Landing		0.11		12.3		<u></u>	
2/17/95	1100	cf707	Prospect Slough		0.036		13.8			
2/21/95	1400	bn96	Greene's Landing		0.048		7 04			I
2/21/95	930	cf708	Greene's Landing		0.069		7 49			
2/23/95	1600	bn97	Greene's Landing		0.053		631			
2/24/95	900	ct711	Greene's Landing		0.057		4 59			
2/28/95	2030	ct712	Greene's Landing	~	0.037		5.85			
2/28/95	800	cf713	Prospect Slough		0.045		283			
3/3/95	1530	cf714	Greene's Landing		0.066		5 79			
3/5/95	1600	cf715	Greene's Landing		0.076		6 56			
3/7/95		cf716	Greene's Landing		0.052		6.18			
3/10/95	1330	bp102	Cottonwood Creek		0.032		233			
3/10/95	1330	bp102	Cottonwood Creek		0.29		189		· · · · · ·	
3/10/95	1000	bp114	East Yolo Bypass		0438		600			
3/10/95	1115	bp106	Little Cow Cr. @ Dersch Br.		0.123		7 98			
3/10/95	1115	bp100	Little Cow Cr. @ Dersch Br.		0.125		62			
3/10/95	1240	bp108	Putah Creek @ Mace Blvd		0.105		88 1		·	<b>├</b> ────
3/10/05	1430	bp105	Sac R @ Bend Rdg		0.7/		52			
3/10/05	2000	bp100	Sac R. @ Coluce Bdg		0 400		266			
3/10/95	1000	bp07	Sac R. @ Cupress Pdg		0.409		200			<u>├</u>
3/10/95	1000	bp08	Sac P. @ Old Formu		0.11		251			
3/10/95	1630	bo00	Sac R. @ Dood a <sup>0</sup>		0.290		102			<u>├</u>
2/10/93	1330	bp107	Sac R. W ROad a-8		0.3/1		492	<b></b>		<b>  </b>
2/10/95	1700	bp107	Sac K. W KOad a-9		0.311		112			<b> </b>
2/10/95	1000	bp103	Sac R. @ Shasta Dam		0.020		2.30	<b> </b>		<b>  </b>
2/10/95	1230	bp104	Sac K. @ Dalls Ferry Bug		0.154		1.41			<b>├</b>
[3/10/93	2230	ισρισι	Sacramento Stougn		0.433	1	120	I		1 1

				Dis	Total	Dis	Total		Total	
Date	Hour	Station #	Station Name	Pb	Cđ	Cđ	NI	DIS NI	AS	Dis As
3/10/95		bp112	Skag Slough		0.057		14.1			
3/10/95		bp113	West Yolo bypass		0.311		165			
3/11/95	1530	bp110	American River @ Sac State		0.017		2.17			
3/11/95	1200	bp109	Cache Creek 102		0.495		651			
3/11/95	1200	bp109	Cache Creek 102		0.311		653			
3/11/95	1630	bp111	Feather River @ Hwy 99	<u>.</u>	0.026		4.06			
3/11/95	1300	CF 800	Greene's Landing		0.16		13.2			
3/11/95	1500	CF 801	Mokelumne River		0.066		2.61			
3/11/95	1500	CF 801	Mokelumne River		0.033		5.72			
3/11/95	1600	CF 802	S.J. River @ Vernalis		0.169		128			
3/13/95	1100	CF 803	Sutter Bypass		0.068		20.4			
3/13/95		bp117	Sycamore					2.86		
3/14/95		bp115	Greene's Landing		0.056		11.1			
3/21/95	1800	CF 807	Prospect Slough		0.072		19.3			
3/22/95	1700	CF 808	Greene's Landing		0.029		5.76			
3/22/95	1700	CF 811	Greene's Landing		0.033		5.72			
3/22/95	1000	CF 809	Mokelumne River		0.095		2.47			
3/22/95	1000	CF 809	Mokelumne River		0.084		1.72			
3/22/95	1400	CF 810	S.J. River @ Vernalis		0.024		3.97			

<u></u>	<u> </u>	[	l	Total	Dis	Total	1	
Date	Hour	Station #	Station Name	Ag	Ag	Fe	Dis Fe	Hardness
1/1/1/93	har an	GL 22	Greene's Landing		0.013			
1/13/93	<u> </u>	GL 23	Greene's Landing	1	0.008			
1/14/93		GL 24	Greene's Landing	0.014	-			(
3/23/93	1030	3	Sac R depth 1			4600		
3/23/93	1030	1 .	Sac R surface 1			3600		
3/23/93	1030	2	Sac R surface 2				410	
3/23/93	1030	4	Sac. R depth 2				600	
4/13/93	1700	36	Sac. River @ Delta					
717193	1510	135	Middle R. @ Bullfrog Ldg.	0.013	· ·			74
7/7/93	1510	136	Middle R. @ Bullfrog Ldg.		0.005			74
7/7/93	1750	149	S.J. River @ Vernalis	0.015				146
7/7/93	1750	150	S.J. River @ Vernalis					146
7/19/93	1038	151	S.J. River @ Antioch	0.01				78
7/19/93	1038	152	S.J. River @ Antioch					78
7/19/93	1300	153	Sac. River @ Hood	0.009	1			48
7/19/93	1300	154	Sac. River @ Hood		0.003			48
7/20/93		Fl	Sac R. @ Rio Vista	0.009		i		-44
7/20/93		F2	Sac R. @ Rio Vista		nd			44
7/20/93		F3	Sac R. @ Rio Vista		< 0.002			44
8/3/93	1311	193	Mokelumne River	0.003	·, · · · · · ·			36
8/3/93	1311	194	Mokelumne River		nd			36
8/3/93		F-11	Sac R. @ Rio Vista		0.006			64
8/3/93		F-12	Sac R. @ Rio Vista	0.007	i			64
8/3/93		F-10/QC	Sac. River @ Hood					66
8/3/93		F-8	Sac. River @ Hood		0.004			66
8/3/93		F-9	Sac. River @ Hood	0.011				66
8/17/93	1200	207	Middle R. @ Bullfrog Ldg.	\	·			48
8/17/93	1200	208	Middle R. @ Bullfrog Ldg.					48
8/17/93	1450	221	S.J. River @ Vernalis					136
8/17/93	1450	222	S.J. River @ Vernalis					136
9/14/93	1200	246	Mokelumne River					32
9/14/93	1200	247	Mokelumne River					32
9/14/93	1200	248	Mokelumne River	0.000				32
9/14/93		13 CF	Sac R. @ Rio Vista	0.006				64
9/14/93	· .	14 CF	Sac R. @ Rio Vista	na	0.000			64
9/14/93	}	IS CF	Sac R. @ RIO VISta	<b> </b>	<0.002	·		04
9/14/93		10 CF	Sac. River @ Hood	ļ				64
9/14/93	2010	17 CF	Sac. River @ Hood	l				04
10/4/93	2030	209	Sac. River @ Freeport	<u> </u>				80
10/4/93	1100	270	Sac. River @ Freeport	<u> </u>				80
10/4/95	1100	271	Sac. River & Fleepolt			····		0
10/4/93		273				·	· · · · · · · · · · · · · · · · · · ·	
10/4/95	1251	275	Makelumne Diver	ļ				
10/14/93	1251	290	Mokelumne River		·			24
10/14/93	11231	1299	See P. @ Pie Viete	0.000			}	56
10/14/93			Sac R. @ Rio Vista	0.008	nd	ļ		56
10/14/93		19 CF	Sac N. W. KIU VISLA	<u> </u>	<u>nu</u>			- 30
10/14/93		20 CF	Sac, River @ Hood					40
10/14/93		121 CF	Sac. River @ Hood	<u> </u>	<u> </u>			40
10/14/93	1020	2205	Middle D. @ Dullfrog I do	<u> </u>				40
10/20/02	0201	212	Middle P. @ Duillfrog Ldg.	┟────			<u> </u>	62
10/20/02	1030	22 CE	S I Diver @ Antioch	<u> </u>		760		676
10/29/93	<u> </u>	23 CF	S.J. River @ Antioch			75		, 020
10/20/02		124 CFIQU	S.J. River @ Antioch	<u> </u>		13	810	626
10/22/23		LAJ UF/QU	DI ANU CANUCI	L	L	ŀ	010	020

Date	Hour	Station #	Station Name	Total Ag	Dis Ag	Total Fe	Dis Fe	Hardness
10/29/93		26 CF	S.J. River @ Stockton					74
10/29/93		27 CF	S.J. River @ Stockton					74
10/29/93		28 CF	S.J. River @ Stockton			· · ·		74
10/29/93		323	S.J. River @ Vernalis					128
10/29/93		324	S.J. River @ Vernalis					128
11/10/93		29 CF	Greene's Landing					60
11/10/93		30 CF A	Greene's Landing					60
11/10/93		30 CF B	Greene's Landing					60
11/11/93		31 CF	Greene's Landing					60
11/11/93		32 CF	Greene's Landing					60
11/11/93		33 CF	Greene's Landing					60
11/11/93	<u> </u>	34 CF	Greene's Landing					60
11/11/93	<u> </u>	35 CF	Greene's Landing					60
11/11/93		36CF	Greene's Landing					60
11/12/03		37 CF A	Greene's Landing					60
11/12/03	}	137 CF R	Greene's Landing					60
11/12/93		137 CF B	Greene's Landing					60
11/12/93		30 CF	Greene's Landing			<u>-</u>		
11/12/93		139 CF	Greene's Landing	0.014			ļ	60
11/29/93		40 CF	S.J. River @ Antioch	0.014				010
11/29/93		41 CF	S.J. River @ Stockton	0.012				188
12/13/93		42 CF	Sac R. @ Rio Vista	0.01	0.000			70
12/13/93	<u> </u>	43 CF	Sac R. @ Rio Vista	0.010	0.002			/6
12/13/93		44 CF	Sac. River @ Hood	0.012				54
12/13/93	ļ	44 CF	Sac. River @ Hood	ļ			ļ	54
12/13/93	ļ	45 CF	Sac. River @ Hood		0.002			54
1/10/94	L	GL 21	Greene's Landing	0.002				64
1/10/94	ļ	46 CF	S.J. River @ Antioch	0.004			 	262
1/10/94		47 CF	S.J. River @ Antioch				L	262
1/10/94		48 CF	S.J. River @ Stockton	L			ļ	204
1/10/94	<u> </u>	48 CF	S.J. River @ Stockton					204
1/10/94		49 CF	S.J. River @ Stockton					204
1/11/94	914	410	Middle R. @ Bullfrog Ldg.					88
1/11/94	914	411	Middle R. @ Bullfrog Ldg.					88
1/11/94	914	412	Middle R. @ Bullfrog Ldg.					88
1/11/94	914	425	S.J. River @ Vernalis					156
1/11/94	914	426	S.J. River @ Vernalis					156
1/11/94	914	427	S.J. River @ Vernalis					156
1/13/94		66	Greene's Landing					66
1/13/94	1	65 A	Greene's Landing					66
1/13/94	1	65 B	Greene's Landing	1	1	1		66
1/18/94	1	25	Greene's Landing	1		<u> </u>	1	60
1/19/94	1	24	Greene's Landing	1		[	1	60
1/23/94	1	27	Greene's Landing	1	1		1	80
1/24/94	+	26	Greene's Landing	+	<u> </u>	t	1	88
1/24/94	1	29	Greene's Landing	+	1	t	1	88
1/25/94	+	28	Greene's Landing	<u> </u>	1	1	+	76
1/26/94	+	30	Greene's Landing	1		<u>+</u>		88
1/26/04	+	131	Greene's Landing		<u> </u>			88
1/27/04		133	Greene's Landing	+	1	1	+	88
1/29/04	+	32	Greene's Landing	+	+	1		64
1/20/94		35	Greene's Landing	+	+	·	+	64
1/20/94		26	Greene's Landing	+	+	+	<u> </u>	64
1/28/94	-000	30	Greene's Landing	+		+	+	66
1/29/94	900	140	Greene's Landing		+			66
1/30/94	1000	38	Greene's Landing			+		4 60
1/30/94	1000	42	Greene's Landing					00

.

	1	T	·	Total	Dis	Total	<u></u>	<u> </u>
Date	Hour	Station #	Station Name	Ag	Ag	Fe	Dis Fe	Hardness
1/31/94		41	Greene's Landing	8				66
2/1/94		44	Greene's Landing		<del></del>			72
2/1/94		48	Greene's Landing					72
2/2/94	1	43	Greene's Landing		,	}		72
2/5/94	1700	55	Greene's Landing		<u></u>			60
2/7/94		50	Greene's Landing					68
2/7/94		53	Greene's Landing		-			68
2/8/94	1.	51	Greene's Landing		-			72
2/8/94	+	52	Greene's Landing		- <del>: +</del>			72
2/9/94		54	Greene's Landing			·		80
2/10/94		56	Greene's Landing		7.6.			54
2/10/94	930	58	Greene's Landing			<u> </u>		54
2/11/94	1000	61	Greene's Landing			1		60
2/11/94	1600	62	Greene's Landing					60
2/12/94	- <u> </u>	60	Greene's Landing			·		64
2/16/94	700	63	Greene's Landing					
2/16/94	700	64	Greene's Landing		,	· · · · ·		
2/17/94	1	67	Greene's Landing			·		80
2/17/94		68	Greene's Landing					80
2/18/94	1200	70	Greene's Landing					80
2/19/94		69	Greene's Landing		,			86
2/19/94	1400	72	Greene's Landing					86
2/19/94	1400	71 A	Greene's Landing					86
2/19/94	1400_	71 B	Greene's Landing					86
2/20/94	1550	74	Greene's Landing					72
2/21/94		73	Greene's Landing					66
2/21/94	1600	76	Greene's Landing		<u>.</u>			· 66
2/22/94	<u> </u>	75	Greene's Landing					56
2/22/94	1	77	Greene's Landing		·	L		56
2/22/94	1600	79	Greene's Landing					56
2/23/94		81	Greene's Landing					58
2/23/94	1700	82	Greene's Landing		·			58
2/24/94		83	Greene's Landing					62
2/24/94	1700	84	Greene's Landing					62
2/25/94		85	Greene's Landing					66
2/25/94	1800	86	Greene's Landing			ļ	[]	66
2/27/94		87	Greene's Landing					80
2/28/94	1.000	89	Greene's Landing			<u> </u>		82
2/28/94	1200	90	Greene's Landing			Į		82
3/1/94		191	Greene's Landing					84
3/1/94		93	Greene's Landing			<u> </u>	 	84
3/4/94	1200	195	Greene's Landing					88
3/4/94	1200	196	Greene's Landing					88
3/9/94	1130	100	Greene's Landing		~ <u></u>			
3/9/94	1130	101	Greene's Landing		·			
2/0/04	1120	102	Greene's Landing			<u> </u>		
3/0/04	1120	103	Greene's Landing			<del> </del>		
3/0/0/	1120	104	Greene's Landing			┢────		
3/9/94	1120	107	Greene's Landing					
2/0/04	1120	1056	Greene's Landing		·,	·		
3/0/04	1130	1050	Greene's Landing				<u> </u>	
2/0/04	1120	1004	Greene's Landing			<u> </u>		
3/10/04	1130	1000	Greene's Landing					76
3/10/94	1800	100	Greene's Landing		,			76
5/10/94	1000	107	Loreene s Landing			1	1	/0

•

â

....

÷

I

,
			[	Total	Dis	Total		le la
Date	Hour	Station #	Station Name	Ap	Ag	Fe	Dis Fe	Hardness
3/15/94		110	Greene's Landing					72
3/15/94		111	Greene's Landing			<u> </u>		72
3/15/94		112	Greene's Landing					72
3/15/94		113	Greene's Landing					72
3/15/04 -		113	Greene's Landing					72
3/15/04	1800	115	Greene's Landing			┼────		72
3/16/04	1000	114	Greene's Londing					72
3/16/04	1100	114	Greene's Landing					72
2/16/04	1100	117	Greene's Landing					
2/16/04		117	Greene's Landing			<i>↓ ·</i>		
2/16/94		110	Greene's Landing			<u> </u>		72
2/16/94		119	Greene's Landing					
2/16/94		120	Greene's Landing					72
2/16/94		121	Greene's Landing					72
3/10/94	<u> </u>	122	Greene's Landing			ļ		12
3/23/94		aa33	French Camp Slough	<u> </u>				44
3/23/94		aa34	French Camp Slough			ļ		44
3/23/94		22	Ulatis Creek			ļ		304
3/23/94	1 400	122	Ulatis Creek			<u> </u>		304
4/12/94	1400	4/4	Mokelumne River			ļ		32
4/12/94	1400	475	Mokelumne River					32
4/12/94	1200	104CF	Sac R. @ Rio Vista	L				76
4/12/94	1200	105CF	Sac R. @ Rio Vista			Į,		76
4/12/94	900	100CF	Sac. River @ Hood					70
4/12/94	900	101CF	Sac. River @ Hood					70
4/12/94	900	102CF	Sac. River @ Hood					70
4/12/94	900	103CF	Sac. River @ Hood					70
4/27/94	1300	497	Middle R. @ Bullfrog Ldg.					124
4/27/94	1300	498	Middle R. @ Bullfrog Ldg.					124
4/27/94	900	106CF	S.J. River @ Antioch					154
4/27/94	900	107CF	S.J. River @ Antioch					154
4/27/94	900	108CF	S.J. River @ Antioch					154
4/27/94	900	109 cf	S.J. River @ Antioch					154
4/27/94	900	110CF	S.J. River @ Stockton					172
4/27/94	900	111CF	S.J. River @ Stockton					172
4/27/94	930	480	S.J. River @ Vernalis					84
4/27/94	930	481	S.J. River @ Vernalis					84
4/27/94	930	482	S.J. River @ Vernalis			1	[	84
4/27/94	930	483	S.J. River @ Vernalis			1		84
4/27/94	930	484	S.J. River @ Vernalis	1				84
4/30/94		aal	Paradise Cut	1			1	432
5/10/94	1	laa6	Duck Slough					98
5/10/94		aa7	Duck Slough	1				98
5/10/94	930	GL 201	Greene's Landing	1			1	66
5/10/94		g1200	Greene's Landing	1	1	1	1	66
5/10/94		gl201	Greene's Landing	1	1	1		66
5/10/94	1200	541	Mokelumne River	1		1	1	30
5/10/94	1200	541/QA	Mokelumne River	1	<u> </u>	1	1	30
5/10/94	1200	1223	Paradise Cut	1			1	396
5/10/94	+	122	Paradise Cut		1			396
5/10/04	+	114cf	Sac R @ Rio Vista	+	1		1	62
5/10/04	+	115cf	Sac R @ Rio Vista	+	1	-	1	62
5/10/04	+	112cf	Sac River @ Hood	+	+			54
5/10/04	+	112cf/0A	Sac River @ Hood		+		+	54
5/10/94	+	11201/QA	Sac Diver @ Hood	+	+	-+	+	54
5/25/04			Old Piver @ Trocy Plud	+	<u> </u>		+	152
15125194	1	10010	JUIU KIVEI @ HACY DIVU.		1	1	I	1.7

Date	Hour	Station #	Station Name	Total Ag	Dis Ag	Total Fe	Dis Fe	Hardness
5/25/94		aa9	Old River @ Tracy Blvd.					152
5/25/94		aa35	Paradise Cut					398
5/25/94		aa8	Paradise Cut					398
6/3/94		aall	Old River @ Tracy Blvd.					238
6/3/94		aal2	Old River @ Tracy Blvd.					238
6/3/94		aa14	Paradise Cut					384
6/3/94		aal5	Paradise Cut					384
7/12/94		aa21	Duck Slough		<del>,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,</del>			72
7/12/94		aa22	Duck Slough					72
7/12/94		aa19	Paradise Cut					400
7/12/94		aa20	Paradise Cut	-				400
7/12/94		aa23	Prospect Slough		÷			84.3
7/12/94		aa24	Prospect Slough					84.3
7/21/94		aa25a	Mokelumne River		0.008			
7/21/94		aa25b/OA	Mokelumne River					
7/21/94		aa26a	Mokelumne River	0.008				
7/21/94		aa26b/OA	Mokelumne River					
8/9/94		bp 27	Duck Slough	•	<u> </u>			68
8/9/94		bp 28	Duck Slough				<u> </u>	68
8/9/94		bp 29	Prospect Slough		* * • • • •			72
8/9/94		bp 30	Prospect Slough					72
9/2/94		bp 30	Duck Slough		4			70
9/2/94			Duck Slough					70
0/2/04		hp?	Duck Slough					70
0/2/04		bp2	French Camp Slough					82
0/2/04		bp6	French Camp Slough					82
0/2/04		hp3	Prospect Slough	-				86
0/2/04		bp3/QA	Prospect Slough					86
0/2/94		bp3/QA	Prospect Slough					86
10/5/04		bp36	5 mile		÷		┼────	80
10/5/94		10030	Groopola Londing					<u> </u>
10/3/94	ļ	10036	Mokolumno Divor					
11/4/04		10027	S L Diver @ Antioch	0.012	· · ·		·	
11/4/94		10029	S.J. River @ Antioch	0.012	0.004			
12/12/04	1245	400	S.J. River @ Andoch		0.004			
12/13/94	1245	400	Mokelumne River				<u> </u>	
12/13/94	1245	401	Makalumna Divar					
12/13/94	1245	402	Wokelumne River				<u> </u>	
12/13/94		aa29	Ulatis Creek					
12/13/94	1.500	aa30	Ulatis Creek		<del></del>			02
1/0/95	1500	10p44	Greene's Landing		<i></i>			92
1/6/95	1500	16p45	Greene's Landing				<u> </u>	92
1///95	<u> </u>	10040	Greene's Landing					00
1/1/95	1000	10047	Greene's Landing				<del> </del>	00
1/8/95	1330	0048	Greene's Landing		. <u></u>		<u> </u>	00
1/8/95	1330	10049	Greene's Landing				<u> </u>	00
1/9/95		0023	Duck Slough				<u> </u>	234
1/10/95	ļ	bp52	Greene's Landing		<u> </u>		l	52
1/10/95	ļ	0053	Greene's Landing				<u> </u>	52
1/10/95		10p54	Prospect Slough					82
1/10/95		bp54/QA	Prospect Slough		<u> </u>		ļ	82
1/11/95	1430	bp55 .	Greene's Landing				ļ	44
1/11/95	1430	bp56	Greene's Landing		ļ		<u> </u>	44
1/11/95	1630	bp59	Prospect Slough				<u> </u>	88
1/12/95	1400	bp61	Greene's Landing					42
1/12/95	1400	bp62/QA	Greene's Landing	1			1	42

	T			Total	Dis	Total	1	
Date	Hour	Station #	Station Name	Ag	Ag	Fe	Dis Fe	Hardness
1/12/95	1400	bp63	Greene's Landing					42
1/12/95	1400	bp64/QA	Greene's Landing	1				42
1/12/95	1030	bp60	Prospect Slough					62
1/13/95	1500	bp65	Greene's Landing					58
1/13/95	1500	bp66	Greene's Landing					58
1/13/95	1000	bp67	Prospect Slough					58
1/14/95	1300	bp69	Greene's Landing	1	···			40
1/14/95	1300	bp70	Greene's Landing					40
1/14/95	1000	bp68	Prospect Slough					82
1/15/95	1400	bp71	Greene's Landing	<u> </u>				44
1/15/95	1400	bp72	Greene's Landing					44
1/15/95	1400	bp77	Greene's Landing				1	44
1/15/95	1000	bp74	Prospect Slough					60
1/15/95	1000	bp75	Prospect Slough					60
1/17/95	1400	bp78	Greene's Landing					44
1/17/95	1400	bp79	Greene's Landing					44
1/17/95	1000	bp80	Prospect Slough		····			48
1/18/95	1400	bp82	Greene's Landing					44
1/18/95	1400	bp83	Greene's Landing					44
1/18/95	1100	bp81	Prospect Slough					
1/20/95	1600	bp86	Greene's Landing					48
1/20/95	1600	bp87	Greene's Landing				·	48
1/22/95	1430	bp90	Greene's Landing					54
1/22/95	1430	bp91	Greene's Landing					54
1/22/95	1200	bp89	Prospect Slough	1				64
1/22/95	1100	bp88	Skag Slough					116
1/23/95	1500	cf500	Greene's Landing					50
1/23/95	1500	cf501	Greene's Landing			-		50
1/23/95	1200	cf502	Prospect Slough					60
1/23/95	1000	cf503	Skag Slough					124
1/24/95	1600	cf504	Greene's Landing					56
1/24/95	1600	cf505	Greene's Landing					56
1/25/95	1500	cf506	Greene's Landing					54
1/25/95	1500	cf507	Greene's Landing					54
1/25/95	1000	cf508	Prospect Slough					64
1/25/95	1000	cf509	Prospect Slough					64
1/26/95	1400	cf512	Greene's Landing					50
1/26/95	1500	cf513	Greene's Landing					50
1/26/95	1600	cf510	Prospect Slough					56
1/26/95	1600	cf511	Prospect Slough			1	1	56
1/27/95	1000	cf514	Greene's Landing					48
1/27/95	1000	cf515	Greene's Landing					48
1/27/95	1530	cf516	Prospect Slough					60
1/28/95	1500	cf517	Greene's Landing				L	48
1/28/95	1500	cf518	Greene's Landing					48
1/28/95	1200	cf519	Prospect Slough					60
1/28/95	1200	cf520	Prospect Slough					60
1/28/95	1000	cf521	Skag Slough					104
1/29/95	1100	bp92	Greene's Landing					44
1/29/95	1100	bp93	Greene's Landing					44
1/29/95		bp94	Greene's Landing					44
1/30/95	1700	cf600	Greene's Landing					48
1/30/95	1700	cf601	Greene's Landing					48
1/31/95	1600	cf602	Greene's Landing					48
1/31/95	1600	cf603	Greene's Landing		1			48

[	1			Total	Dis	Total		]
Date	Hour	Station #	Station Name	Ag	Ag	Fe	Dis Fe	Hardness
1/31/95	1600	cf604	Greene's Landing			1		48
1/31/95	1600	cf605/QA	Greene's Landing					48
1/31/95	1600	cf607	Greene's Landing					48
1/31/95	1600	cf610	Greene's Landing					48
1/31/95	1600	cf611	Greene's Landing					48
1/31/95	1200	cf606	Prospect Slough					68
2/1/95	1300	cf608	Greene's Landing					50
2/1/95	1600	cf609	Greene's Landing					50
2/2/95	1600	cf612	Greene's Landing			<u> </u>		50
2/3/95	1400	cf613	Greene's Landing					48
2/3/95	1000	cf614	Prospect Slough					68
2/5/95	1500	cf615	Chipps Island		•	1		62
2/5/95	1500	cf625	Chipps Island		- i i i i i i i i i i i i i i i i i i i	1		62
2/5/95	1300	cf616	Grizzly Bay					66
2/5/95	1300	cf623	Grizzly Bay					66
2/5/95	1600	cf617	Martinez					72
2/5/95	1000	cf624a	Martinez				· · ·	72
2/5/95	1000	cf624b/OA	Martinez			1		72
2/6/95	1600	cf619	Greene's Landing					46
2/6/95	1600	cf622	Greene's Landing		1	1		46
2/6/95	1400	cf618	Prospect Slough					46
2/10/95	1600	cf701a	Greene's Landing					52
2/10/95	1600	cf701b/OA	Greene's Landing		·			52
2/10/95	1600	cf702a	Greene's Landing					52
2/10/95	1300	cf702b/OA	Greene's Landing			1		52
2/10/95	1400	cf700	Prospect Slough					66
2/14/95	1600	cf703	Greene's Landing					62
2/14/95	1300	cf704	Prospect Slough			<u> </u>		80
2/14/95	1000	cf705	Skag Slough			†		192
2/17/95	1350	cf706	Greene's Landing				•	56
2/17/95	1100	cf707	Prospect Slough					148
2/21/95	1400	bp96	Greene's Landing			1		56
2/21/95	930	cf708	Greene's Landing					56
2/23/95	1600	bp97	Greene's Landing			1		64
2/24/95	900	cf711	Greene's Landing				1	64
2/28/95	2030	cf712	Greene's Landing			1		64
2/28/95	800	cf713	Prospect Slough			1		244
3/3/95	1530	cf714	Greene's Landing		1			58
3/5/95	1600	cf715	Greene's Landing		,			50
3/7/95		cf716	Greene's Landing			1		46
3/10/95	1330	bp102	Cottonwood Creek					60
3/10/95	1330	bp102	Cottonwood Creek			1		60
3/10/95	1	bp114	East Yolo Bypass		-	1		148
3/10/95	1115	bp106	Little Cow Cr. @ Dersch Br.			1		36
3/10/95	1115	bp106	Little Cow Cr. @ Dersch Br.				<u> </u>	36
3/10/95	1240	bp108	Putah Creek @ Mace Blvd.		,	1		112
3/10/95	1430	bp105	Sac R. @ Bend Bdg			1	<u> </u>	36
3/10/95	2000	bp100	Sac R. @ Colusa Bdg		1	1	1	48
3/10/95	1000	bp97	Sac R. @ Cypress Bdg	İ		1	1	40
3/10/95	1830	bp98	Sac R. @ Old Ferry		1	1	1	48
3/10/95	1550	bp99	Sac R. @ Road a-8				1	54
3/10/95	1700	bp107	Sac R. @ Road a-9		·	1		136
3/10/95	800	bp103	Sac R. @ Shasta Dam			1	1	46
3/10/95	1230	bp104	Sac R. @ Balls Ferry Bdg	1		1		38
3/10/95	2230	bp101	Sacramento Slough		1		1	108

÷,

				Total	Dis	Total		· ]
Date	Hour	Station #	Station Name	Ag	Ag	Fe	Dis Fe	Hardness
3/10/95		bp112	Skag Slough					220
3/10/95		bp113	West Yolo bypass					62
3/11/95	1530	bp110	American River @ Sac State					28
3/11/95	1200	bp109	Cache Creek 102					128
3/11/95	1200	bp109	Cache Creek 102					128
3/11/95	1630	bp111	Feather River @ Hwy 99					28
3/11/95	1300	CF 800	Greene's Landing		1			30
3/11/95	1500	CF 801	Mokelumne River					22
3/11/95	1500	CF 801	Mokelumne River					22
3/11/95	1600	CF 802	S.J. River @ Vernalis					114
3/13/95	1100	CF 803	Sutter Bypass					46
3/13/95		bp117	Sycamore					128
3/14/95		bp115	Greene's Landing					30
3/21/95	1800	CF 807	Prospect Slough					56
3/22/95	1700	CF 808	Greene's Landing					56
3/22/95	1700	CF 811	Greene's Landing					56
3/22/95	1000	CF 809	Mokelumne River					36
3/22/95	1000	CF 809	Mokelumne River					36
3/22/95	1400	CF 810	S.J. River @ Vernalis					84

# **APPENDIX C:**

Quality Assurance/Quality Control Methods and Results

#### METHODS

#### METAL ANALYSES

*Field* The field portion of the QA program consisted of collecting blanks and field duplicates. Field blanks were collected to insure that samples were not contaminated by any aspect of the collecting procedure. A five gallon carboy of ultra pure water was brought to a field site. Water was pumped from the carboy following the same procedures which were used when a routine field sample was collected.

On 64 occasions duplicate water samples were collected from randomly selected sites to characterize field variability and the reproducibility of the measurements performed by the Trace Metal Laboratory and the Mussel Watch Laboratory. Field duplicates consisted of collecting two samples with a ten minute lapse between samples. This field duplicate collection method does not allow precision to be evaluated rigorously, for any observed variability could be a combination of inter-laboratory variability and real changes in the system during the ten minute lag in sample collection. Therefore, the measured variability could be considered a maximum with the true inter-laboratory precision being lower.

The laboratory component of the OA program was focused toward characterizing Laboratorv contamination of sampling equipment and assessing measures of precision and accuracy. Laboratory blanks were collected to insure that the sampling equipment was not contaminated. This procedure consisted of pumping ultra pure water (18 megaohm deionized) water through the peristaltic tubing and filter apparatus into an analysis bottle. Precision is a measure of the reproducibility of a test method when it is repeated under controlled conditions. As described in the QA/QC documents (Goetzl et al., 1994; 1995), precision was evaluated by two methods: (1) inter-laboratory analyses of field duplicates (see sample collection description above) between the Trace Metal Laboratory and Mussel Watch Laboratory, and 2) an intra-laboratory repeated analysis of the standard reference materials (SRMs) by the Mussel Watch Laboratory. The agreement between the amount of a component measured by the test method and the amount actually present is a measure of accuracy of the test method. To measure accuracy, one SRM was run for approximately every 25 samples analyzed. The standard reference materials used were Riverine Water SLRS-2 and SLRS-3 (for 1993-94 samples and 1994-95 samples, respectively) from the National Research Council of Canada. Certified values for the SRMs used in this study can be found in the QA/QC reports (Goetzl et al., 1994, 1995).

#### TOXICITY ASSESSMENT

Standard procedures were followed in all aspects of the toxicity assessment. Monthly reference toxicant tests, consisting of five to six known concentrations of NaCl in laboratory control water, were conducted for each species. Chronic  $LC_{50}$  and  $EC_{50}$  concentrations were calculated to ascertain changes in animal sensitivity throughout the time period of the study. A complete description of quality assurance measures can be found in the Delta Monitoring Quality Assurance Project Plans (Connor *et al.*, 1995; Nielsen *et al.*, 1995).

#### RESULTS

#### METAL ANALYSES

**Field** On nine occasions field blanks were collected; twice for dissolved metals and seven times for total recoverable metals (Table C-1). Contamination was negligible with no metals detected above 1  $\mu$ g/l. This finding is consistent with the minimal contamination reported when the technique was applied to quantify metal concentrations in Central Valley reservoir releases (Goetzl and Stephenson, 1993). Field duplicates were collected on 64 occasions with a resulting average difference between the two laboratories of 16% (Table C-2; Goetzl *et al.*, 1995). Differences between the two laboratories were found to be random, with neither laboratory consistently higher or lower than the other. This value incorporates both a measure of the ten minute lag in sample collection of the duplicates and inter-laboratory variability. Values not detected by either laboratory or very close to the detection limit (e.g., cutoff point at 5x the detection limit) were not included.

*Laboratory* Laboratory blanks were collected on 11 occasions with 65% of the individual metals data quantified as below the detection limits from the method (Table C-3). Contamination was negligible with only one metal detected above 1  $\mu$ g/l on one occasion when metals were detected in the laboratory blanks. These findings were consistent with those in Goetzl and Stephenson (1993), indicating the sampling gear was relatively free of metal contamination. Laboratory blanks were also collected to determine if filtration of samples prior to conducting toxicity tests resulted in contamination (Table C-4). Of three laboratory blanks tested for filtration effects, there was no consistent pattern of removal or contamination for the seven metals. Although 0.45  $\mu$ m filtration of field samples may have removed colloids and possibly resulted in sorption of metals on the membrane. Since filtration effects were not assessed for field samples, the concentrations reported for metals in this study are conservative estimates and may somewhat underestimate the actual values.

Intra-laboratory precision was assessed between five and 11 times depending on the metal. The average difference between the certified and mean detected values ranged from 2 to 20% (Goetzl *et al.*, 1994; 1995). All values were between the 99% confidence limits for the SRMs (Goetzl *et al.*, 1994; 1995). Inter-laboratory precision, which incorporated a measure of inter-laboratory and field variability, was shown to be within an average of 14% and 18% of each other for the 1993-94 and 1994-95 samples, respectively (Table C-2; Goetzl *et al.*, 1995). Values that were not detected by either lab or values that were very close to the detection limit (i.e., cutoff point at 5x the detection limit) were not included in the precision calculation. In addition, the calculation did not include values that differed between labs by a large amount (e.g., outliers). Those values were highlighted in the reports (Goetzl and Stephenson, 1993; Goetzl *et al.*, 1995). Single-laboratory precision was analyzed using the SRM SLRS-2 and SRM SLRS-3 for the 1993-94 and 1994-95 samples, respectively. All of the values for the elements were within the 99% confidence limits of the SRMs.

Approximately one standard reference material (SRMs) was analyzed for every 25 samples to address the accuracy of the evapoconcentration method. The SRM metal values were all greater than ten times the detectable limits with the exception of silver (1993-94 and 1994-95 samples) and lead (1994-95 samples) (Goetzl et al., 1994; 1995). All of the 1993-94 SRMs were within the warning limits, which are  $\pm 15\%$  greater than the 95% SRM confidence limits. All of the 1994-95 SRMs were within the warning limits, with the exception of lead. The SRM for lead used with the 1994-95 samples was considerably lower than the lead SRM used with the 1993-94 samples. The 1994-1995 value was very close to the detection limit, making it difficult to analyze. All values (in both years) were within the warning and control limits ( $\pm 20\%$  greater than the 95% SRM confidence limits) with the exception of lead. All but one lead SRM value in the 1994-95 document was between the warning and control limits. These results indicate, with few exceptions, a high level of accuracy and precision were associated with the evapoconcentration method utilized in this program. Analysis of SRMs can be used to describe the expected accuracy of field samples if the certified SRM values are similar to mean ambient metal concentrations. The certified SRM values in this study ranged from 31% to 99% lower than the mean metal concentrations measured in field samples collected from 1993 to 1995. Obtaining similar certified SRM values and mean field concentrations was inhibited by the nature of sampling which occurred over a wide spatial and temporal scale. This resulted in considerable spatial and temporal differences in metal concentrations over the course of the study.

## TOXICITY ASSESSMENT

Between test variability was assessed for this study with reference toxicant tests. USEPA (1994) recommends reference toxicant testing to ascertain whether changes in animal sensitivity occurred. Of particular interest are the detection of outlier values exceeding the upper or lower 95 percent confidence limits of the long term mean or of general trends in changing animal sensitivity. During the 1993-1994 phase of testing, neither were noted in the control charts of any of the test species (Deanovic et al., 1996). One outlier occurred in the LC<sub>50</sub> chart for Pimephales mortality. In this particular case, the fathead minnow was less sensitive to NaCl. All quality control measurements showed acceptable characteristics suggesting toxicity test data were reliable. One outlying value each occurred in the Ceriodaphnia reproduction and survival test, the Selenastrum and Pimephales growth assays, and the fish mortality data during the 1994-1995 phase of testing (Deanovic et al., 1998). The USEPA (1994) suggests one outlying value may be expected to occur by chance when 20 or more events are compared. Twenty-one to twenty-four data points were presented in the control charts, therefore, quality control measurements were acceptable and indicated the bioassay data were reliable. A more complete description of the Quality Assurance information for the toxicity studies can be found in the toxicity reports (Deanovic et al., 1996; 1998).

Sample ID	Cu	Zn	Cr	Pb	Cd	Ni	As
dissolved (cf630)	<.04	0.04	<.05	<.01	0.011	0.25	
total recoverable (cf805)	<.04	<.01	· <.05	<.01	<.002	<.02	
total recoverable (cf603)	0.02	0.599	0.09	<.01	<.002	0.18	<.1
total recoverable (cf804)	<.04	0.01	<.05	<.02	<.002	<.02	
total recoverable (51)	0.16	0.16	<.05	<.01	<.002	<.02	
total recoverable (110)	<.04	0.11	<.05	<.01	0.01	<.02	
total recoverable (117)	<.04	0.43	<.05	<.01	0.01	0.32	
total recoverable (481)	<.04	0.24	<.05	<.01	<.002	<.02	
dissolved (cf105)	0.07	0.09	0.08	<.01	0.003	0.1	

Table C-1. Summary of field blanks (18 megaohm deionized water) run through field sampling equipment at various sampling sites. Values are expressed as  $\mu g/l$ . Sample sites are in parentheses.

ю

	Metal Species						
Station Code	Cu	Zn	Cr	Pb	Cd	Ni	As
1994							
F9/F10 (TR)	10	2	1	13	5	11	
F2/F3 (D)	7	47	18	20	33	24	
246/247 (TR)	12	15	29	10	16	10	
270/271 (TR)	25		_		32		
272/273 (TR)	13	36	9	18	18	5	
14CF/15CF (D)	6	37	5	67	17	10	
21CF/22CF (D)	1	26	35	13	14	6	
26CF/27CF (TR)	77	11	28	13	36	3	
44CFA/44CFB (TR)	1	1	15	2	13	6	
48CFA/48CFB (TR)	7	5	30.		0	9	
401/402 (D)	3		6	0	0	1	
410/411 (TR)		23	57	73	50		
425/426 (D)	30	23					
30CFA/30CFB (D)	11	13	11	19	7	1	
37CFA/37CFB (TR)	24	18	7	12	0	11	
25/25B (D)	1		28	67	0	29	
30/30B (TR)	2	30	12	1	0	8	
33/34 (D)	1		19	15	50	12	
38/39 (TR)	14		2	7	14	11	
44/45 (TR)	8		4	24	33	2	
46A/46B (TR)	14	20	10	0	5	7	
47A/47B (TR)	9	33	11	9	<u> </u>	13	

Table C-2. Percent Difference Between Duplicate Analyses for Total Recoverable and Dissolved Concentrations of Seven Metals in Field Samples Collected from the Sacramento/San Joaquin Delta Estuary. (D) = dissolved; (TR) = total recoverable.

.

.

٠

.

Table C-2 (cont.). Percent Difference Between Duplicate Analyses for Total Recoverable and Dissolved Concentrations of Seven Metals in Field Samples Collected from the Sacramento/San Joaquin Delta Estuary. (D) = dissolved; (TR) = total recoverable.

	Metal Species						
Station Code	Cu	Zn	Cr	Pb	Cd	Ni	As
48/49 (D)	6	27	0 .	36	50	12	
56/57 (TR)	9	27	5	41	10	1	
58/59 (D)	3	4	10	28	20	1	
65A/65B (TR)	3	41	1	7	0	3	
71A/71B (TR)	24	21	8	16	0	6	
77/78 (TR)	4	15	6	15	13	2	
79/80 (D)	2	22	3	6	50	5	х. х
91/92 (TR)	6	34	8	18	0	1	
93/94 (D).	29	18	7	20	0	13	
105A/105B (TR)	2	9	. 3	23	23	2	
106A/106B (TR)	4	6	6	26	0	10	
111A/111B (TR)	4	24	7	20	12	5	
113/113QC (TR)	6	6	8	24	8	4	
121/121QC (D)	7.	4	1	5	26	9	
GL131/GL132 (D)	8	28	-3	0	16	13	· · · · · · · · · · · · · · · · · · ·
483/484 (D)	42	11	15	22		9	
100CF/101CF (TR)	2	18	23	0	18	- 19	
102CF/103CF (D)	2	27	15	7	0	19	
CF106/CF107 (TR)	3	8	14	29	6	19	
CF108/CF109 (D)	2	16	22	8	19	28	
bpl (TR)	9	. 4	5	14	10	4	45
bp3/bp32 (TR)	5	8	3	8	14	11	35

Table C-2 (cont.). Percent Difference Between Duplicate Analyses for Total Recoverable and Dissolved Concentrations of Seven Metals in Field Samples Collected from the Sacramento/San Joaquin Delta Estuary. (D) = dissolved; (TR) = total recoverable.

.

.

.

.

	Metal Species						
Station Code	Cu	Zn	Cr	Pb	Cd	Ni	As
bp10/bp11 (TR)	11	14	12	13	18	21	20
bp15/bp16 (TR)	15	20	14	21	9	13	15
112cf (TR)	11	26	11	15	28	22	6
541 (TR)	15	36	11	16	50	20	14
380/381 (TR)	1	27	1	4	23	18	20
aa25a/aa25b (D)	9	2	31	0	53	6	25
aa26a/aa26b (TR)	7	16	21	17	8	7	21
bp51 (TR)	20	0	1	22	8	18	
bp54 (TR)	24	18	11	31	9	2	
bp61/bp62 (TR)	13	1	2	41	3	5	
bp63/bp64 (D)	32	31	5	47	15	43	
cf604/cf605 (TR)	4	28	2	34	12	6	
cf624a/cf624b (D)	18	24	9	44	14	20	
cf701A/cf701B (TR)	18	21	12	40	30	12	
cf702A/cf702B (D)	2	12	3	38	40	4	
bp102 (TR)	5	20	24	10	30	19	
bp106 (TR)	12	20	26	7	15	22	
bp109 (TR)	14	15	14	4	37	0	
cf801 (TR)	10	61	38	32	50	54	
cf809 (TR)	10	27	7	32	12	30	
Mean % Difference	10	19	13	20	17	11	31
SD	9	13	12	17	16	11	11

Mean % Difference WY94 = 14%; Mean % Difference WY95 = 18%; Overall Mean% Difference WY94 & WY95 = 16%

Sample ID	Cu	Zn	Cr	Pb	Cd	Ni	As
total recoverable (bp7)	<.04	0.05	<.05	<.01	<.002	0.02	<.03
total recoverable (bp32)	0.13	0.22	<.05	0.03	0.002	0.04	<.03
total recoverable (bp26)	<.04	0.04	<.05	. <.01	<.002	<.02	0.12
dissolved (cf628)	<.04	0.39	<.05	<.01	0.009	0.24	
total recoverable (50)	< 04	0.14	<.05	<.01	<.002	<.02	
total recoverable (cf607)	0.18	1.81	0.2	<.01	0.008	0.91	
total recoverable (62)	<.04	<.01	<.05	<.01	<.002	<.02	
total recoverable (cf804)	<.04	<.01	<.05	<.01	<.002	<.02	
total recoverable (116)	0.14	0.03	<.05	0.01	<.002	<.02	
total recoverable (480)	<.04	0.08	<.05	<.01	<.002	<.02	· · ·
dissolved (cf104)	<.04	<.01	0.08	<.01	0.005	<.02	

Table C-3. Summary of laboratory blanks (18 megaohm deionized water) run through field sampling equipment. Values are expressed as  $\mu g/l$ . Sample numbers are in parentheses.

#	Cu	Zn	Cr	Pb	Cd	Ni	As
1 Unfiltered	0.09	0.2	nd	nd	nd	nd	0.18
1 Filtered	0.06	0.36	nd	nd	nd	nd	0.18
2 Unfiltered	nd	0.08	nd	nd	0.01	0.11	0.14
2 Filtered	0.02	0.28	nd	0.06	nd	nd	nd
3 Unfiltered	nd	0.84	nd	nd	0.009	nd	
3 Filtered	nd	0.26	nd	nd	nd	nd	

Table C-4. Summary of toxicity study blanks (deionized water) analyzed to assess potential addition of metals via filtration. Filtered treatments were passed through a through 0.45  $\mu$ m filter. Values are expressed as  $\mu$ g/l. nd = non-detect

.

.

é

## APPENDIX D

J

.

•

-

## **Metals Source Pilot Study**

#### INTRODUCTION

÷

Water samples were collected for a one-time pilot study during a major storm event in March 1995 to assess the relative metal load contribution from sources upstream of the Delta, primarily in the Sacramento River Watershed. The study was designed to assess metal loads, therefore only total recoverable concentrations were quantified. No toxicity samples were collected and the lack of dissolved metals analyses prohibited an assessment of water quality objective exceedances. Although the objective of the pilot study was to track sources of metals during a high flow event, the data could not be used to quantify the load contribution from mines in the area of Lake Shasta and Keswick Reservoir because discharges from the reservoirs were maintained at low levels to minimize downstream flooding. This resulted in samples downstream of the reservoirs which were negligibly affected by runoff from this mining region.

#### MATERIALS AND METHODS

Sample collection and metal analyses followed the ultra-clean methods described in the main body of this report. Load calculations were point estimates because samples were only collected once. Loads were calculated by simply multiplying the total recoverable metal concentrations by flow measurements.

#### Sample Locations

A special study was undertaken from 10 March to 13 March 1995 to track sources of metals into the Delta. Samples were collected from 22 stations including nine Sacramento River stations downstream of Shasta Dam, four western valley drainages (i.e., Cottonwood Creek, Putah Creek, Cache Creek, and Skag Slough), four major river inputs (i.e., Feather, American, Mokelumne, and San Joaquin), and the Yolo and Sutter Bypass (Fig. D-1; Table D-1).

#### RESULTS

#### HYDROLOGICAL CONDITIONS

The samples were collected during the largest storm of the year when combined outflows from the basin peaked on 13 March at 297,000 CFS (Fig. D-2). Discharges from Shasta Dam were maintained at low levels during this special study (e.g., 2,300 CFS on 10 March), to minimize downstream flooding. Peak releases of approximately 68,000 CFS from Shasta Dam did not occur until 17 March (Markham *et al.*, 1996). This was also true for Keswick Reservoir which had a mean daily release of 16,100 CFS on 10 March and did not reach the peak release for WY95 of 74,800 until 17 March (Markham *et al.*, 1996). Therefore, potentially substantial metal loading, especially of cadmium, copper, and zinc, from historic mines above Shasta Dam and from the historic mines which drain into Keswick Reservoir would not have been represented in the Sacramento River for this study.

Results from this study characterize a temporal period when the basin is rapidly filling with water (Table D-2). Flows were low on the Sacramento River from Shasta Dam and Keswick Dam but increased downstream and peaked at 129,000 CFS at the Ord Ferry Bridge. The majority of river volume originated between Bend Bridge (Site 6) and Woodsen Bridge (Site 8). Sources of water in this region include several undammed creeks such as Spring (near the town of Bend), Willow, Reeds, Red Bank, Elder, Paynes, Antelope, and Mill (Table D-2). Over approximately the next 80 river miles flows decreased reaching 42,000 CFS at the City of Colusa where a weir diverts water into the Sutter Bypass. The decrease in volume from Ord Ferry to Colusa is primarily accounted for by the timing of sample collection; the pulse of water at Ord Ferry had not yet reached the Colusa site.

### METAL CONCENTRATIONS

Both metal concentrations and flow estimates are need to calculate loads. A description of metal concentrations is provided below to provide a picture, independent of flow, of the total concentration of each metal from each sampling location. The following section then combines the concentration data with flow measurements to provide an estimate of loads.

The highest total recoverable metal concentrations in the upper Sacramento River Watershed were seen in Cottonwood Creek approximately four miles upstream of the confluence with the Sacramento River. (Table D-2; Figs. D-3 to D-8). Montoya and Pan (1992) was the only reference found which indicates historic mineral activity in this watershed. Chromium was extracted from the Round Bottom mine while gold was mined from the Midas mine site. Trace metal analyses were performed on one sample collected downstream from each mine in July 1989 when flows ranged from a slow seep to less than two liters per minute (Montoya and Pan, 1992). Total concentrations of cadmium, chromium, and nickel in the Round Bottom sample were 1.2, 16, and 54  $\mu$ g/l, respectively (Montoya and Pan, 1992). Only trace concentrations of arsenic were detected at the Midas Mine (Montoya and Pan, 1992). By comparison, total recoverable cadmium, chromium, and nickel concentrations measured near the confluence of Cottonwood Creek and the Sacramento River in this study were 0.35, 150, 211  $\mu$ g/l. However there is not enough information in the literature to definitively identify the mines as the source of the high metal concentrations. Increased drainage from the mine(s) and erosion of metal rich geological deposits are other potential sources of metal enrichment measured during this storm event.

Concentrations decreased from the confluence of Cottonwood Creek and the Sacramento River to the Bend Bridge station, with an associated increased river volume (Figs. D-3 to D-8). However, concentrations increased again at Road a-8 which is near the input of many of the undammed creeks mentioned above. These data indicate the undammed creeks may be an important source of metal enrichment in the river during high flows. Concentrations of all metals measured except nickel decreased downstream from Road a-8 then increased again at the Colusa Bridge station where values were close to the those at Road a-8. This again points to undammed creeks, such as Deer and Big Chico, as potential sources for metal enrichment.

Other studies reported unknown sources of metals upstream of Sacramento were responsible for increased metal concentrations in the lower Sacramento River (Larry Walker & Associates, 1997; Alpers, written comm.; Foe and Croyle, 1998). Larry Walker & Associates (1997) reported the largest loads of mercury in the Sacramento River occurred during storm events and originated from above the Feather River. Alpers (written comm.) conducted a metals transport study during both wet and dry weather and consistently noted an increase in mercury load in the Sacramento River between Redding and Colusa. Increased loads of other metals, such as lead and copper, were noted for the Sacramento River between Keswick Dam and Bend Bridge (Charlie Alpers, written comm.). However, neither study identified the source(s). In addition, it is not clear from these studies if other metals are enriched along this stretch of river. To address this question, one must compare the results of this study with those of Foe and Croyle (1998). Samples for both studies were collected at the same time for the metals source components. Mercury followed the same pattern in upper Sacramento River, with enrichment between Bend Bridge and Ord Ferry (Foe and Croyle, 1998). Detailed follow-up studies are needed to identify the major source(s) of these metals along this stretch of river.

During high flow conditions, a weir is opened on the Sacramento River near the Colusa station. River water enters the Sutter Bypass which eventually drains into the Yolo Bypass. Samples collected from the Sutter Bypass downstream of the Colusa station had greatly reduced metal concentrations, suggesting a dilution effect or settling (Table D-2; Figs. D-9 to D-14). However, Sacramento Slough which runs parallel to the Bypass had concentrations as high as those measured in Cottonwood Creek. Both the Sutter Bypass and Sacramento Slough are not well mixed at the sample stations during high flow events and can contain water from the Sacramento River, the Colusa Basin Drain, and several small creeks and sloughs. The complex hydrology in the Sutter Bypass and Sacramento Slough during high flows makes interpretation of metal concentrations at these stations difficult.

Several stations which discharge into the Yolo Bypass, and eventually the north Delta, were monitored for total recoverable metals. Cache Creek was sampled a short distance upstream of where it discharges into the Bypass. Concentrations of all metals were 150% to approximately 300% higher than at Cottonwood Creek (Table D-2; Figs. D-9 to D-14). Concentrations in Putah Creek prior to discharging into the Bypass were much higher than most main river stations. The west and east side of the Yolo Bypass was monitored near Interstate 80 in the region receiving water from Cache Creek, Putah Creek, Colusa Basin Drain, the Sacramento River, and the Sutter Bypass. Concentrations on the east side were consistently higher than those on the west side, indicating the Bypass is not well mixed during such high flow events. Concentrations on the east side were by far the highest concentrations measured during this survey.

One station was selected to quantify metal concentrations entering the Delta from the San Joaquin River. Metal concentrations in the San Joaquin River at Vernalis were moderately high when compared to those in the upper Sacramento River and Yolo Bypass (Table D-2; Figs. D-9 to D-14).

The pattern of total recoverable metal concentrations was quite different in the lower Sacramento River. The Feather and American Rivers are the primary tributaries which enter the Sacramento River in the lower watershed. Metal concentrations in the Feather and American Rivers were much lower than the upper Sacramento River (Table D-2; Figs. D-9 to D-14). Water from the Sacramento River above the Feather and American Rivers begins to enter the Yolo Bypass when flows exceed 60,000 CFS. All additional water in the river is diverted into the Bypass when flows reach 100,000 CFS. The combined discharges of the Feather and American River was approximately 112,000 CFS on 11 March. Therefore, most of the water reaching Greene's Landing during this study is expected to have come from these two watersheds while most water in the upper Sacramento River would flow into the Bypass. For reasons which are unclear, metal concentrations at Greene's Landing were greater than those in the Feather and American Rivers. Possible explanations include, but are not limited to, a sediment bedload source during high flows, urban runoff from storm drains in Sacramento and West Sacramento, and/or municipal sewage treatment plants along the Sacramento River, although municipal sources were unlikely to be of sufficient magnitude.

### METAL LOADS

Load calculations were point estimates for the load tracking study because a one time analysis of metals was performed at each station.

Overall conclusions for load estimates in this study may be limited or incomplete due to the lack of measured flows at several stations. In addition, flows out of Shasta Dam and Keswick Reservoir were maintained at low levels during the storm event which resulted in an incomplete description of metal loading from mines which drain into these two water bodies. However, similar patterns determined for the metal analysis component of the source study emerged when metal loads were assessed. A significant sources of metal load to the upper Sacramento River during the storm was Cottonwood Creek (Table D-2; Figs. D-3 to D-8). Additional significant sources of metal loads entered the river between Bend Bridge and the Ord Ferry Road Bridge, again pointing toward undammed creeks as sources along this stretch of river. Cache Creek contributed significant loads to the lower stretches of the watershed (Table D-2; Figs. D-9 to D-14). In fact, Cache Creek loads exceeded those of Cottonwood Creek. These results confirm that Cache Creek is a major source of metals during high flow years. Although metal concentrations in Putah Creek were among the highest measured in the study, loads were relatively low due to low flows when compared to other stations. Many of the load estimates measured during the short sampling period for the metal source study exceeded the average daily loads entering the Delta during WY95 (Table 57 & 59). Data obtained from this study indicate major storm events can contribute significant metal loads to the river. However, stations monitored for the metals source study did not provide an assessment of metal loads in the entire Sacramento River Watershed because samples were not collected from sites where metal loads are most heavily influenced by upstream sources of metals such as historic base-metal mining. Additional studies should be performed to identify sources of loads between Bend Bridge and the Ord Ferry Road Bridge. In addition, this study should be repeated over a wider temporal period, should include flow

measurements at all stations to better characterize loads into the system, and incorporate stations which would permit a characterization of metal loading from mining activities.

### SUMMARY OF RECOMMENDATIONS

2

1. Repeat the metals source study on the Sacramento River from Shasta Dam to Greene's Landing and the Yolo Bypass during major rain events to better characterize metal and sediment loads in the system. Incorporate flow measurements at all stations where such studies are performed to permit calculations of loads. In addition, apply more rigorous load calculation methods such as those in Cohn *et al.*, (1989). Measurements of dissolved metals should be incorporated into future studies in this region to permit an assessment of compliance with water quality objectives. Furthermore, a toxicity assessment should be incorporated into the overall study design.

2. Conduct a special study on the Sacramento River downstream from the Bend River Bridge to the Ord Ferry Bridge during major storm events to characterize the sources of increased flows, metal concentrations, and loads. Monitoring should include stations in undammed creeks including Spring (near the town of Bend), Reeds, Red Bank, Elder, Paynes, Antelope, and Mill. Dissolved metal concentrations should be measured as well to permit an assessment of water quality objective exceedances. Load calculations should follow current methods which are more rigorous than those applied in this report.

3. Conduct a special study on the Sacramento River downstream from County Road A-8 to Colusa during major storm events to characterize sources of enriched metal concentrations along this stretch of the Sacramento River. Samples should be collected from Big Chico and Mill Creeks which are sources of water to the river in this area. Dissolved metal concentrations should be measured as well to permit an assessment of water quality objective exceedances.

4. Additional studies should be performed during high flow years when the Yolo Bypass is operational to better characterize the source(s) of elevated metal concentrations at Greene's Landing reported in this study when compared to concentrations in the American and Feather River.

## **DESCRIPTION OF SAMPLING LOCATIONS**

Sacramento River @ Shasta Dam (site 1): Sample collected from east bank below Shasta Dam at Powerhouse.

Sacramento River @ Cypress Bridge (site 2): Sample collected in mid channel from Cypress Avenue bridge.

Little Cow Creek (site 3): Sample collected from mid channel off the Dersch Road Bridge outside of Anderson.

Sacramento River @ Balls Ferry (site 4): Sample collected in mid channel from Balls Ferry Road bridge.

<u>Cottonwood Creek (site 5)</u>: Sample collected in mid channel off HWY 5 frontage road bridge about one mile south of the town of Cottonwood.

Sacramento River @ Bend (site 6): Sample collected in mid channel from Bend bridge Park.

Sacramento River @ Road a-8 (site 7): Sample collected in mid channel off County Road A8 bridge near Tehema and the Mills Creek Recreation Area.

<u>Sacramento River @ Road a-9 (site 8)</u>: Sample collected in mid channel from South Avenue bridge at Woodson State Recreation Area.

Sacramento River @ Ord Ferry (site 9): Sample collected in mid channel from Ord Ferry Road bridge.

Sacramento River @ Colusa (site 10): Sample collected on west side of channel off River Road bridge.

<u>Sutter Bypass (site 11)</u>: Sample collected about one third of way across Bypass on north side of channel off HWY 113 bridge.

Sacramento Slough (site 12): Sampled from the Reclamation District pumphouse at Karnack.

**Feather River (site 13)**: Sample collected by wading off intersection of Garden Highway and Lee Road.

American River (site 14): American River sample collected in mid channel off bridge at Sacramento State University in the City of Sacramento.

Site Name	Date Sampled
American R. Sac State	3/11/95
Cache Creek @ Road 102	3/11/95
Cache Creek @ Road 102	3/11/95
Cottonwood Creek	3/10/95
Cottonwood Creek	3/10/95
East Yolo bypass	3/10/95
Feather R. @ Highway 99	3/11/95
Little Cow Cr. Dersch Br.	3/10/95
Little Cow Cr. Dersch Br.	3/10/95
Mokelumne River	3/11/95
Mokelumne River	3/11/95
Putah Creek @ Mace Blvd.	3/10/95
Sac R. @ Shasta Dam	3/10/95
Sac R. @ Balls Ferry Br.	3/10/95
Sac R. @ Bend Bridge	3/10/95
Sac R. @ Colusa Bridge	3/10/95
Sac R. @ Cypress Bridge	3/10/95
Sac R. @ Ord Ferry	3/10/95
Sac R. @ Road a-8	3/10/95
Sac R. @ Road a-9	3/10/95
Sacramento Slough	3/10/95
Skag Slough	3/10/95
Sutter Bypass	3/13/95
S.J. River @ Vernalis	3/11/95
West Yolo Bypass	3/10/95

Table D-1. Sites and Dates of Sampling for the Metals Source Study

¢,

					Total Cu	Cu Load	Total Zn	Zn Load	Total Cr	Cr Load
Date	Hour	Station #	Station Name	Flow (cfs)	(μg/l)	(kg)	(µg/l)	(kg)	(µg/l)	(kg)
3/10/95	800	bp103	Sac. River @ Shasta Dam	2300	1.23	6.92	4.6	25.87	1.44	8.10
3/10/95	1000	bp97	Sac. River @ Cypress Br.	18000	8.23	362.20	18.7	822.99	2.03	89.34
3/10/95	1115	bp106	Little Cow Creek @ Dersch Br.	10000	12.4	303.18	33	806.85	7.39	180.56
3/10/95	1230	bp104	Sac. River @ Balls Ferry Br.		10.7		29.6		6.5	
3/10/95	1330	bp102	Cottonwood Creek	21000	92.4	4744.28	170	8728.65	150	7701.75
3/10/95	1430	bp105	Sac. River @ Bend Br.	67000	28.8	4717.87	68.8	11270.47	39.6	6487.07
3/10/95	1550	bp99	Sac. River @ Road a-8		70.4		157		150	
3/10/95	1700	bp107	Sac. River @ Road a-9	102000	56.6	14115.47	134	33418.26	99.6	24839.24
3/10/95	1830	bp98	Sac. River @ Ord Ferry	129000	46.8	14760.95	97.2	30657.37	75.7	23876.16
3/10/95	2000	bp100	Sac. River @ Colusa Br.	42000	58.1	5966.29	129	13247.01	94.8	9735.01
3/11/95	1630	bp111	Feather R. Highway 99	34500	4.54	382.96	6.29	530.58	3.14	264.87
3/11/95	1530	bp110	American R. @ Sac. State	77800	1.15	218.75	3.87	736.16	1.28	243.48
3/11/95	1300	CF 800	Sac. River @ Greene's Landing	99000	8.6	2081.67	19.8	4792.69	13.8	3340.36
3/11/95	1500	CF 801	Mokelumne River		4.55		11.19		3.14	
3/13/95	1100	CF 803	Sutter Bypass		12		24.8		17.6	
3/10/95	2230	bp101	Sacramento Slough		73.2		173		122	
3/11/95	1200	bp109	Cache Creek @ Road 102	17500	140.5	6011.64	288.5	12344.19	291	12451.16
3/10/95	1240	bp108	Putah Creek @ Mace Blvd.	682	76.9	128.23	253	421.87	98.4	164.08
3/10/95		bp114	East Yolo Bypass		121		333		303	
3/10/95		bp113	West Yolo Bypass		43		144		90	
3/10/95		bp112	Skag Slough		5.22		15.3		4.82	
3/11/95	1600	CF 802	Vernalis	7830	34.1	652.82	107	2048.45	69.1	1322.87

~

Table D-2. Total recoverable metal concentrations, metal loads, and flows in the Sacramento River Watershed during the largest storm event of the year in March 1995.

10

.

.

.

Date	Hour	Station #	Station Name	Flow (cfs)	Total Pb (µg/l)	Pb Load (kg)	Total Cd (µg/l)	Cd Load (kg)	Total Ni (µg/l)	Ni Load (kg)
3/10/95	800	bp103	Sac. River @ Shasta Dam	2300	2.68	15.07	0.026	0.15	2.36	13.27
3/10/95	1000	bp97	Sac. River @ Cypress Br.	18000	0.83	36.53	0.11	4.84	2.3	101.22
3/10/95	1115	bp106	Little Cow Creek @ Dersch Br.	10000	6.9	168.71	0.114	2.79	7.09	173.35
3/10/95	1230	bp104	Sac. River @ Balls Ferry Br.		4.32		0.154		7.41	
3/10/95	1330	bp102	Cottonwood Creek	21000	19.9	1021.77	0.353	18.12	211	10833.80
3/10/95	1430	bp105	Sac. River @ Bend Br.	67000	7.68	1258.10	0.2	32.76	52	8518.38
3/10/95	1550	bp99	Sac. River @ Road a-8		15.7		0.371		492	
3/10/95	1700	bp107	Sac. River @ Road a-9	102000	12.9	3217.13	0.377	94.02	112	27931.68
3/10/95	1830	bp98	Sac. River @ Ord Ferry	129000	10.2	3217.13	0.296	93.36	251	79166.66
3/10/95	2000	bp100	Sac. River @ Colusa Br.	42000	12.1	1242.55	0.409	42.00	266	27315.54
3/11/95	1630	bp111	Feather R. Highway 99	34500	0.72	60.73	0.026	2.19	4.06	342.47
3/11/95	1530	bp110	American R. @ Sac. State	77800	0.44	83.70	0.017	3.23	2.17	412.78
3/11/95	1300	CF 800	Sac. River @ Greene's Landing	99000	3.04	735.85	0.16	38.73	13.2	3195.13
3/11/95	1500	CF 801	Mokelumne River		3.93		0.05		4.17	
3/13/95	1100	CF 803	Sutter Bypass		4.88		0.068		20.4	
3/10/95	2230	bp101	Sacramento Slough		17.5		0.433		120	
3/11/95	1200	bp109	Cache Creek @ Road 102	17500	30.6	1309.30	0.403	17.24	652	27897.45
3/10/95	1240	bp108	Putah Creek @ Mace Blvd.	682	28	46.69	0.47	0.78	88.1	146.91
3/10/95		bp114	East Yolo Bypass		33.3		0.438		600	
3/10/95		bp113	West Yolo Bypass		15.6		0.311		165	
3/10/95		bp112	Skag Slough		4.66		0.057		14.1	
3/11/95	1600	CF 802	Vernalis	7830	17.6	336.94	0.169	3.24	128	2450.48

Table D-2 (cont). Total recoverable metal concentrations, metal loads, and flows in the Sacramento River Watershed during the largest storm event of the year in March 1995.

÷.

Ð



Figure D-1. Map of the Sacramento River Watershed and its major tributaries. Numbers refer to sample stations described in Appendix A.



Figure D-2. Precipitation and flow pattern in the Sacramento Basin during the winter and spring of 1995. Arrow indicates sampling for the metals source study.



Figure D-2. Precipitation and flow pattern in the Sacramento Basin during the winter and spring of 1995. Arrow indicates sampling for the metals source study.

,



Figure D-3. Schematic of copper loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



Figure D-4. Schematic of zinc loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



Figure D-5. Schematic of chromium loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



Figure D-6. Schematic of lead loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



Figure D-7. Schematic of cadmium loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



c

Figure D-8. Schematic of nickel loads, total recoverable concentrations, and water flow in the upper Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest an unknown riverine cadmium source between Bend (site 6) and Woodson Bridge (site 8).



¢

Figure D-9. Schematic of copper loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).



Figure D-10. Schematic of zinc loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).

. . .


e

Figure D-11. Schematic of chromium loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).



Figure D-12. Schematic of lead loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).



e

Figure D-13. Schematic of cadmium loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).



Figure D-14. Schematic of nickel loads, total recoverable concentrations, and water flow in the lower Sacramento River during the largest storm event of the year in March 1995. Small circles with numbers represent stations described in Appendix A. Results suggest enrichment of cadmium at Cache Creek (site 16), Putah Creek (site 17), and the Sacramento River at Greene's Landing (site 15).