Regional Harbor Monitoring Program
Pilot Project
2006-07

Final Report

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EXECUTIVE SUMMARY

The Regional Harbor Monitoring Program (RHMP) was developed by the Port of San Diego, City of San Diego, City of Oceanside, and County of Orange to address questions regarding the general water quality and condition of aquatic life in the four harbors within Region 9 (San Diego) of the State Water Resources Control Board. The RHMP was developed as a core monitoring program to address the overall condition of the harbors with supplemental focused studies to answer specific questions. The core monitoring program assesses the conditions found in the harbors and compares to reference values that are based on historical data from the four harbors.

A Pilot Program for the RHMP began in 2005 and is designed to run for three years. The objective of the Pilot Program is to implement the RHMP core monitoring on a limited scale to verify the study design. Data from two strata (marinas and freshwater influence) will be statistically evaluated to establish the frequency of core monitoring needed to assess trends in water and sediment quality. Comparison of the pilot project data to historical data used in setting the threshold levels and target percentages is not a direct comparison because the historical data were collected throughout the harbors and include data from potentially cleaner sediments in the more open parts of the harbors. The comparisons in this report are made to verify elements of the study design and are not intended to make conclusions on the conditions in the harbors. This report only presents the results of the water and sediment sampling performed in August 2006, and provides indications of the conditions in the marina and freshwater influenced areas of the harbors during the August 2006 sampling event. The Pilot Program will be assessed as to the appropriateness of the study design upon completion of the third year of sampling to be performed during the summer of 2007.

Based on the current year of data, the following statements can be made:

- Water column concentrations of copper in marinas are higher than those found in freshwater influenced areas and the proportion of marina samples with elevated concentrations is higher than the proportion found historically throughout the harbors.
- Concentrations of other metals and polyaromatic hydrocarbons in the water column were below water quality objectives.
- All bacterial concentrations were below AB 411 levels.
- Measurements of sediment quality were mixed compared to historical concentrations for metals.
- Biological indicators for benthic infauna suggest poorer habitat quality in both the marinas and freshwater influenced areas than was found historically throughout the harbors.
- Sediment toxicity tests may imply healthier conditions in both the marinas and freshwater influenced areas than found historically.

1.0 INTRODUCTION

The Regional Harbor Monitoring Program (RHMP) was developed by the Port of San Diego, County of Orange, the City of San Diego, and the City of Oceanside in response to a July 24, 2003 request by the San Diego Regional Water Quality Control Board (SDRWQCB) under §13225 of the California Water Code. The RHMP is a comprehensive effort to survey the general water quality and condition of aquatic life in the harbors and to determine whether beneficial uses are being met for the following four local harbors: San Diego Bay, Mission Bay, Oceanside Harbor, and Dana Point Harbor. The program is composed of a core monitoring program with potential focused studies to answer specific questions. The core monitoring program was designed to address the following five major questions posed in the SDRWQCB's request:

1. What are the contributions and spatial distributions of inputs of pollutants to harbors in the San Diego Region and how do these inputs vary over time?
2. Are the waters in the harbors safe for body contact activities?
3. Are fish in harbors safe to eat?
4. Do the waters and sediments in the harbors sustain healthy biota?
5. What are the long-term trends in water quality for each harbor?

Implementation of the RHMP began with a Pilot Program to verify the design of the program. Prior to the initiation of sampling in the Pilot Program the following tasks were completed to finalize the design:

- Acquire and analyze relevant available historical information.
- Complete detailed mapping to verify stratum areas.
- Identify the indicators to be monitored.
- Establish the threshold levels.
- Establish the preset target proportions.
- Develop a Quality Assurance Project Plan.

A key element of the RHMP is the identification of strata within and across the harbors. This element was modeled after the regional Southern California Bight (Bight) and national Environmental Monitoring and Assessment Program (EMAP). The use of strata allows delineation of harbor inputs based on activities within each area. Five strata were identified for monitoring in the core program: marinas, industrial/port, freshwater influenced, shallow water, and deep water. The shallow and deep water strata encompass all areas within the harbors not within the other three specific strata. The freshwater influenced stratum includes areas that may be affected by input from streams or storm water runoff (Weston, 2005a).

Sampling for the Pilot Program began in August 2005. The Pilot Program is designed to run for three years with sampling conducted during the summer months; full implementation of the RHMP will occur in 2008 and coincide with the next regional Southern California Bight monitoring program. The Pilot Program is a scaled down version of the RHMP that focuses on a limited number of indicator measurements and samples in two of the five identified strata. The strata sampled in the Pilot Program, marinas and freshwater influenced, were selected because the variability within them is anticipated to be greater than in the other three strata and thus will provide a conservative estimate of the amount of sampling needed to detect trends.

Statistical analysis of the data obtained in the Pilot Program is performed with the application of a binomial model. This approach compares proportions of samples below (or above) an established threshold with an established target proportion and determines whether there is a significant increase in the proportion of samples below the threshold compared to the historical data. The proportions can be tracked through time as the program progresses to measure improvement in the health of the harbors.

This report presents the results of the second year of the Pilot Program for the RHMP. As such, it focuses on the measurements made in August 2006 for the marina and freshwater influenced strata in the four harbors, making comparisons using the binomial approach to historical data and compares the results of the two sampled strata to each other.

2.0 METHODS

2.1 Field Sampling

2.1.1 Station Selection

Station selection in Dana Point Harbor, Oceanside Harbor, Mission Bay, and San Diego Bay is based on a stratified random sampling design similar to that used in southern California regional monitoring programs. Uniformly sized hexagons were overlaid on maps of each of the bays. The size of the hexagons was determined by the smallest freshwater influenced area that could be safely sampled. Hexagons were set at 100 feet (ft) (~65 meters) per side with the nominal sampling station at the center of the hexagon. Ten stations were then randomly selected in the marina and freshwater influenced strata with the stipulation that at least one station was set in each harbor. Because Oceanside Harbor has no freshwater influenced strata, two marina stations were selected for that water body. All of the sampling occurred within a 30 meter (m) radius of the nominal station coordinates. The coordinates of the actual sampling locations were recorded when sampling occurred.

The locations of the sampling stations in each of the harbors are shown in Figures 2-1 to 2-6. Marina stations in Dana Point Harbor and Oceanside Harbor were located near boat slips throughout the harbors. Marina stations in Mission Bay were located in Quivera Basin and Santa Barbara Cove. In San Diego Bay, marina stations were located in the Shelter Island Yacht Basin, America's Cup Harbor, Sunroad Resort Marina, and the Coronado Cays. Some marina stations in both Mission Bay and San Diego Bay were in the same marina areas that were sampled in 2005, while other stations were in different marina areas.

The one freshwater influenced station in Dana Point Harbor was located adjacent to a storm drain. No freshwater influenced areas were identified in Oceanside Harbor. Freshwater influenced stations in Mission Bay were located near Rose Creek Inlet. In San Diego Bay, they were located in South Bay near the power plant and at the mouths of Chollas Creek and Sweetwater River. As with the marina stations, some of these areas were previously sampled in 2005 while others were sampled for the first time in 2006.

Figure 2-1. Sampling stations in Dana Point Harbor for 2006

Figure 2-2. Sampling stations in Oceanside Harbor for 2006

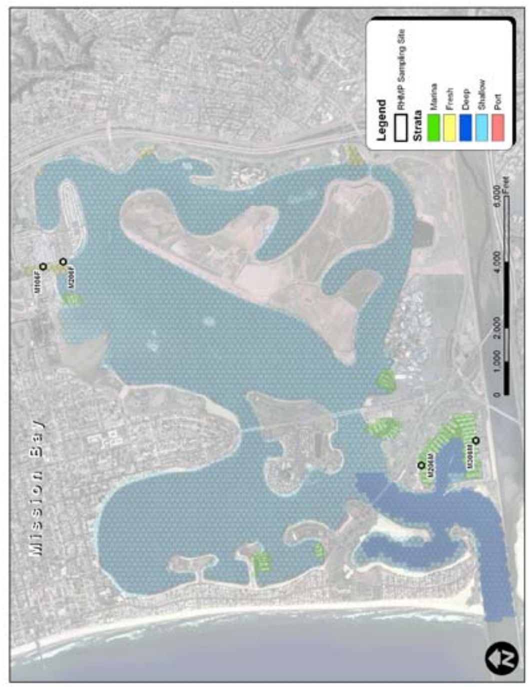

Figure 2-3. Sampling stations in Mission Bay for 2006

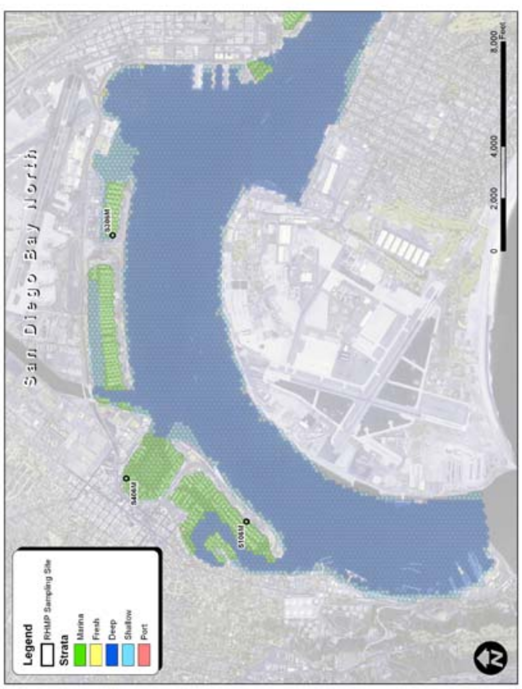

Figure 2-4. Sampling stations in northern San Diego Bay for 2006

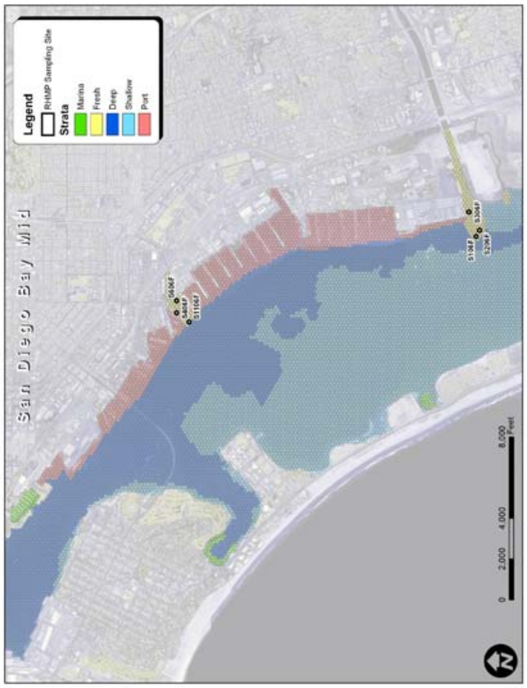

Figure 2-5. Sampling stations in central San Diego Bay for 2006

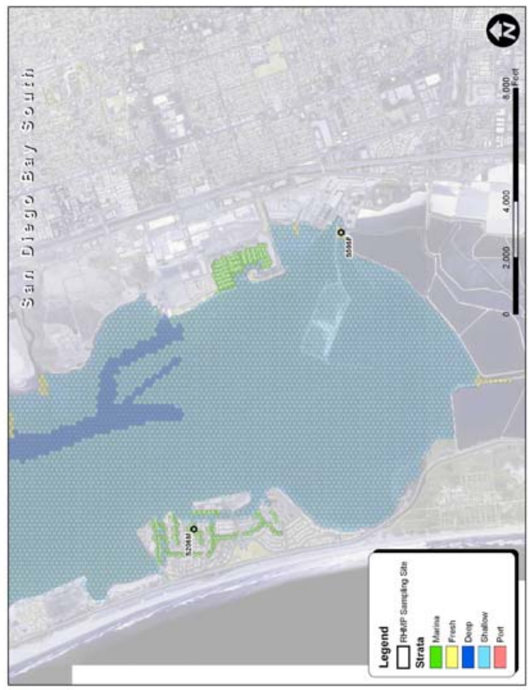

Figure 2-6. Sampling stations in southern San Diego Bay for 2006

2.1.2 Water Quality Sampling

Water column sampling was performed by Weston in August 2006. A total of twenty stations were sampled, ten marina and ten freshwater influenced. Actual coordinates for sample locations were recorded on field data sheets. The actual locations of the stations and sampling dates are listed in Table 2-1.

Water column sampling was conducted using a Seabird SBE-25 Sealogger CTD (conductivity-temperature-depth) equipped with sensors that measure temperature, specific conductance, dissolved oxygen (DO), pH, and light transmission. Casts were taken at stations located with a differential global positioning system (dGPS). Dissolved oxygen and pH sensors were calibrated prior to the week of monitoring. Transmissivity, conductivity, and temperature are calibrated annually by Sea-Bird Electronics, Inc. Before beginning a cast, a 3-minute equilibration was performed to bring the CTD sensors to thermal equilibration with the ambient sea water and to ensure that all of the pumps have turned on. The CTD was lowered at a speed of 0.25-0.50 m/sec until it was within 1m of the bottom. The instrument operated at a scan rate that recorded 8 scans/sec.

After casts in each harbor were performed, the data were uploaded and saved onto a field computer. Data were checked to ensure the CTD turned on properly, the depth was accurate, and that all water quality measurements were recorded throughout a cast. Data were transferred to a disk upon returning to the laboratory. A post cruise calibration was performed following the week of sampling.

Discrete water samples were collected at each station one meter below the surface using a Niskin bottle. Water samples were transferred to labeled sample containers. Additional data such as the weather, wind speed and direction, and water color and odor were recorded on field data sheets. Samples to be analyzed for total organic carbon (TOC), dissolved organic carbon (DOC), dissolved and total metals, total hardness (as $CaCO_3$), and polynuclear aromatic hydrocarbons (PAHs) were sent to CRG Marine Laboratories, Inc. (CRG) for chemical analyses. The CTD profiles and the samples for indicator bacteria, Enterococcus, were retained and analyzed by Weston Solutions, Inc. All of the samples were sent to the designated laboratories under the proper storage conditions and within holding times.

2.1.3 Sediment Sampling

Sediment sampling was performed in August 2006 at the same stations as those sampled for water quality using a dGPS to locate the stations. Field observations and actual coordinates for sample locations were recorded on sediment sampling data forms. Table 2-1 shows the actual locations of the stations and sampling dates.

Table 2-1. Station locations and the dates they were sampled

Harbor	Station ID	Strata	Date Sampled	Water Sampling		Sediment Sampling	
				Latitude	Longitude	Latitude	Longitude
Dana Point Harbor	D106M	M	August 21, 2006	33º 27' 39.90" N	117º42' 03.24" W	33º 27' 40.14" N	117º42' 03.78" W
	D206M	M	August 21, 2006	33º 27' 38.58" N	117º42' 01.68" W	33º 27' 38.22" N	117º42' 01.44" W
	D106F	F	August 21, 2006	33º 27' 36.00" N	117º41' 40.20" W	33º 27' 35.76" N	117º41' 40.62" W
Oceanside Harbor	O406M	M	August 21, 2006	33º 12' 18.66" N	117º 23' 24.36" W	33º 12' 18.72" N	117º 23' 24.18" W
	O506M	M	August 21, 2006	33º 12' 45.42" N	117º 23' 41.04" W	32º 12' 45.30" N	117º 23' 40.44" W
Mission Bay	M206M	M	August 22, 2006	32º 45' 52.32" N	117º 14' 22.92" W	32º 45' 52.26" N	117º 14' 22.62" W
	M306M	M	August 22, 2006	32º 45' 36.12" N	117º 14' 12.72" W	32º 45' 36.12" N	117º 14' 12.78" W
	M106F	F	August 22, 2006	32º 47' 45.60" N	117º 13' 13.02" W	32º 47' 44.82" N	117º 13' 13.08" W
	M206F	F	August 22, 2006	32º 47' 38.94" N	117º 13' 11.40" W	32º 47' 39.54" N	117º 13' 11.40" W
San Diego Bay	S106M	M	August 23, 2006	32º 42' 42.60" N	117º 13' 49.74" W	32º 42' 42.66" N	117º 13' 49.20" W
	S206M	M	August 23, 2006	32º 37' 38.76" N	117º 07' 58.08" W	32º 37' 38.64" N	117º 07' 58.56" W
	S306M	M	August 23, 2006	32º 42' 34.14" N	117º 11' 41.82" W	32º 43' 34.20" N	117º 11' 42.24" W
	S406M	M	August 23, 2006	32º 43' 28.50" N	117º 13' 29.76" W	32º 43' 28.56" N	117º 13' 30.66" W
	S106F	F	August 23, 2006	32º 38' 46.44" N	117º 07' 22.38" W	32º 38' 46.32" N	117º 07' 22.14" W
	S206F	F	August 23, 2006	32º 38' 44.94" N	117º 07' 18.78" W	32º 38' 44.70" N	117º 07' 18.60" W
	S306F	F	August 23, 2006	32º 38' 49.98" N	117º 07' 08.22" W	32º 38' 49.98" N	117º 07' 08.16" W
	S406F	F	August 24, 2006	32º 41' 10.86" N	117º 08' 07.56" W	32º 41' 10.20" N	117º 08' 07.98" W
	S506F	F	August 23, 2006	32º 36' 48.00" N	117º 05' 55.44" W	32º 36' 47.76" N	117º 05' 55.26" W
	S606F	F	August 24, 2006	32º 41' 11.10" N	117º 08' 01.20" W	32º 41' 10.80" N	117º 08' 01.14" W
	S1106F	F	August 24, 2006	32º 41' 05.10" N	117º 08' 12.96" W	32º 41' 04.74" N	117º 08' 13.02" W

Benthic sediments were collected using a stainless steel, 0.1m^2 Van Veen grab sampler. A minimum of four sediment grabs per station were collected for the following: benthic infauna, acute toxicity, grain size, and chemistry including TOC, total metals, and PAHs. A sample was determined to be acceptable if the surface of the grab was even, there was minimal surface disturbance, and there was a penetration depth of at least 5 centimeters (cm). Rejected grabs were discarded and re-sampled.

Samples collected for infaunal analysis were rinsed through a 1.0 mm mesh screen and transferred to a labeled quart jar. A 7% magnesium sulfate ($MgSO_4$) seawater solution was added for approximately 30 minutes to relax the collected specimens. The samples were then fixed in a 10% buffered formalin solution. Infaunal samples were retained and analyzed by Weston.

Samples for acute toxicity and sediment chemistry were collected from the top 2 cm of the grab. Sediment within 1 cm of the sides of the grab was avoided. A total of 3 liters (L) of sediment was collected for toxicity and placed in three 1-L jars. Toxicity samples were then kept at 4°C on ice in coolers. Sediment for trace metals and organics (PAHs) analysis was placed in one 4-ounce jar. These samples were stored at 4°C on ice and frozen within 24 hours. Approximately 150-200 grams of sediment were collected for TOC and grain size. Samples were each placed in one quart sized Ziploc™ bag and kept on ice. The TOC samples were frozen within 24 hours of collection and the grain size stored at 4°C. The samples for acute toxicity, grain size, and TOC were retained and analyzed by Weston. The samples for trace metals and PAHs were shipped frozen to CRG Marine Laboratories, Inc. within a week of collection.

2.2 Laboratory Analysis

2.2.1 Chemistry

Chemical analyses were performed on both water and sediment samples; a complete list of chemical analytes with analytical methods and detection limits is provided in Table 2-2. For the water samples, the analyses included total organic carbon (TOC) and dissolved organic carbon (DOC), total and dissolved metals, total hardness measured as $CaCO_3$, and polyaromatic hydrocarbons (PAHs). For the sediment samples, TOC, trace metals and PAHs were analyzed. All chemical analyses were conducted to meet or exceed the specifications of the Surface Water Ambient Monitoring Program (SWAMP). Sediment samples were also analyzed for grain size to provide data on the size distributions of the sediment (gravel, sand, silt, and clay).

Table 2-2. RHMP Constituents to be Monitored and Corresponding Analytical Methods

Analyte	Method	Reporting Limit	Units
Water Samples			
pH	Collected in field	-	-
Specific Conductance	Collected in field	-	µS/cm
Dissolved Oxygen	Collected in field	-	mg/L
Temperature	Collected in field	-	ºC
Salinity	Collected in field	-	PSU
Transmissivity	Collected in field	-	%
Total Organic Carbon	EPA 415.1	1	%
Dissolved Organic Carbon	EPA 415.1	0.5	%
Total Hardness as CaCO3	SM 2340B	5.00	mg/L
Enterococcus	SM 9223	< 10	MPN/100ml
Dissolved Metals			
Aluminum (Al)	EPA 1640	0.125	µg/L
Antimony (Sb)	EPA 1640	0.015	µg/L
Arsenic (As)	EPA 1640	0.015	µg/L
Beryllium (Be)	EPA 1640	0.01	µg/L
Cadmium (Cd)	EPA 1640	0.01	µg/L
Chromium (Cr)	EPA 1640	0.01	µg/L
Cobalt (Co)	EPA 1640	0.01	µg/L
Copper (Cu)	EPA 1640	0.01	µg/L
Iron (Fe)	EPA 1640	0.025	µg/L
Lead (Pb)	EPA 1640	0.01	µg/L
Manganese (Mn)	EPA 1640	0.01	µg/L
Mercury (Hg)	EPA 245.7	0.02	µg/L
Molybdenum (Mo)	EPA 1640	0.01	µg/L
Nickel (Ni)	EPA 1640	0.01	µg/L
Selenium (Se)	EPA 1640	0.015	µg/L
Silver (Ag)	EPA 1640	0.01	µg/L
Thallium (Tl)	EPA 1640	0.01	µg/L
Tin (Sn)	EPA 1640	0.01	µg/L
Titanium (Ti)	EPA 1640	0.01	µg/L
Vanadium (V)	EPA 1640	0.01	µg/L
Zinc (Zn)	EPA 1640	0.01	µg/L
Total Metals			
Aluminum (Al)	EPA 1640	0.125	µg/L
Antimony (Sb)	EPA 1640	0.015	µg/L
Arsenic (As)	EPA 1640	0.015	µg/L
Beryllium (Be)	EPA 1640	0.01	µg/L
Cadmium (Cd)	EPA 1640	0.01	µg/L
Chromium (Cr)	EPA 1640	0.01	µg/L
Cobalt (Co)	EPA 1640	0.01	µg/L
Copper (Cu)	EPA 1640	0.01	µg/L
Iron (Fe)	EPA 1640	0.025	µg/L
Lead (Pb)	EPA 1640	0.01	µg/L
Manganese (Mn)	EPA 1640	0.01	µg/L
Mercury (Hg)	EPA 245.7	0.02	µg/L
Molybdenum (Mo)	EPA 1640	0.01	µg/L
Nickel (Ni)	EPA 1640	0.01	µg/L

Table 2-2. RHMP Constituents to be Monitored and Corresponding Analytical Methods

Analyte	Method	Reporting Limit	Units
Selenium (Se)	EPA 1640	0.015	µg/L
Silver (Ag)	EPA 1640	0.01	µg/L
Thallium (Tl)	EPA 1640	0.01	µg/L
Tin (Sn)	EPA 1640	0.01	µg/L
Titanium (Ti)	EPA 1640	0.01	µg/L
Vanadium (V)	EPA 1640	0.01	µg/L
Zinc (Zn)	EPA 1640	0.01	µg/L
Polynuclear Aromatic Hydrocarbons			
1-Methylnaphthalene	EPA 625	5	ng/L
1-Methylphenanthrene	EPA 625	5	ng/L
2,3,5-Trimethylnaphthalene	EPA 625	5	ng/L
2,6-Dimethylnaphthalene	EPA 625	5	ng/L
2-Methylnaphthalene	EPA 625	5	ng/L
Acenaphthene	EPA 625	5	ng/L
Acenaphthylene	EPA 625	5	ng/L
Anthracene	EPA 625	5	ng/L
Benz[a]anthracene	EPA 625	5	ng/L
Benzo[a]pyrene	EPA 625	5	ng/L
Benzo[b]fluoranthene	EPA 625	5	ng/L
Benzo[e]pyrene	EPA 625	5	ng/L
Benzo[g,h,i]perylene	EPA 625	5	ng/L
Benzo[k]fluoranthene	EPA 625	5	ng/L
Biphenyl	EPA 625	5	ng/L
Chrysene	EPA 625	5	ng/L
Dibenz[a,h]anthracene	EPA 625	5	ng/L
Fluoranthene	EPA 625	5	ng/L
Fluorene	EPA 625	5	ng/L
Indeno[1,2,3-c,d]pyrene	EPA 625	5	ng/L
Naphthalene	EPA 625	5	ng/L
Perylene	EPA 625	5	ng/L
Phenanthrene	EPA 625	5	ng/L
Pyrene	EPA 625	5	ng/L
Sediment Samples			
Total Organic Carbon	EPA 415.1	0.05	%
Grain Size Analysis	Plumb 1981	-	-
Acute Toxicity	EPA/600/R-94/025	-	%
Benthic Infauna	-	-	-
Total Metals			
Aluminum (Al)	EPA 6020	5	mg/kg
Antimony (Sb)	EPA 6020	0.05	mg/kg
Arsenic (As)	EPA 6020	0.05	mg/kg
Barium (Ba)	EPA 6020	0.05	mg/kg
Beryllium (Be)	EPA 6020	0.05	mg/kg
Cadmium (Cd)	EPA 6020	0.05	mg/kg
Chromium (Cr)	EPA 6020	0.05	mg/kg
Cobalt (Co)	EPA 6020	0.05	mg/kg
Copper (Cu)	EPA 6020	0.05	mg/kg
Iron (Fe)	EPA 6020	5	mg/kg
Lead (Pb)	EPA 6020	0.05	mg/kg

Table 2-2. RHMP Constituents to be Monitored and Corresponding Analytical Methods

Analyte	Method	Reporting Limit	Units
Manganese (Mn)	EPA 6020	0.05	mg/kg
Mercury (Hg)	EPA 245.7	0.02	mg/kg
Molybdenum (Mo)	EPA 6020	0.05	mg/kg
Nickel (Ni)	EPA 6020	0.05	mg/kg
Selenium (Se)	EPA 6020	0.05	mg/kg
Silver (Ag)	EPA 6020	0.05	mg/kg
Strontium (Sr)	EPA 6020	0.05	mg/kg
Thallium (Tl)	EPA 6020	0.05	mg/kg
Tin (Sn)	EPA 6020	0.05	mg/kg
Titanium (Ti)	EPA 6020	0.05	mg/kg
Vanadium (V)	EPA 6020	0.05	mg/kg
Zinc (Zn)	EPA 6020	0.05	mg/kg
Polynuclear Aromatic Hydrocarbons			
1-Methylnaphthalene	EPA 8270C	5	µg/kg
1-Methylphenanthrene	EPA 8270C	5	µg/kg
2,3,5-Trimethylnaphthalene	EPA 8270C	5	µg/kg
2,6-Dimethylnaphthalene	EPA 8270C	5	µg/kg
2-Methylnaphthalene	EPA 8270C	5	µg/kg
Acenaphthene	EPA 8270C	5	µg/kg
Acenaphthylene	EPA 8270C	5	µg/kg
Anthracene	EPA 8270C	5	µg/kg
Benz[a]anthracene	EPA 8270C	5	µg/kg
Benzo[a]pyrene	EPA 8270C	5	µg/kg
Benzo[b]fluoranthene	EPA 8270C	5	µg/kg
Benzo[e]pyrene	EPA 8270C	5	µg/kg
Benzo[g,h,i]perylene	EPA 8270C	5	µg/kg
Benzo[k]fluoranthene	EPA 8270C	5	µg/kg
Biphenyl	EPA 8270C	5	µg/kg
Chrysene	EPA 8270C	5	µg/kg
Dibenz[a,h]anthracene	EPA 8270C	5	µg/kg
Fluoranthene	EPA 8270C	5	µg/kg
Fluorene	EPA 8270C	5	µg/kg
Indeno[1,2,3-c,d]pyrene	EPA 8270C	5	µg/kg
Naphthalene	EPA 8270C	5	µg/kg
Perylene	EPA 8270C	5	µg/kg
Phenanthrene	EPA 8270C	5	µg/kg
Pyrene	EPA 8270C	5	µg/kg

2.2.2 Toxicity

Solid phase (SP) bioassays were performed to estimate the potential toxicity of the collected sediments to benthic organisms. The test used was the same one performed to analyze toxicity in sediments collected for the 2003 Regional Monitoring Program (Bight '03). The sediments were tested in a 10-day SP test using the marine amphipod *Eohaustorius estuarius*. On the day before test initiation a 2 cm aliquot of sample sediment was placed in a test chamber followed by prepared seawater. The samples were left overnight to allow establishment of equilibrium between the sediment and overlying water. On day one of the test, 20 amphipods were randomly placed in each of the test chambers. Any amphipods that did not bury in the sediment within 5-10 minutes were removed and replaced. Control sediment was used to determine the health of the amphipods; this was done by exposing the amphipods to clean sediment according to the same protocols used for the test sediments. Samples were monitored daily for the emergence of amphipods. At the end of the test, organisms were removed from the test chambers by sieving the sediment through a 0.5-mm mesh screen and the numbers of live and dead amphipods in each test chamber were recorded. The percent survival was calculated for the control and test sediments. The acceptability of the test was determined by evaluating the response of the control organisms. The test was considered acceptable if there was 90 percent mean control survival.

Standard procedure calls for measuring pore water ammonia in the sediments prior to test initiation to determine whether the concentration is within acceptable limits. If concentrations exceed 60 mg/L there is a potential that any observed toxicity may be due to high ammonia rather than some other contaminant. The pore water ammonia concentration in sample M2M-06 (94.0 mg/L) was found to be significantly elevated above recommended limit of 60 mg/L. Consequently, to reduce ammonia concentrations, sample material was acclimated by performing daily renewals with fresh seawater until pore water ammonia concentrations were reduced to 60 mg/L or below. An acclimated control was set up concurrently with sample M2M-06 to ensure that the acclimation procedure did not contribute to toxicity of *E. estuarius*. An un-acclimated control was also set up the day before test initiation to reflect normal test procedures.

A 96-hour reference toxicity test was also conducted concurrently with the sediment test to establish sensitivity of the test organisms used in the evaluation of the sediments. The reference toxicant test was performed using the reference substance, cadmium chloride, with concentrations of 2.5, 5, 10, 20, and 40 mg Cd^{2+}/L. Ten test organisms were added to each of these concentrations. The concentration that caused 50% mortality of the organisms (the median lethal concentration, or LC_{50}) was calculated from the data. The LC_{50} values were then compared to historical laboratory data for the test species with the reference substance. The results of this test were used in combination with the control mortality to assess the health of the test organisms.

An additional reference toxicant test was also conducted using ammonium chloride with target concentrations of 15.625, 31.25, 62.5, 125, and 250 mg NH_4/L to evaluate potential influence of ammonia toxicity on the test results of the sediments.

2.2.3 Infauna

The benthic samples were brought back from the field to the laboratory where they remained in a formalin solution for 7 days. The samples were then transferred from formalin to 70% ethanol for laboratory processing. The organisms were sorted using a dissecting microscope into five main taxonomic groups: polychaetes, crustaceans, molluscs, echinoderms, and miscellaneous minor phyla. While sorting, technicians kept a rough count for QA/QC purposes. Qualified taxonomists identified each organism and kept an actual count. The organisms were identified to the lowest possible taxon for each phylum.

A QA/QC procedure was performed on each of the sorted samples to ensure a 95% sorting efficiency. A 10% aliquot of a sample was re-sorted by a senior technician trained in the QA/QC procedure. The number of organisms found in the aliquot was divided by 10% and added to the total number found in the sample. The original total was divided by the new total to calculate the percent sorting efficiency. When the sorting efficiency of the sample was below 95%, the remainder of the sample (90%) was re-sorted.

2.2.4 Microbiology

Water samples were analyzed for the indicator bacteria, Enterococcus, using IDEXX Enterolert™ methodology. All results were reported to a most probable number value with a minimum reporting limit of <10 MPN/100mL and a maximum reporting limit of 24,196 MPN/100mL. All samples were delivered to the analytical laboratory within the 6 hour holding time requirement. Sample analysis was initiated immediately upon receipt.

2.2.5 Profile Data Processing

Sea-Bird profile scans were processed by Sea-Bird data processing software to convert recorded voltages to parametric values. Scans were averaged into one meter bins for analysis.

2.3 Data Analysis

A binomial model was selected to assess changes in sediment and water quality over time and to make statistically valid statements about present day conditions in the four water bodies that comprise the harbor monitoring program. In Phase I of this project, historical data were compiled to establish threshold levels and preset targets by which to measure changes in the harbors (Table 2-3). The majority of the data were from the 1998 Regional Monitoring Program (Bight 98') and the Bay Protection and Toxic Cleanup Program (BPTCP). Data that had similar detection limits (chemistry), test species (toxicity), and sampling equipment and screen size (benthic infauna) were used to determine a threshold level (Weston, 2005b).

The selection of which indicators were going to be monitored in the Pilot Program was based on whether there was sufficient historical data to create a threshold level. The threshold levels were established as concentration levels for chemical constituents, toxicity levels for bioassays, and diversity measures and the Benthic Response Index (BRI) for infauna (Smith et al., 2003). Preset targets were determined by defining the proportion of historical samples previously collected in these harbors that were below the established threshold levels. Preset target proportions then became the constant in the binomial model for comparison to newer data from the harbors. Proportions of samples collected in the Pilot Program were compared to the preset target in order to measure changes in the harbors. If there is a significantly greater proportion of current samples above the preset targets than it would indicate that water or sediment quality conditions were improving (Weston, 2005b). A summary of the established threshold levels and preset targets is presented in Table 2-4.

Indicators for the study were selected when there was sufficient historical data that could be used to compare to current data collected in the harbors. Primary indicators for the study were selected because they are either major constituents of concern (e.g., copper in water) or they provide information on a suite of measurements. Secondary indicators are used as supporting data to enhance the interpretation of the primary indicators (Weston, 2005b). The selection of individual primary and secondary indicators for water column chemistry, sediment chemistry, sediment toxicity, and benthic infauna will be further discussed in Sections 2.3.1 through 2.3.4.

Table 2-3. Studies used for establishing reference ambient values

Study Name	Year	Dana Point Harbor	Oceanside Harbor	Mission Bay	San Diego Bay
Sediment Chemistry					
America's Cup Harbor	2001				X
Bight 98	1998	X	X		X
BPTCP	1994, 1996	X	X	X	X
Central SD Bay Nav. Channel Deepening	1998, 2003				X
Chollas Creek	2003				X
10th Avenue Marine Terminal	2002				X
National City Wharf Extension	1999				X
Navy Arco	2000				X
Navy P-326	2000				X
Paleta Creek	2003				X
Reference reconnaissance	2003				X
Sediment sampling	2003	X			
Toxic Hot Spots Sediment	2003				X
Water and Sediment Testing Project	2001-2003			X	
Benthic Infauna					
Ambient Bay and Lagoon Monitoring	2003		X	X	
America's Cup Harbor	2002				X
Bight 98	1998	X	X		X
Reference reconnaissance	2003				X
Switzer Creek	2002				X
Sediment Toxicity					
Bight 98	1998				X
Benthic Infauna Analysis	2003-2004	X			
National City Wharf Extension	1999				X
Water and Sediment Testing Project	2001-2003			X	
Water Column Chemistry					
Baywide Copper	2002				X
Dana Point monitoring	1992-2002	X			
Paco Bay Water measurements	1992-1999				X

Table 2-4. Summary of Reference Ambient Values and Preset Targets

Measure	Reference Ambient Value	Preset Target
Primary Indicators		
Dissolved Copper (water)	4.8 µg/L	70%
Total Copper (water)	5.8 µg/L	26%
ER-M Quotient	0.2	48%
BRI	31	37%
E. estuarius mortality	20%	51%
Secondary Indicators		
Dissolved Zinc (water)	90 µg/L	100%
Total Zinc (water)	95 µg/L	97%
Dissolved Nickel (water)	74 µg/L	100%
Total Nickel (water)	75 µg/L	100%
Sediment Cadmium	1.2 mg/kg	90%
Sediment Chromium	81 mg/kg	78%
Sediment Copper	175 mg/kg	68%
Sediment Lead	46.7 mg/kg	74%
Sediment Nickel	20.9 mg/kg	80%
Sediment Zinc	150 mg/kg	45%
Sediment Total PAHs	4022 µg/kg	74%
Shannon-Wiener diversity	2	90%
Number of taxa	24	92%

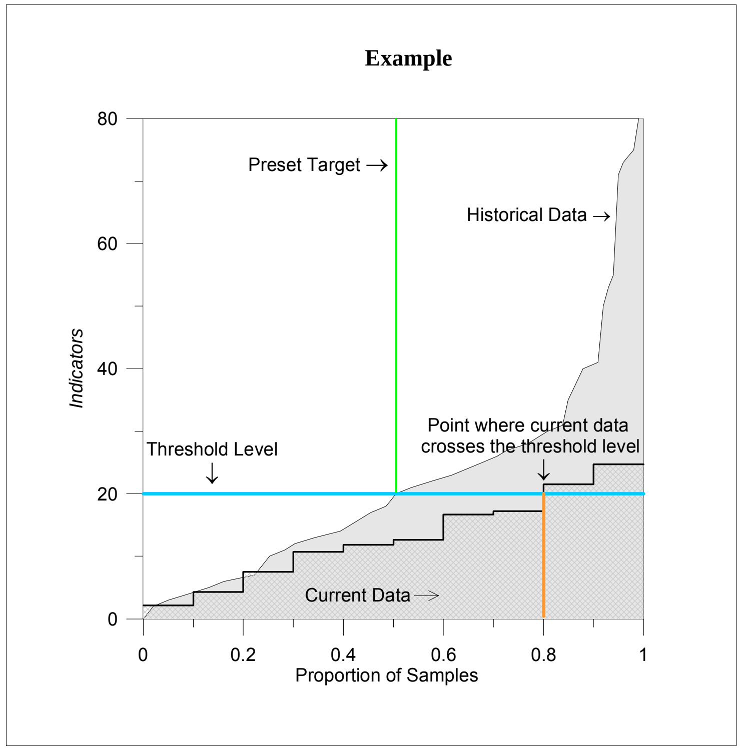

Figure 2-7. Example of a distribution curve that can be used as a reference guide

Each of the indicators measured in the Pilot Program were plotted for visual comparison to the threshold levels and preset targets. Figure 2-7 shows an example of a distribution curve that can be used as a reference guide. Both the historical and current data are plotted as distribution curves with the current data overlying the historical data. The current data is shown as a step plot rather than a smooth curve because there are only ten samples analyzed from each stratum compared to the larger amount of samples used from historical data. The horizontal blue line is the threshold level for each indicator. The vertical green line is the preset target. The orange line represents where the distribution curve for the 2006 data crosses the threshold level. When the orange line is to the left of the preset target then the proportion of samples that are below the threshold level is lower than the proportion of samples historically observed below this level. This would indicate that water or sediment quality conditions for that particular indicator are poorer than historically throughout the harbors. If the orange line is to the right of the preset target then the proportion of samples below the threshold level is greater than the proportion of samples historically observed below the threshold. This would indicate progress towards improved water or sediment quality in the harbors.

Results for each indicator were compared to the preset target to determine if they showed an increase in the proportion of samples below the threshold level. When the proportion of samples in the current year was higher than the preset target, the two proportions were compared to determine whether the increase was statistically significant. The null hypothesis was that the proportion of current samples below the threshold level was the same as the historical proportion of samples below the threshold level. The proportions were compared using methods described by Cohen (1977, Chapter 6). When the null hypothesis was rejected, it was determined that the current result is significantly greater than the preset value. This means that the current samples indicate an improved state when compared to historical data.

2.3.1 Water Column Chemistry

Historic observations of water column metal concentration were available for dissolved and total copper, nickel, and zinc (Weston, 2005b). This data along with benchmark values from the California Toxics Rule (CTR) and the California Ocean Plan (COP) were evaluated to establish threshold levels. The CTR was created using both literature and toxicity test data, thus making the CTR the best threshold level to use for aqueous metals (CTR, 2000). Only dissolved and total copper were selected as primary indicators for aqueous metals because of the large numbers of historical observations above the CTR. Dissolved and total zinc and nickel are used as secondary indicators. If the proportion of current samples below the threshold level is larger than the preset target it would indicate that water quality in the harbors was improving (Weston, 2005b). The threshold levels and preset targets for these metals are listed in Table 2-4.

2.3.2 Sediment Chemistry

For sediment chemistry, the mean ER-M quotient is the primary indicator for comparing results in the monitoring program to preset targets. Briefly, the effects range-low (ER-L) and effects range-median (ER-M) are two effects-based sediment quality values developed to help interpret sediment chemistry measurements and their potential for causing adverse biological effects (Long et al., 1995). These parameters were developed from an extensive database of sediment toxicity bioassays and chemistry measurements. The ER-L was calculated as the lower tenth percentile of the observed effects concentrations and the ER-M as the 50th percentile of observed effects concentrations.

The ER-M quotient, which is the ratio of sample concentration to the ER-M, can be used to evaluate the likelihood of benthic effects based on cumulative sediment chemistry. The quotient is derived by dividing each measured sediment chemical concentration by its respective ER-M. The mean ER-M quotient calculates an average quotient based on concentrations of all known contaminants relative to the ER-M values. Therefore, the ER-M quotient is a method of integrating the effects from multiple contaminants (Wenning et al., 2005). For the Pilot Program, the mean ER-M quotient was calculated using concentrations of arsenic, cadmium, chromium, copper, lead, mercury, nickel, silver, zinc, and total detectable PAHs.

Using historical data, the threshold level for the mean ER-M quotient was determined to be 0.2 based on recent projects with the San Diego RWQCB. Samples with ER-M quotients above 0.2 are more likely to have adverse benthic effects associated with the sediment chemistry. Based on historical data, the preset target for the ER-M quotient was established as 48%. If the proportion of current sediment samples with a mean ER-M quotient below 0.2 is significantly higher than 48%, then it would indicate that the overall conditions of sediment quality have improved. If the proportion of samples continues to be lower than the preset target over the course of the program then other indicators such as individual chemical constituents can be evaluated in conjunction with the ER-M quotient to help determine which contaminants are problematic in the harbors (Weston, 2005b).

Total PAHs and metals including cadmium, chromium, copper, lead, nickel, and zinc are used as secondary indicators for the Pilot Program. These measures will be used to help interpret the mean ER-M quotient by showing which of the parameters are predominant or changing in the mean ER-M quotient. For total PAHs and all of the metals except copper, the ER-L was determined to be the best threshold level. The threshold level for copper was based on the level at which anthropogenic origins may be contributing to the overall copper concentrations in the sediment. To determine this concentration, historical data were used to plot copper concentrations against iron concentrations, both of which are common in harbor sediments. Normalization to iron is a common approach to understanding the influence of potential enrichment via anthropogenic inputs. This is because iron is a reliable indicator of "geological background" levels. Trace metals such as copper that reliably scale with iron are generally viewed as being within geological background. From our historical data, at lower concentrations of copper there is a constant relationship with iron. Therefore, lower concentrations of copper that we observe may be natural and within geological background levels (Schiff and Weisberg, 1999). However, this relationship changes at a copper concentration of about 175 mg/kg as shown in Figure 2-8. At levels higher than this, copper scales differently. This is the basis for using 175 mg/kg as the threshold level for sediment copper, since above this point, the slope of the relationship changes markedly. A higher proportion of current samples below the threshold level compared to the preset target suggests that sediment quality conditions in the harbors are better than historically observed (Weston, 2005b). Table 2-4 shows the threshold levels and preset targets.

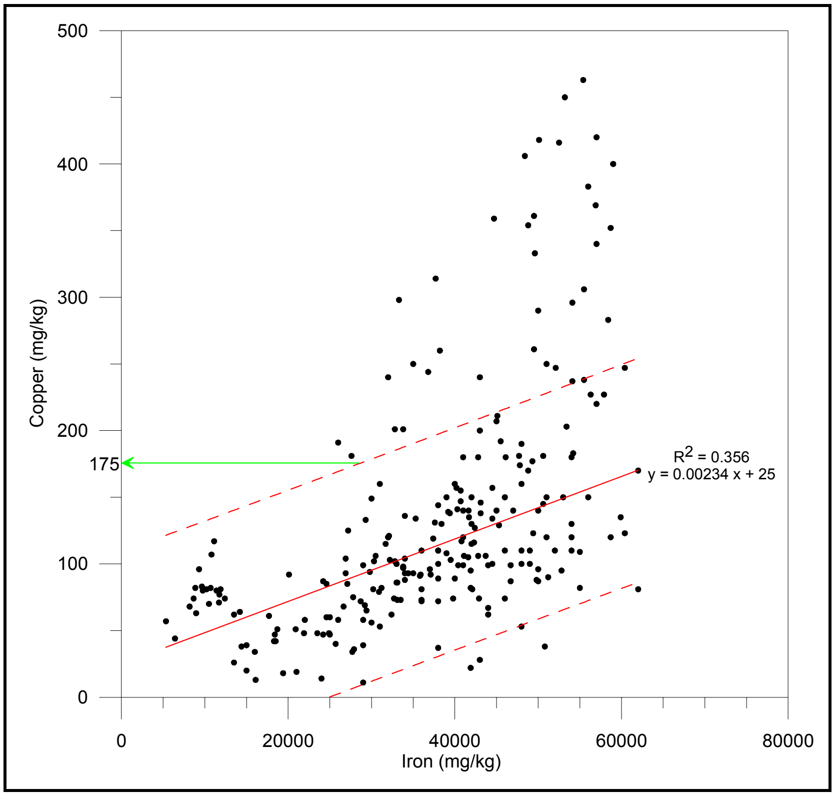

Figure 2-8. Relationship of copper to iron

2.3.3 Sediment Toxicity

Historical toxicity test results for *Eohaustorius estuarius* were used to establish the threshold levels for sediment toxicity. *Eohaustorius estuarius* was selected as a primary indicator of improving harbor conditions because of its relative sensitivity and the large amount of data that exist on this species that can be used for comparison. Mortality of the test species rather than survival was used for analysis for purposes of consistency with other indicators (i.e. higher numbers represent poorer conditions). Test results were adjusted for control survival prior to analysis of the data. For this primary indicator, the threshold level was set at 20% mortality; a value that is typically used as an indicator of non-toxic sediments. Conditions are considered to have improved if significant changes over the preset target were observed (i.e., more than 51% of samples show less than 20% mortality) (Weston, 2005b).

2.3.4 Benthic Infauna

Benthic infauna data from each of the harbors was assessed using various indices common to ecological community structure evaluations including Benthic Response Index (BRI), Shannon-Wiener diversity index, number of taxa, and abundance.

The BRI[1] is the primary indicator for evaluating infaunal changes in the harbors. The numerical criterion (i.e. community response levels) for this index is calculated by applying an abundance-weighted-average gradient that is correlated with sediment/habitat quality to the pollution tolerance of infaunal species. A reference threshold and four response levels help to characterize the degrees to which habitat conditions are deviating from reference conditions. Response level 1 is characterized as marginal deviation. Level 1 includes BRI values at which 5% of the reference species were lost. Response Levels 2-4 indicate increasingly disturbed benthic environments. Response level 2 is characterized as biodiversity loss which indicates a loss of 25% of reference species. Response level 3 is when there is a community function loss. BRI values at this level indicate a loss of 50% of reference species. Response level 4 is characterized by defaunation which indicates a loss of 80% of reference species (Ranasinghe et al., 2003). The range of BRI levels for each of these response levels is shown in Table 2-5.

Table 2-5. Characterization and BRI Ranges for Response Levels of Benthic Community Conditions

BRI Threshold	Level	Characterization	Definition
<31	Reference		
31-42	Response Level 1	Marginal deviation	>5% of reference species lost
42-53	Response Level 2	Biodiversity loss	>25% of reference species lost
53-73	Response Level 3	Community function loss	>50% of reference species lost
>73	Response Level 4	Defaunation	>80% of reference species lost

The BRI threshold level for the Pilot Program was set at 31 which is the currently established value for reference conditions in embayments. After applying this value to historical data, a preset target was determined to be 37%. When more than 37% of samples are below the threshold level of 31 then it would be indicative of a healthier benthic community compared to what was observed historically.

The Shannon-Wiener diversity and number of taxa are used as secondary indicators. Threshold levels were determined to be 2 and 24, respectively, for these indices. The preset targets were set at 90% for the Shannon-Wiener diversity and 92% for the number of taxa. For these indicators only, the proportion of samples above the threshold is the target of interest. When more than 90% (Shannon-Wiener diversity) or 92% (number of taxa) of the samples are above their respective threshold levels then it would also be indicative of a healthier benthic community than historically observed throughout the harbors (Weston, 2005b).

[1] The BRI used here is the first iteration of the index for enclosed bays. The index is currently under revision by SCCWRP. The new version is expected to be released later in 2007.

3.0 RESULTS

3.1 Water Quality

3.1.1 Chemistry

Surface water samples collected from marina and freshwater influenced stations in 2006 were analyzed for total and dissolved metals, hardness, dissolved organic carbon (DOC), total organic carbon (TOC), and PAHs. Concentrations of the indicator metals, DOC, TOC, total PAHs and physical characteristics are shown in Table 3-1. The results for all chemical analyses are provided in Appendix A.

Metals

Water samples from the marina stratum exceed the CTR for copper in each of the harbors. All of the sample concentrations from Dana Point Harbor and Oceanside Harbor are above the CTR for both dissolved and total copper (Table 3-1). In Mission Bay, one of the two stations has both dissolved and total copper exceeding CTR values. Two of the four marina stations in San Diego Bay have concentrations above the CTR for dissolved and total copper. No samples have concentrations that exceed the CTR for nickel or zinc.

Only two samples from freshwater influenced stations have copper concentrations that exceed the CTR. In Dana Point Harbor, one station has copper concentrations well above both the CTR for dissolved copper at 12.5 µg/L and for total copper with a concentration of 16.4 µg/L. One station in San Diego Bay has concentrations that exceed the CTR for dissolved copper with a concentration of 4.8 µg/L; no sample concentrations are above the CTR for total copper. Copper concentrations at both freshwater stations in Mission Bay are below the CTR concentration. Nickel and zinc concentrations are below their respective CTR at all stations.

Distribution curves for dissolved and total metals are presented in Figure 3-1 and Figure 3-2. The threshold level for each metal was based on the benchmark values from the CTR (Weston, 2005b). A lower percentage of marina samples collected in 2006 have concentrations of dissolved copper (30%) below the threshold level than the preset target of 70%, while the percentage of total copper samples (30%) in the marinas was above the preset target of 26%. All of the 2006 marina samples have concentrations of nickel and zinc, both dissolved and total, below the threshold level. None of these three metals show a significantly greater proportion of values below the threshold level in 2006 compared to historical data. However, the data do indicate that dissolved copper concentrations in the water column may be higher in the marinas than observed historically, while concentrations of zinc appear to be remaining at similar levels and concentrations of nickel are lower than observed historically.

Table 3-1. Surface water chemistry results for marina and freshwater influenced stations

Marina Water Chemistry Results - Surface															
Analyte	Units	MDL	RL	CTR	COP	Station Code									
						D106M	D206M	O406M	O506M	M206M	M306M	S106M	S206M	S306M	S406M
Dissolved Copper (Cu)	µg/L	0.01	0.02	4.8	25	**15.80**	**11.64**	**7.29**	**8.92**	4.41	**9.23**	**9.91**	3.32	**7.34**	3.32
Dissolved Nickel (Ni)	µg/L	0.005	0.01	74	50	0.52	0.55	0.47	0.45	0.40	0.39	0.48	0.81	0.59	0.55
Dissolved Zinc (Zn)	µg/L	0.005	0.01	90	189	42.06	34.85	19.25	22.33	15.36	35.08	19.88	9.27	19.94	9.88
Total Copper (Cu)	µg/L	0.01	0.02	5.8	30	**20.77**	**14.61**	**9.52**	**12.46**	5.53	**10.69**	**12.29**	4.38	**9.65**	5.12
Total Nickel (Ni)	µg/L	0.005	0.01	75	50	0.46	0.44	0.31	0.33	0.22	0.24	0.33	0.70	0.47	0.43
Total Zinc (Zn)	µg/L	0.005	0.01	95	200	52.12	36.84	21.68	25.15	15.58	30.98	21.95	10.05	21.71	11.19
Hardness as $CaCO_3$	mg/L	1.00	5.00	-	-	5620	5620	5490	5540	5540	5560	5720	5770	5640	5800
Total detectable PAHs	ng/L	1.00	5.00	-	-	3.20	0.00	6.20	3.20	5.90	16.00	43.60	5.10	167.80	72.50
TOC	mg/L	0.05	0.15	-	-	0.64	0.55	0.58	0.58	0.50	0.79	0.43	0.82	0.45	0.47
DOC	mg/L	0.05	0.15	-	-	0.46	0.42	0.46	0.48	0.47	0.50	0.47	0.79	0.52	0.50
Temperature	ºC	-	-	-	-	19.11	19.26	21.21	20.59	18.17	19.21	20.73	25.27	21.69	20.19
Salinity	PSU	-	-	-	-	33.42	33.39	33.29	33.29	33.47	33.50	33.54	34.97	33.74	33.64
Conductivity	µS/cm	-	-	-	-	45120	45232	47007	46403	44271	45316	46844	53313	48040	46447
pH	pH units	-	-	-	-	7.97	8.02	8.00	8.07	8.06	8.06	7.25	7.93	7.19	7.93
Dissolved Oxygen	mg/L	-	-	-	-	6.72	6.73	6.50	6.77	7.37	7.02	7.16	5.55	6.44	7.21
Transmissivity	%	-	-	-	-	62.77	63.82	75.15	44.72	85.18	88.16	74.77	68.52	47.55	56.94
Enterococci	MPN/100 mL	1.00	10.00	-	-	<10	<10	<10	<10	<10	<10	10	<10	<10	10

Freshwater Influenced Water Chemistry Results - Surface															
Analyte	Units	MDL	RL	CTR	COP	Station Code									
						D106F	M106F	M206F	S106F	S206F	S306F	S406F	S506F	S606F	S1106F
Dissolved Copper (Cu)	µg/L	0.01	0.02	4.8	25	**12.51**	0.90	1.31	2.67	2.52	**4.82**	3.19	2.25	3.29	2.91
Dissolved Nickel (Ni)	µg/L	0.005	0.01	74	50	0.96	0.48	0.47	0.81	0.75	1.90	0.74	0.97	0.72	0.63
Dissolved Zinc (Zn)	µg/L	0.005	0.01	90	189	38.14	4.20	2.98	5.53	6.25	4.94	8.39	3.32	10.99	6.84
Total Copper (Cu)	µg/L	0.01	0.02	5.8	30	**16.35**	1.22	2.10	3.70	3.31	2.91	3.30	2.97	3.68	2.79
Total Nickel (Ni)	µg/L	0.005	0.01	75	50	0.86	0.37	0.37	0.82	0.66	0.62	0.50	0.96	0.60	0.49
Total Zinc (Zn)	µg/L	0.005	0.01	95	200	39.15	3.82	4.29	6.23	7.68	6.69	7.43	5.52	12.45	6.59
Hardness as $CaCO_3$	mg/L	1.00	5.00	-	-	5600	5520	5790	5750	5820	5750	5910	6110	5780	5800
Total detectable PAHs	ng/L	1.00	5.00	-	-	0.00	9.20	8.90	19.70	22.90	19.80	38.60	7.90	20.70	50.40
TOC	mg/L	0.05	0.15	-	-	0.53	0.70	0.70	0.59	0.66	0.64	0.70	0.80	0.48	0.51
DOC	mg/L	0.05	0.15	-	-	0.44	0.67	0.72	0.62	0.67	0.71	0.62	0.89	0.59	0.59
Temperature	ºC	-	-	-	-	19.21	22.70	23.02	24.54	24.32	25.44	23.58	28.08	23.45	23.51
Salinity	PSU	-	-	-	-	33.38	33.35	33.68	34.48	34.44	34.62	34.11	35.60	34.08	34.11
Conductivity	µS/cm	-	-	-	-	45165	48543	49295	51903	51618	53019	50426	57172	50251	50350
pH	pH units	-	-	-	-	8.02	7.83	7.29	8.05	8.01	8.12	7.87	7.86	7.97	7.99
Dissolved Oxygen	mg/L	-	-	-	-	6.83	4.76	2.16	6.36	6.32	7.28	6.27	5.24	6.13	6.39
Transmissivity	%	-	-	-	-	63.49	74.38	62.54	80.22	79.43	77.89	83.43	41.78	87.42	83.18
Enterococci	MPN/100 mL	1.00	10.00	-	-	10	<10	<10	<10	20	<10	<10	10	<10	<10

Bold-Above CTR
MDL = Method Detection Limit
RL = Reporting Limit
CTR = California Toxics Rule
COP = California Ocean Plan

A high percentage of freshwater influenced samples have concentrations of total copper (90%) below the threshold level compared to the preset target of 26%. The percentage of dissolved copper samples with concentrations above the threshold was also above the preset target with 80% above compared to a target of 70%. All of the freshwater influenced samples have nickel and zinc concentrations below their respective threshold levels in both the dissolved and total fractions. In this stratum, the number of samples with concentrations of total copper below the threshold levels is significantly higher than the preset target. This indicates that concentrations of total copper are lower in freshwater influenced regions of the harbors than in the historical data. There were no statistically significant changes for any of the other indicator metals in the percentage of samples with concentrations below the threshold level compared to the preset targets. Data collected in 2006 also indicates that zinc concentrations in the harbors are remaining at similar levels as seen in the past and nickel concentrations are lower than in the historical data. More metals data needs to be collected from both the marina and freshwater stratum in order to determine whether these assessments are valid.

Dissolved and Total Organic Carbon

DOC and TOC analyses were conducted on all 20 of the surface water samples collected in 2006 (Table 3-1). Concentrations of DOC in the marina samples range from 0.42 mg/L to 0.52 mg/L, with one higher value of 0.79 mg/L in San Diego Bay. Marina TOC concentrations range from 0.43 mg/L to 0.82 mg/L.

Results of DOC concentrations in the freshwater influenced samples vary a bit more than those in marina samples, ranging from 0.44 mg/L in Dana Point Harbor to 0.89 mg/L in San Diego Bay. TOC concentrations in the freshwater influenced samples range from 0.48 mg/L to 0.80 mg/L.

The highest concentrations of DOC and TOC were found in the southern part of San Diego Bay for both strata.

Total Detectable PAHs

The results for total detectable PAHs in surface water samples are presented in Table 3-1. The detection limits for this constituent exceeded SWAMP requirements. For samples collected at marina stations, concentrations range from zero (non-detectable) in Dana Point Harbor to 167.8 ng/L in San Diego Bay. Concentrations of PAHs in freshwater influenced samples comparatively range from zero (non-detectable) in Dana Point Harbor to 50.4 ng/L in San Diego Bay. Results for individual PAH compounds are shown in Appendix A. Insufficient data were available to establish threshold values and preset targets for PAHs in water; PAH data are collected in the Pilot Program to begin establishing a baseline for future comparisons.

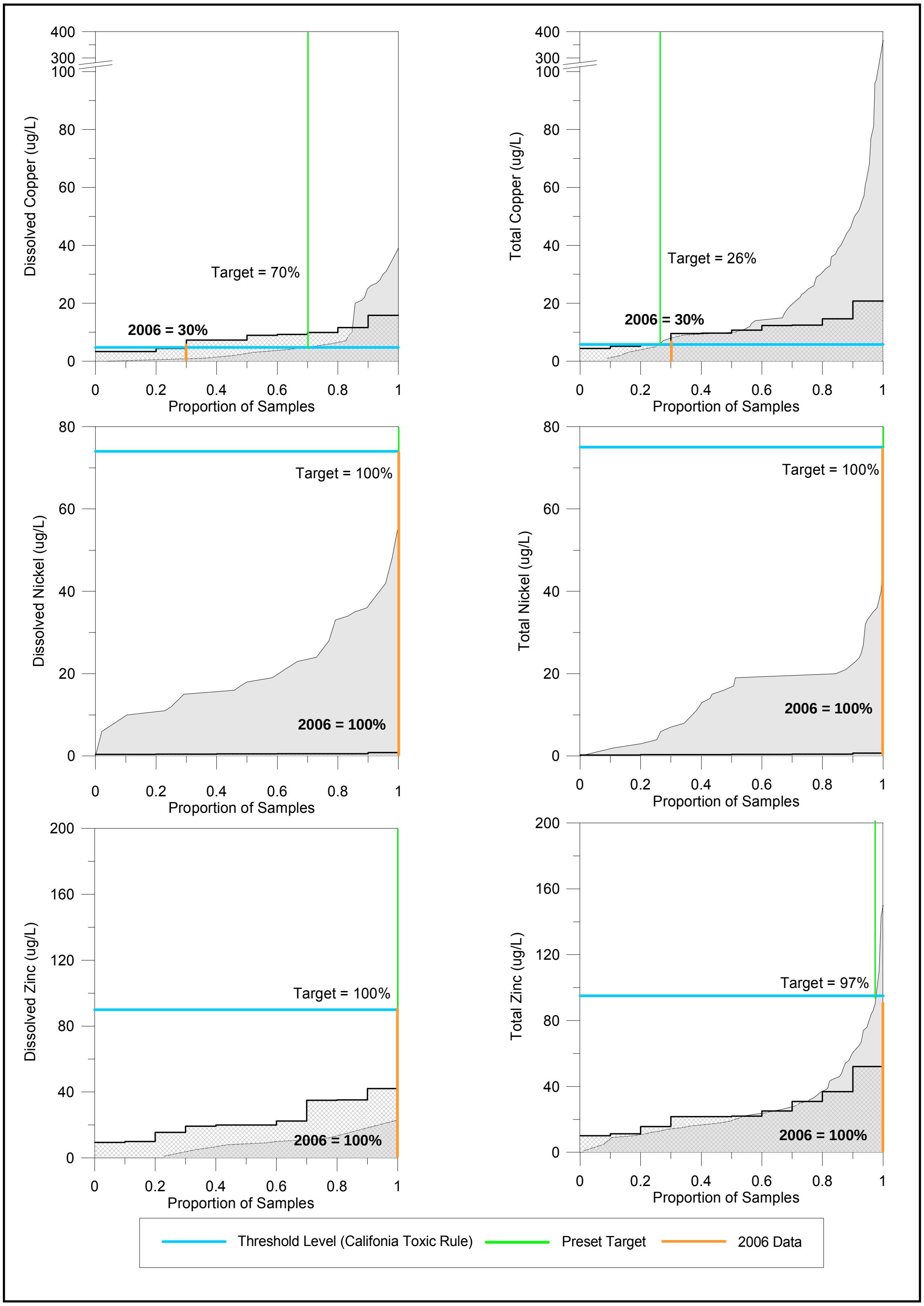

Figure 3-1. Distribution curves for dissolved and total water column metals in marina stations

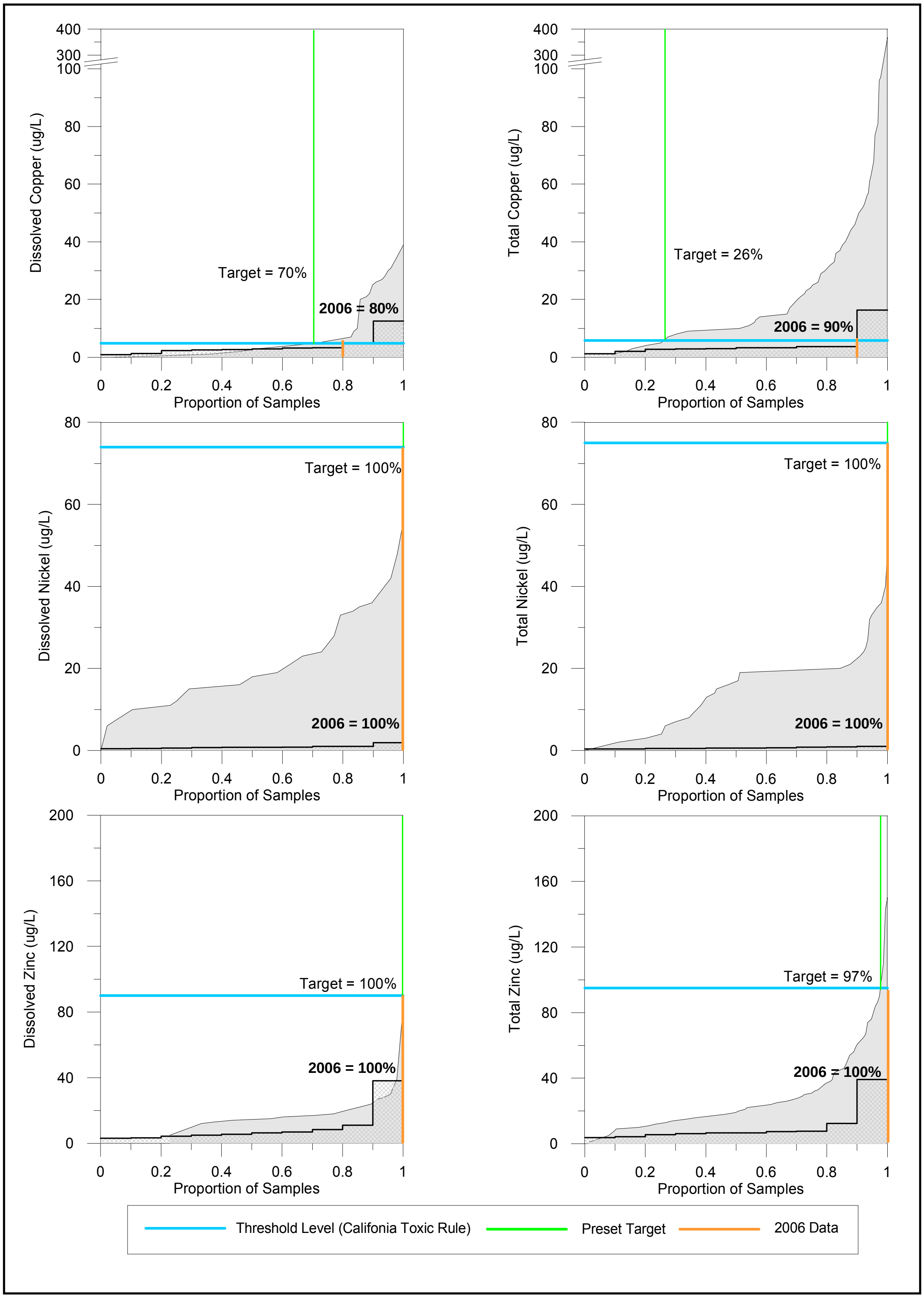

Figure 3-2. Distribution curves for dissolved and total water column metals in freshwater influenced stations

3.1.2 Bacteria

The results of the water analysis for the indicator bacteria, enterococcus, are presented in Table 3-1. None of the stations exceed the AB411 standards of 104 MPN/100mL. All stations have bacteria counts of 20 MPN/100mL or below. A threshold level and preset target have not been established for the indicator bacteria enterococcus, however tracking concentrations of enterococcus can help in determining whether the waters in the harbors are safe for body contact activities.

3.1.3 Water Column Profiles

Physical water column measurements for the 20 stations sampled in 2006 are presented in Figure 3-3 and Figure 3-4. Surface water data for individual monitoring stations are provided in Table 3-1. Summary data for all depths are provided in Appendix B. Measurements include temperature, salinity, pH, dissolved oxygen, and transmissivity. These measures, while not being compared to threshold levels, are useful to provide information about water quality that can help explain biological results and determine if the harbor waters can sustain a healthy biota.

Temperature

Maximum differences between surface and bottom temperatures for individual stations are less than 4 degrees Celsius (°C). Rapid changes of temperature with depth, indicative of a strong thermocline, are not evident at any of the sites. Surface temperatures in marina stations range from 18.2°C to 25.3°C and bottom temperatures range from 16.4°C to 25.3°C.

Water temperatures measured in freshwater influenced stations are similar to those in marina stations. The presence of a thermocline is not evident at any of the stations. Surface water temperatures range from 19.2°C to 28.1°C while bottom temperatures range from 18.2°C to 27.8°C.

Salinity and pH

Salinity values varied little between any of the marina stations with values for both surface and bottom waters ranging between approximately 28 to 36 psu; typical of saline environments. Surface water pH values range from 7.2 to 8.1 and vary little with depth.

Freshwater influenced stations have salinity and pH conditions similar to those observed at marina stations.

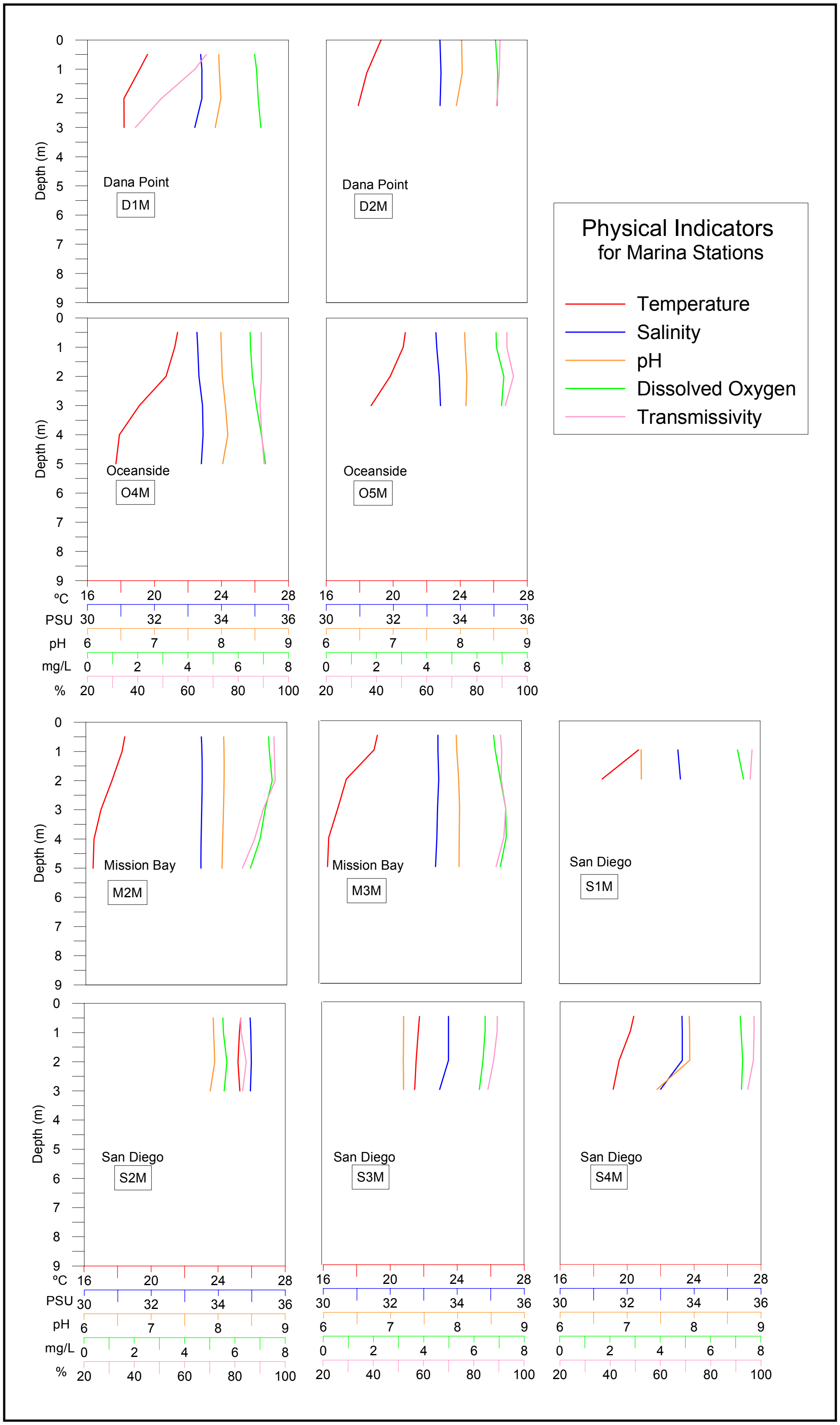

Figure 3-3. Physical indicators for marina stations

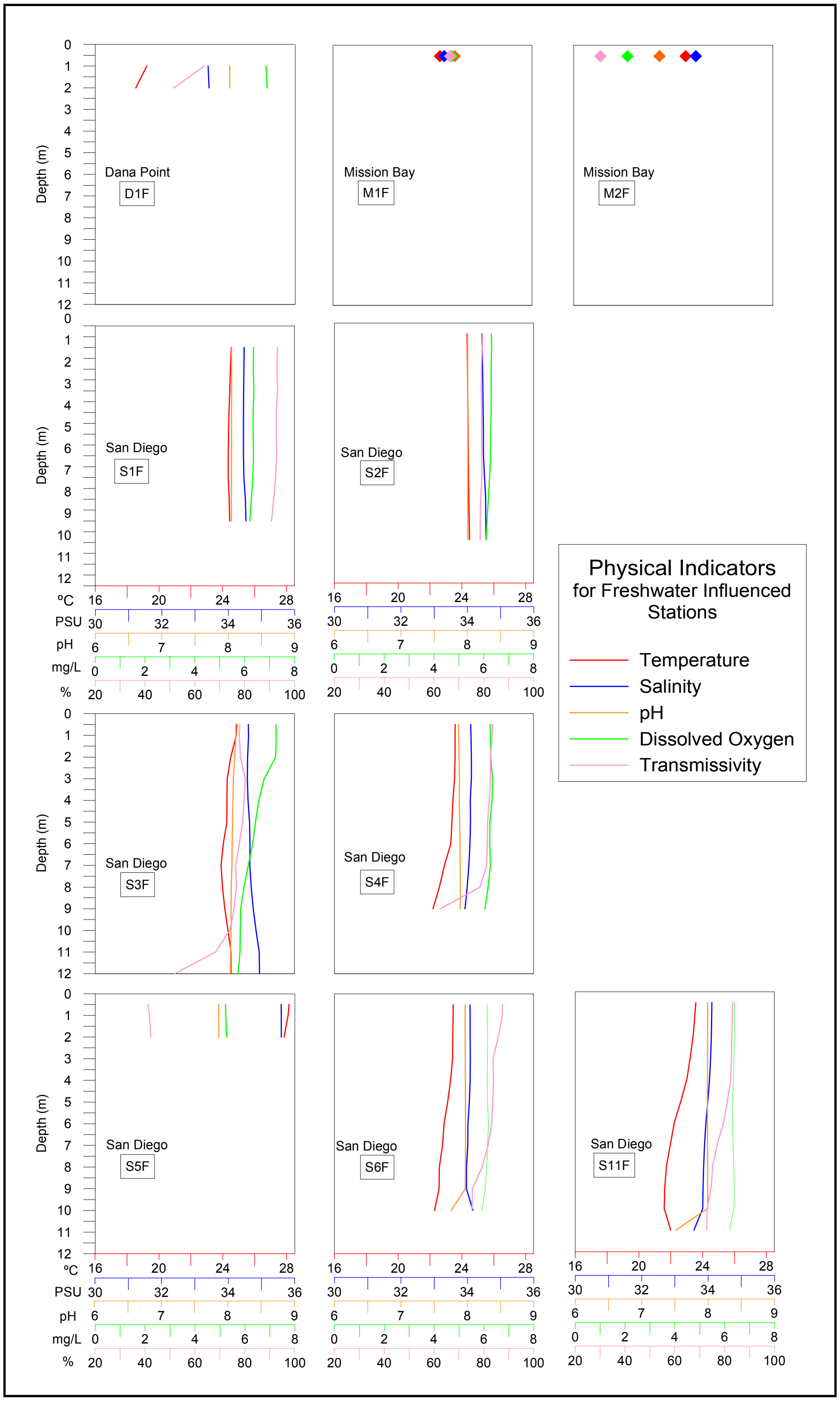

Figure 3-4. Physical indicators for freshwater influenced stations

Dissolved Oxygen
Dissolved oxygen concentrations in all of the marina stations are fairly low. Dissolved oxygen concentrations in the surface waters of marina stations range from 5.6 mg/L to 7.4 mg/L and in bottom waters range from 5.6 mg/L to 7.4 mg/L.

Dissolved oxygen concentrations in surface waters of freshwater influenced stations range from 2.2 mg/L to 7.3 mg/L while concentrations in bottom waters range from 5.3 mg/L to 6.8 mg/L. Maximum differences in concentrations between surface and bottom waters at these individual stations are 1.54 mg/L. Two shallow (~2m depth) stations in Mission Bay are the only stations with dissolved oxygen concentrations below 5 mg/L

Transmissivity
Surface water values for light transmittance for all of the marina stations range from 45% to 88%. Relatively lower transmissivity values are typically found in bottom waters compared to surface and mid-depth waters.

Freshwater influenced stations have transmissivity values in surface waters similar to those found in the marina stations ranging from 42% to 87%. Station S506F, located in South Bay, has the lowest transmissivity values compared to all of the other freshwater influenced stations.

3.2 Sediment Analysis

3.2.1 Chemistry

Mean ER-M Quotient
The mean ER-M quotient is one of the primary indicators for sediment quality for the Pilot Program. The mean ER-M quotient for marina stations ranges from 0.05 to 1.36 (Table 3-2) with 40% of the samples having an ER-M quotient below the threshold level of 0.2 (Figure 3-5). The freshwater influenced stations have mean ER-M quotients ranging from 0.10 to 0.32 with 60% of the samples below the threshold level. In comparison, the mean ER-M quotient for historical data from all strata in the harbors ranged from near zero to 16 with 48% of samples having an ER-M quotient below 0.2. This would indicate that conditions of sediment chemistry in the marinas are similar to those seen historically in the harbors while conditions in areas that have a freshwater influence are better than observed in the past; however, additional data need to be collected before making this assessment.

Metals
Six metals were identified as secondary indicators of sediment chemistry conditions. Results for samples collected in 2006 are shown in Table 3-2. The results for all of the total and dissolved metals that were analyzed are provided in Appendix A.

Table 3-2. Sediment chemistry results for marina and freshwater influenced stations

Marina Sediment Chemistry Results															
Analyte	Units	MDL	RL	ER-L	ER-M	Station Code									
						D106M	D206M	O406M	O506M	M206M	M306M	S106M	S206M	S306M	S406M
Cadmium (Cd)	mg/kg	0.025	0.05	1.2	9.6	0.20	0.20	0.50	0.20	0.10	0.40	<0.025	0.10	0.30	0.30
Chromium (Cr)	mg/kg	0.025	0.05	81	370	24.20	28.20	49.50	38.60	18.90	50.30	4.50	52.50	**88.30**	82.10
Copper (Cu)	mg/kg	0.025	0.05	34	270	**118.20**	**147.30**	**2721.00**	**199.10**	**121.70**	**272.10**	**34.40**	**112.50**	**225.40**	**539.20**
Lead (Pb)	mg/kg	0.025	0.05	46.7	218	9.44	14.35	21.13	15.99	16.64	34.14	4.66	23.10	45.35	**119.50**
Nickel (Ni)	mg/kg	0.025	0.05	20.9	51.6	10.40	11.90	19.40	13.80	5.70	14.90	1.20	14.20	12.00	15.60
Zinc (Zn)	mg/kg	0.025	0.05	150	410	109.70	128.60	**287.20**	**273.50**	114.60	**261.60**	37.10	**249.80**	**317.70**	**337.90**
Total Detectable PAHs	µg/kg	-	-	4022	44792	220.49	169.04	123.11	147.21	101.71	358.52	102.56	89.59	830.00	3975.32
ERM-Q	-	-	-	-	-	0.12	0.14	1.18	0.32	0.12	0.28	0.05	0.21	0.35	1.36
Gravel	%	-	-	-	-	0.98	0.45	0.64	0.08	0.06	0.00	0.00	0.01	1.08	0.00
Sand	%	-	-	-	-	58.53	61.56	25.57	33.15	75.43	12.80	95.55	15.63	28.65	8.42
Silt	%	-	-	-	-	22.97	24.27	45.87	45.71	10.14	52.13	2.59	33.26	26.76	52.72
Clay	%	-	-	-	-	17.53	13.71	27.92	21.06	14.37	35.07	1.86	51.11	43.51	38.86
Median	microns	-	-	-	-	116.51	106.97	14.84	29.41	179.07	10.47	191.68	3.56	6.87	8.18
Fines (silt + clay)	%	-	-	-	-	40.50	37.98	73.79	66.77	24.51	87.20	4.45	84.37	70.27	91.58
TOC	%	0.01	0.05	-	-	1.08	0.90	1.46	1.00	0.96	2.75	0.15	1.13	1.44	2.07
E. estuarius mortality	%	-	-	-	-	5.00	6.00	0.00	0.00	7.00	0.00	0.00	12.00	0.00	1.00

Freshwater Influenced Sediment Chemistry Results															
Analyte	Units	MDL	RL	ER-L	ER-M	Station Code									
						D106F	M106F	M206F	S106F	S206F	S306F	S406F	S506F	S606F	S1106F
Cadmium (Cd)	mg/kg	0.025	0.05	1.2	9.6	0.30	0.60	0.40	0.10	0.10	0.20	0.30	0.30	0.40	0.20
Chromium (Cr)	mg/kg	0.025	0.05	81	370	40.00	34.90	19.80	38.40	34.90	40.30	57.00	22.30	46.20	34.10
Copper (Cu)	mg/kg	0.025	0.05	34	270	**282.30**	**65.50**	27.40	**104.90**	**100.90**	**96.80**	**196.60**	**48.40**	**125.30**	**109.70**
Lead (Pb)	mg/kg	0.025	0.05	46.7	218	18.68	39.36	21.26	21.56	19.71	19.16	**61.27**	11.63	45.39	32.57
Nickel (Ni)	mg/kg	0.025	0.05	20.9	51.6	17.50	15.80	9.70	10.70	9.90	12.80	15.30	12.30	14.30	8.40
Zinc (Zn)	mg/kg	0.025	0.05	150	410	**234.90**	**237.40**	118.80	**152.60**	137.80	141.00	**295.40**	108.30	**236.30**	**193.20**
Total Detectable PAHs	µg/kg	-	-	4022	44792	160.41	134.30	147.59	182.14	179.31	176.56	2212.85	1935.75	1683.96	888.72
ERM-Q	-	-	-	-	-	0.25	0.19	0.10	0.16	0.15	0.15	0.32	0.11	0.25	0.20
Gravel	%	-	-	-	-	2.57	0.11	0.11	2.25	0.01	1.34	3.60	1.11	0.00	0.21
Sand	%	-	-	-	-	4.95	6.30	25.62	49.70	51.52	39.46	18.93	63.42	27.60	73.16
Silt	%	-	-	-	-	55.20	37.89	51.21	19.34	21.22	28.03	30.56	12.38	38.24	11.69
Clay	%	-	-	-	-	37.27	55.70	23.06	28.71	27.25	31.17	46.91	23.10	34.16	14.95
Median	microns	-	-	-	-	8.13	2.26	35.59	71.21	68.73	29.70	5.29	126.19	17.54	215.73
Fines (silt + clay)	%	-	-	-	-	92.47	93.59	74.27	48.05	48.47	59.20	77.47	35.48	72.40	26.64
TOC	%	0.01	0.05	-	-	1.60	3.70	2.33	1.26	1.04	0.94	2.26	1.25	1.74	0.96
E. estuarius mortality	%	-	-	-	-	16.00	19.00	0.00	6.00	0.00	5.00	23.00	22.00	2.00	0.00

Bold - Above ER-L
Shaded - Above ER-M
MDL = Method Detection Limit
RL = Reporting Limit
ER-L = Effects Range-Low
ER-M = Effects Range-Median

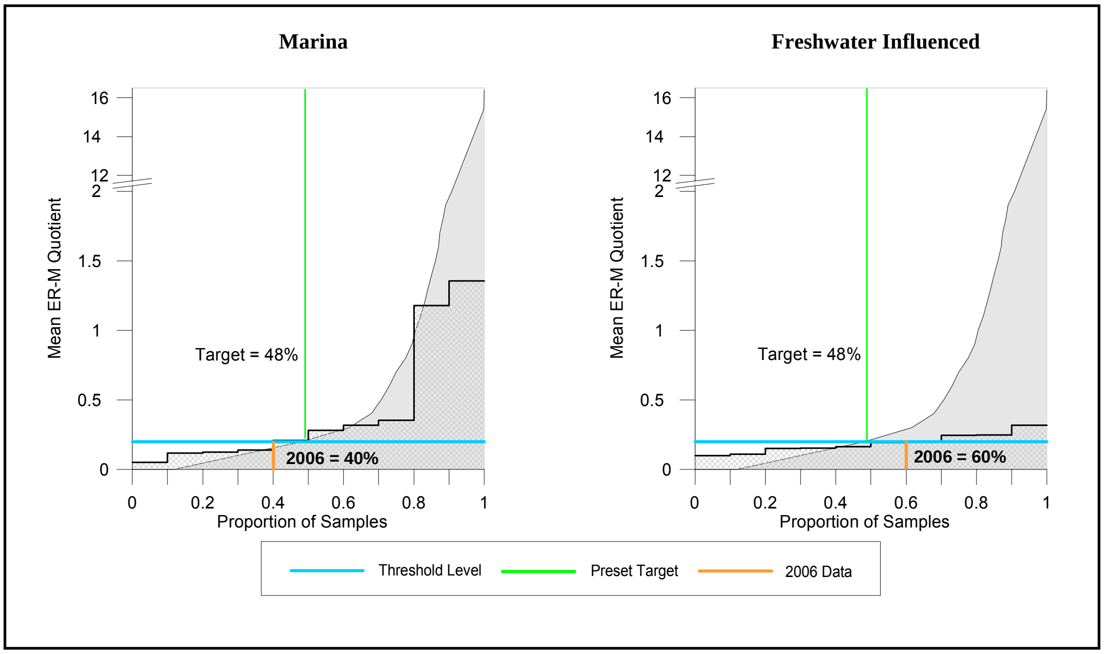

Figure 3-5. Distribution of the mean ER-M Quotient for marina and freshwater influenced sediments

Marina sediments in all four of the harbors have levels of copper that exceed the ER-L value of 34 mg/kg. Copper concentrations range from 34 mg/kg to 2721 mg/kg. Oceanside Harbor, Mission Bay, and San Diego Bay each have one station that exceeds the copper ER-M value of 270 mg/kg. Zinc exceeds the ER-L value of 150 mg/kg in all of the stations located in Oceanside Harbor, at one station in Mission Bay, and three stations in San Diego Bay; none of these concentrations exceeded the ER-M. Chromium exceeds the ER-L value of 81 at one station in San Diego Bay. Lead is also observed above the ER-L at one station in San Diego Bay. Cadmium and nickel concentrations are below their respective ER-L's in all of the harbors.

Freshwater influenced sediments at all of the stations, except one in Mission Bay, have copper concentrations that exceed the ER-L. Copper concentrations range from 27 mg/kg to 282 mg/kg. One freshwater influenced station in Dana Point Harbor has concentrations of copper above the ER-M value. Lead exceeds the ER-L value of 46.7 mg/kg at one station in San Diego Bay while zinc exceeds the ER-L in Dana Point Harbor, Mission Bay and San Diego Bay at various stations with concentrations ranging from 153 mg/kg to 295 mg/kg. Concentrations of cadmium, chromium, and nickel are below their respective ER-L's in all of the harbors.

Distribution curves for concentrations of cadmium, chromium, copper, lead, nickel, and zinc overlaid with historical data are shown in Figure 3-6 and Figure 3-7. A high proportion (90-100%) of marina samples have concentrations of cadmium, chromium, lead, and nickel below the established threshold level with each of these metals exceeding their respective preset targets. Metals that show a significantly greater proportion of samples below the threshold level than the preset targets are cadmium, lead, and nickel. This would indicate that current concentrations of these three metals in marinas are better than historic conditions. In contrast, the percentage of marina samples with copper and zinc below the threshold level are 50% and 40%, respectively, compared to the preset targets of 68% (copper) and 45% (zinc). This indicates that concentrations of copper and zinc in 2006 are worse than those observed historically.

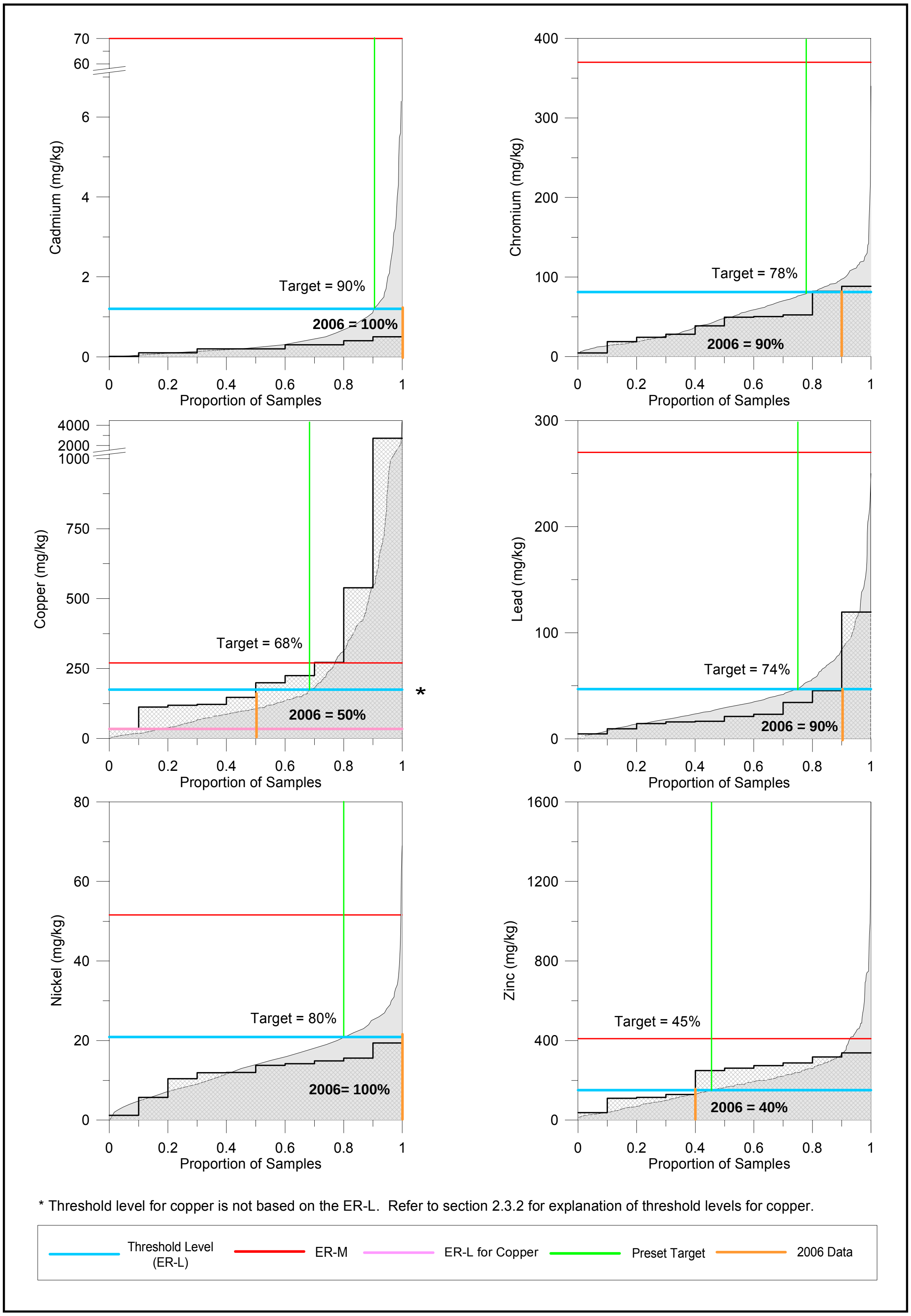

Figure 3-6. Distribution curves for metal concentrations in marina sediments

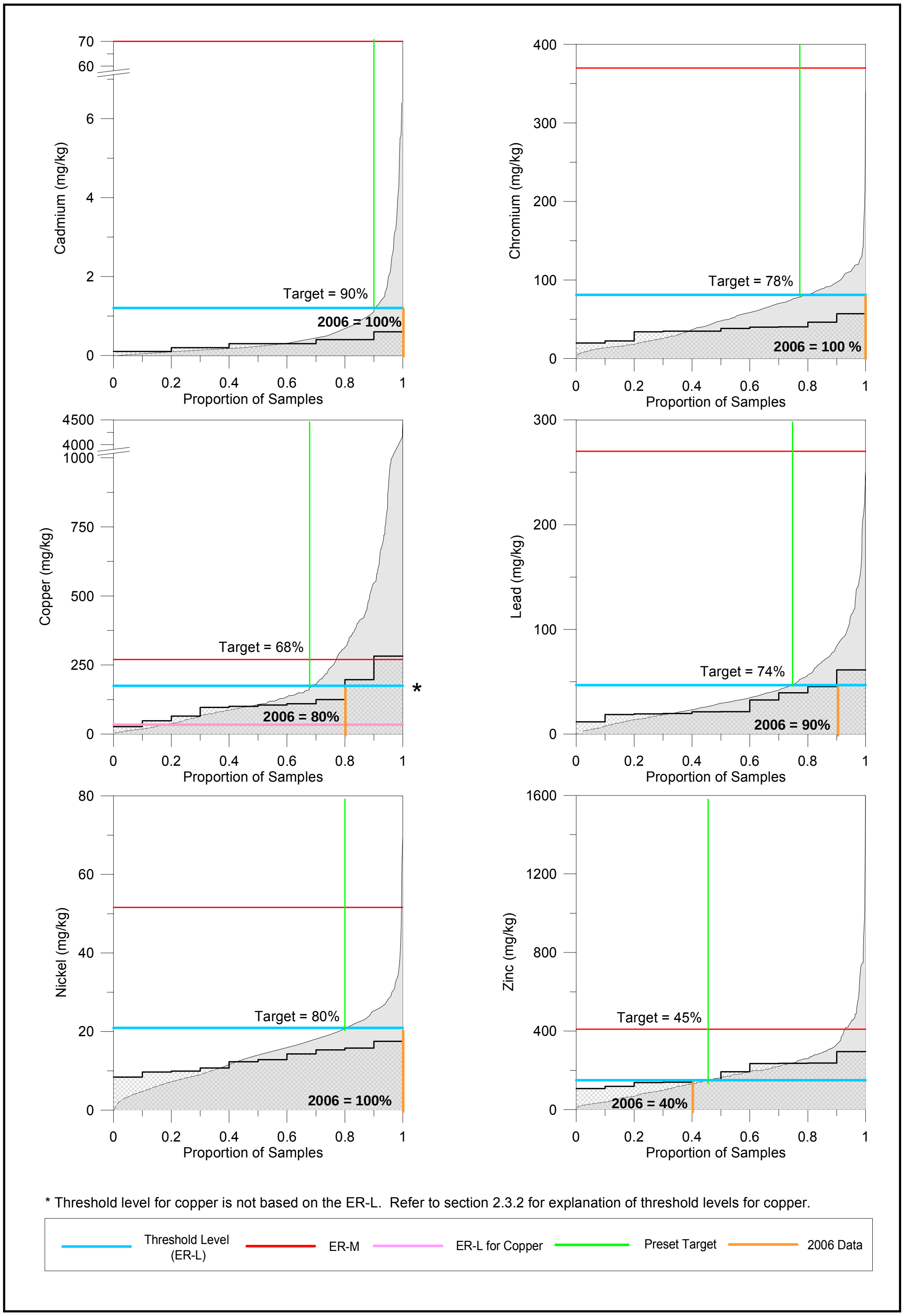

Figure 3-7. Distribution curves for metal concentrations in freshwater influenced sediments

A high proportion (80-100%) of freshwater influenced samples have concentrations of cadmium, chromium, copper, lead, and nickel below the established threshold level. All five of these metals exceeded their respective preset targets. Cadmium, chromium, lead, and nickel had a significantly greater proportion of samples below the threshold level than the preset targets. This would indicate that concentrations of these five metals in freshwater influenced sediments are better than those observed historically. For zinc concentrations, only 40% of the samples from 2006 were below the threshold level compared to the preset target of 45%. This would indicate that zinc concentrations in freshwater influenced regions of the harbor are about the same as historically observed; however, more metals data needs to be collected from both the marina and freshwater stratum in order to determine whether these assessments are valid.

Total Detectable PAHs

The results for total detectable PAHs are also presented in Table 3-2. Concentrations of PAHs in marina sediments range from 89.6 μg/kg to 3975 μg/kg, with the maximum value observed in San Diego Bay. Concentrations of PAHs in freshwater sediments range from 134 μg/kg in Mission Bay to 2213 μg/kg in San Diego Bay. None of the PAH values exceed the ER-L value of 4022 μg/kg.

Distribution curves for concentrations of total detectable PAHs with their ER-L and ER-M levels are shown in Figure 3-8. All of the marina and freshwater influenced samples collected in 2006 are below the threshold level of 4022 μg/kg compared to 74% of historical samples. There is a significantly greater difference between the results of the samples collected in 2006 and the preset target. This would indicate that PAH levels are better than historic conditions; however, more data needs to be collected in order to determine whether these assessments are valid.

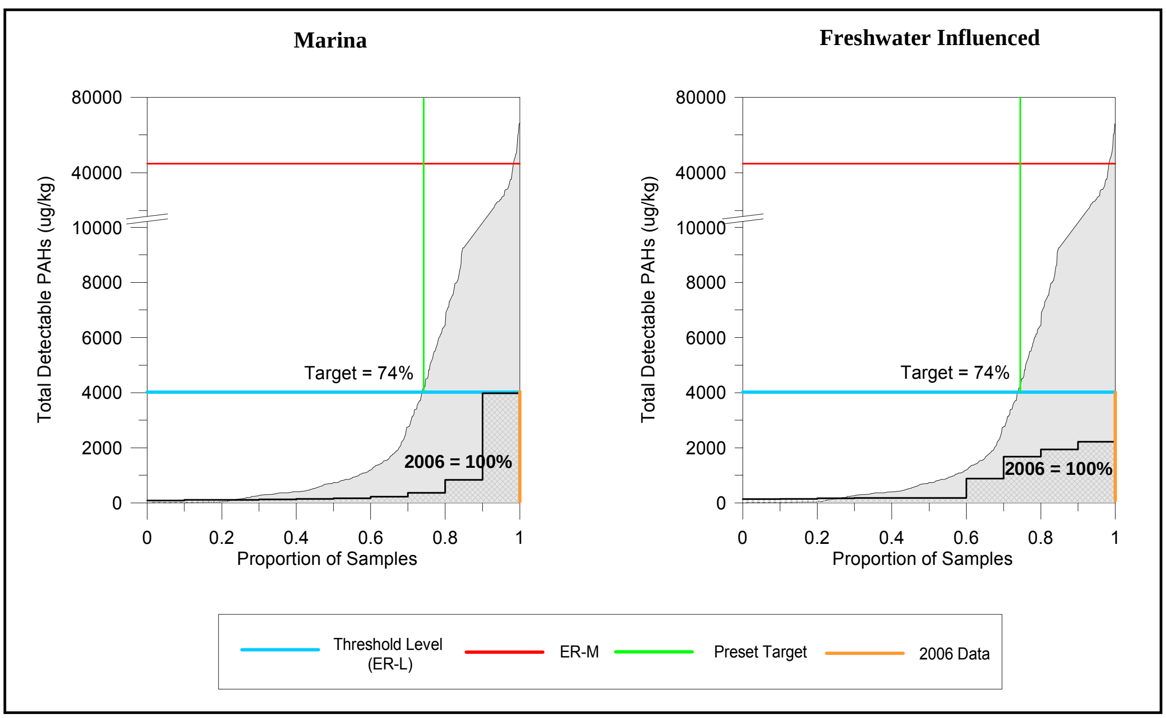

Figure 3-8. Distribution curves for total detectable PAHs in marina and freshwater sediments

3.2.2 Toxicity

The results of the sediment toxicity test conducted with both the marina and freshwater influenced sediments are presented in Table 3-2. Additional supporting data are provided in Appendix C. The control adjusted percent mortality for *E. estuarius* in the marina sediments ranges from 0% to 12%. The percent mortality for freshwater sediments ranges from 0% to 23%.

Distribution curves for *E. estuarius* are presented in Figure 3-9. The threshold level used for this toxicity test is 20% mortality. Historical data show that 51% of samples had less than 20% mortality. Current results show that 100% of marina samples and 80% of freshwater samples are below the threshold level. The proportion of both marina and freshwater influenced samples with mortality below 20% is significantly greater than the preset target of 51%. This would suggest sediment conditions in the harbors may be better in both of these strata than historically. More data needs to be collected in order to validate this assessment.

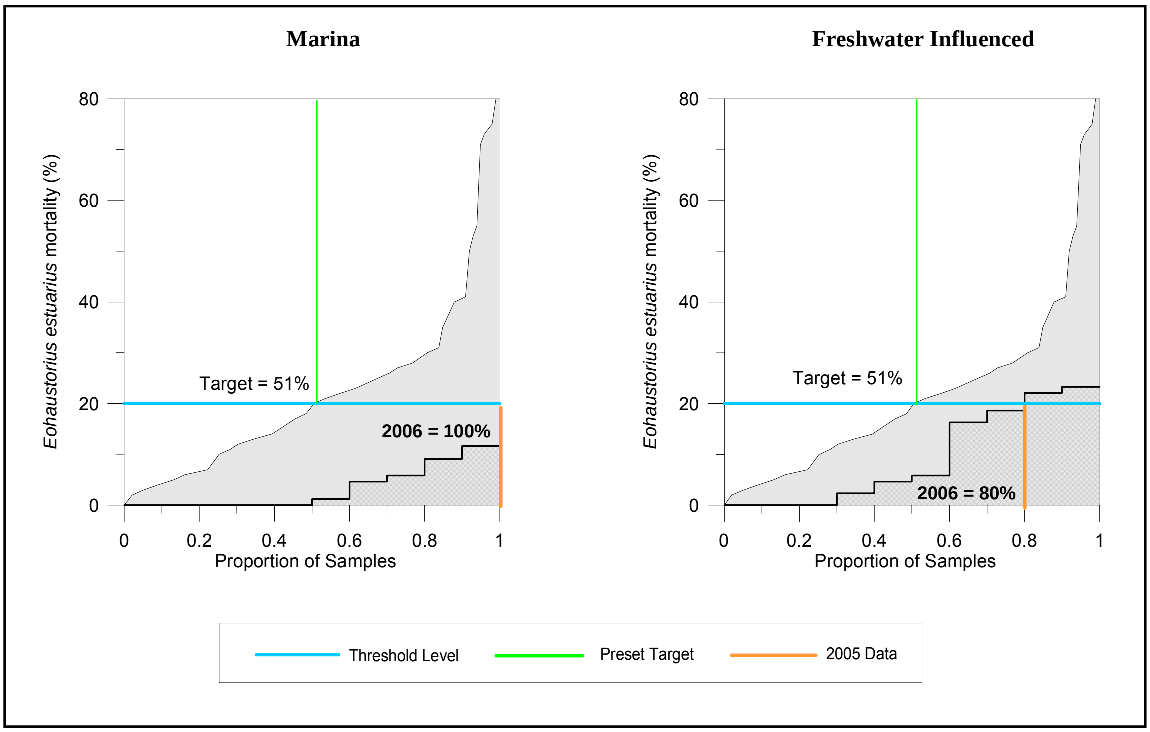

Figure 3-9. Distribution curves for toxicity in marina and freshwater influenced sediments

3.2.3 Benthic Infauna

Benthic infaunal samples were collected and analyzed in the four harbors from 10 stations in two strata: marina and freshwater influenced. The number of taxa, abundance, Shannon-Wiener diversity, and BRI were calculated for each of the samples (Table 3-3). Species names and abundances for each taxon are provided in Appendix D. The number of unique taxa identified in marina samples range from 20 to 54 taxa; total abundance ranges from 89 organisms to 2043 organisms per 0.1 m^2. Abundance is somewhat related to the number of taxa with higher numbers of organisms found at stations with greater diversity. The Shannon-Wiener diversity value for marina samples ranges from 1.83 to 3.14. BRI scores vary widely within and among the harbors ranging from 30 to 50.

Table 3-3. Results of benthic infauna community measures for marina and freshwater influenced sediments

Marina Sediments										
Station	D106M	D206M	O406M	O506M	M206M	M306M	S106M	S206M	S306M	S406M
Number of Taxa	20	27	30	29	54	46	24	28	29	22
Total Count	516	476	89	218	2043	199	158	183	278	101
Shannon-Wiener Diversity Index	2.22	1.99	3.14	2.43	1.83	3.07	2.22	2.52	2.35	2.55
BRI Score[1]	50	42	36	36	37	31	30	44	49	30

Freshwater Influenced Sediment										
Station	D106F	M106F	M206F	S106F	S206F	S306F	S406F	S506F	S606F	S1106F
Number of Taxa	13	42	35	45	26	23	24	39	45	29
Total Count	107	4297	2348	806	109	100	140	4552	798	186
Shannon-Wiener Diversity Index	1.96	1.65	1.67	2.85	2.79	2.42	2.57	1.21	1.91	2.65
BRI Score[1]	49	50	47	47	37	48	52	39	43	38

[1] The BRI used here is the first iteration of the index for enclosed bays. The index is currently under revision by SCCWRP. The revised index is expected to be released later in 2007.

The number of taxa collected at freshwater influenced stations also varied, ranging from 13 to 45; total abundance ranges from 100 organisms to 4552 organisms. The Shannon-Wiener diversity value for freshwater influenced stations ranges from 1.21 to 2.85. Low Shannon-Wiener diversity values suggest the influence of large abundances of a few taxa, particularly at station S506F and M106F. Station S506F was largely dominated by the mollusk species *Musculista senhousei*, and M106F by the gastropod, *Barleeia sp*. The BRI scores are similar in range to those observed at marina stations.

Distribution curves for BRI values calculated from the 2006 data for both marina and freshwater influenced sediments appear in Figure 3-10. The threshold level for the BRI was set at 31, the current level designating reference conditions in embayments. Samples that have BRI values below 31 are considered to be indicative of a healthy benthic community. Only 30% of the marina samples have BRI values below the threshold level in comparison to the preset target of 37%. The proportion of freshwater samples with BRI scores below the threshold level is even lower at zero percent.

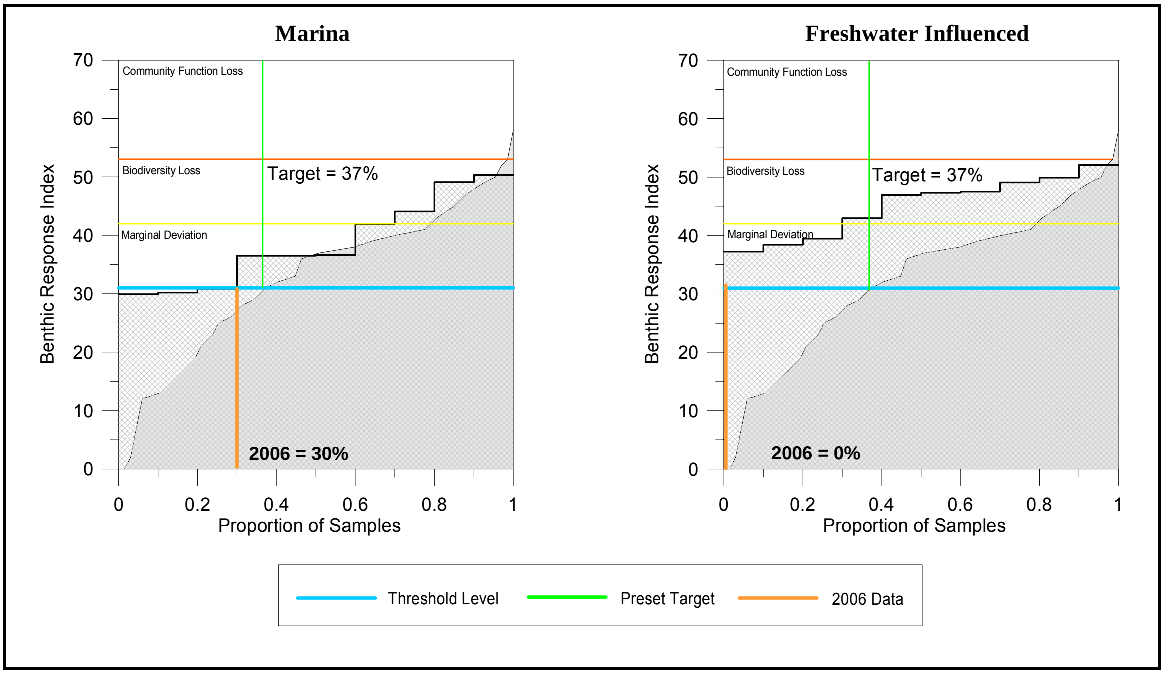

Figure 3-10. Distribution curves for the Benthic Response Index (BRI) for marina and freshwater influenced sediments.

The Shannon-Wiener diversity and number of taxa were used as secondary indicators. The curves for these indicators determined from the 2006 data are presented in Figure 3-11. For these indicators only, the comparison is the proportion of samples with values above the threshold level. Values that are above the threshold level of 2 for Shannon-Wiener diversity and 24 for number of taxa are considered to be an indication of healthier benthic communities. The preset target for the Shannon-Wiener diversity is 90% and for the number of taxa it is 92%. In comparison, the proportion of marina samples collected in 2006 that have Shannon-Wiener diversity and number of taxa above the threshold level are 80% and 70%, respectively. For sediments collected from freshwater influenced stations, 50% of the samples have a Shannon-Wiener diversity and 70% have a number of taxa above the threshold level. No benthic infaunal indicator shows proportions that are significantly greater than the preset targets. These results may imply that benthic community assemblages in both the marina and freshwater influenced stratum may be worse when compared to historic conditions; however, further data will need to be collected over time to determine if these evaluations are robust.

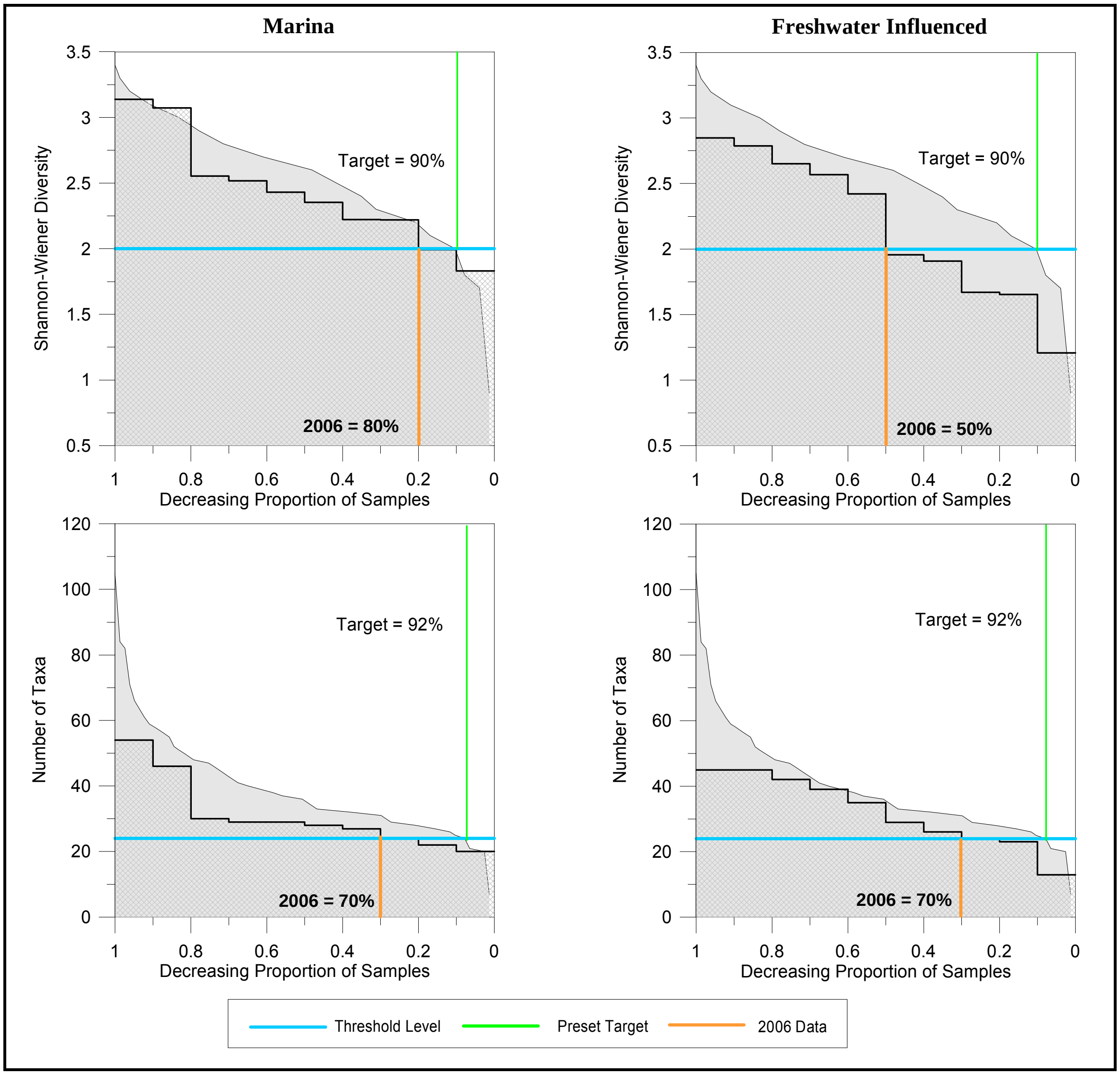

Figure 3-11. Distribution curves for benthic infauna community measures for marina and freshwater influenced sediments

3.2.4 Grain Size and TOC

Sediment grain size and TOC for the 20 stations sampled in 2006 are summarized in Table 3-2. These measurements have no threshold levels for comparison; however, they can be used to help interpret biological responses. Marina samples collected in Dana Point Harbor mainly consist of sand particles while those collected in Oceanside Harbor are largely dominated by fine-grained sediments (particle diameter < 63 microns). Samples from both of these harbors have relatively low TOC values ranging from 0.9% to 1.5%. Marina samples collected in Mission Bay and San Diego Bay vary in sediment size. In Mission Bay, two samples were collected in Quivera Basin one mainly dominated by sand (75%) and the other by fine-grained particles (87%). In San Diego Bay, sand is the dominant sediment constituent in Shelter Island Yacht Basin (S106M) while the other three marina samples, collected from America's Cup Harbor, Sunroad Resort Marina, and the Coronado Cays, consisted mainly of fine-grained particles ranging from 70% to 92%. TOC values in Mission Bay and San Diego Bay range from 0.96% to 2.8%.

Particle size varied within sediments collected from freshwater influenced stations in San Diego Bay. Three samples were collected at the mouth of Sweetwater River. Station S306F is dominated by fine-grained particles (59%) while stations S106F and S206F are approximately an equal mix of both sand and fine grains. Samples collected at the mouth of Chollas Creek are mainly fine-grained except at station S1106F which is dominated by sand (73%). Station S506F, located in South Bay, consisted mainly of sand (63%). Freshwater influenced stations in Dana Point Harbor and Mission Bay are mostly dominated by fine grains. TOC values for these stations range from 0.94% at station S306F to 3.7% at station M106F. There were no samples collected from freshwater influenced stations in Oceanside Harbor.

4.0 DISCUSSION

The objective of the Pilot Program is to implement the RHMP core monitoring on a limited scale to verify the study design. Data are to be statistically evaluated to establish the frequency of core monitoring needed to assess trends in water and sediment quality. Comparison of the pilot project data to historical data used in setting the threshold levels and target percentages is not a direct comparison because the historical data were collected throughout the harbors and include data from potentially cleaner sediments in the more open parts of the harbors. The comparisons made in this report are made to verify elements of the study design and not to make conclusions on the conditions in the harbors.

Water Column Measures

Sample results for the 2006 survey show differing results between the marina and freshwater influenced strata when compared to the preset targets for the primary indicators, dissolved and total copper (Table 4-1). In the marina stratum, both indicators have a percentage of samples above the threshold level that were either lower or similar to the preset targets. In the freshwater stratum, both indicators have a higher percentage of samples above the threshold level compared to the preset targets, but only total copper was significantly higher. This suggests that conditions in the freshwater influenced area are better than what has been observed historically throughout the harbors; conversely, conditions in the marinas are worse than those historically observed. For the secondary indicators nickel and zinc, both dissolved and total, all stations in both strata had concentrations below the threshold levels. This is consistent with the preset targets. Other measured metals and PAHs are all at concentrations below their respective CTRs. Therefore, the potentially degraded conditions in the water column of the marina stratum may be limited to copper contamination, a documented contaminant in San Diego Bay marinas (McPherson and Peters, 1995; SDRWQCB, 2005) as well as others in the San Diego region (Schiff et al., 2006).

Sediment Measures

The mean ER-M quotient is the primary indicator for sediment quality in the RHMP. Comparison of the percent of samples below a mean quotient of 0.2 shows that the marina strata has only 40% of the samples below 0.2 compared to the preset target of 48%, while the freshwater influenced strata has a slightly higher percentage of 60% (Table 4-1). This may indicate that conditions in the marinas are worse, when compared to the historical conditions of the harbors. In contrast, conditions of areas of the harbor with a freshwater influence may be better than observed in the past. This finding triggers examination of the secondary indicators for sediment chemistry, metals and total PAHs. As can be seen in Table 4-1, copper results support the mean ER-M quotient conclusion. However, comparison of the percent of samples below the threshold levels for other metals and PAHs tend to contradict the conclusion based on the mean ER-M quotient. Cadmium, chromium, lead, nickel, and total PAHs have equal or higher percentages of samples below the threshold levels than the target percentage in both strata. Because the mean ER-M quotient is comprised of more than just these secondary indicators, other measured constituents were reviewed to see their contribution to the mean ER-M quotient (see Appendix A, Tables A-3 and A-4). Copper and zinc are contributors as are silver and mercury in the stations with the higher mean ER-M quotients in the marina stratum, while silver is also influential in the freshwater stratum. These metals were not identified as secondary indicators because they were not consistently available in the historical data. Bringing this information into the picture supports the use of the mean ER-M quotient as the primary indicator of sediment chemical quality.

Table 4-1. Comparison to reference ambient values

Measure	Reference Ambient Value	Preset Target	Marina	Freshwater Influenced
Primary Indicators				
Dissolved Copper (water)	4.8 μg/L	70%	30%	80%
Total Copper (water)	5.8 μg/L	26%	30%	90%*
ER-M Quotient	0.2	48%	40%	60%
E. estuarius mortality	20%	51%	100%*	80%*
BRI	31	37%	30%	0%
Secondary Indicators				
Dissolved Zinc (water)	90 μg/L	100%	100%	100%
Total Zinc (water)	95 μg/L	97%	100%	100%
Dissolved Nickel (water)	74 μg/L	100%	100%	100%
Total Nickel (water)	75 μg/L	100%	100%	100%
Sediment Cadmium	1.2 mg/kg	90%	100%*	100%*
Sediment Chromium	81 mg/kg	78%	90%	100%*
Sediment Copper	175 mg/kg	68%	50%	80%
Sediment Lead	46.7 mg/kg	74%	90%*	90%*
Sediment Nickel	20.9 mg/kg	80%	100%*	100%*
Sediment Zinc	150 mg/kg	45%	40%	40%
Sediment Total PAHs	4022 μg/kg	74%	100%*	100%*
Shannon-Wiener diversity	2	90%	80%	50%
Number of taxa	24	92%	70%	70%

* Indicates results significantly higher than preset target

E. estuarius toxicity in sediments, measured by the percent of samples with mortality less than 20%, is significantly above the preset target in both marina and freshwater influenced sediments. This may imply that harbor sediments are less toxic than historical data suggest.

The primary benthic infaunal indicator, the BRI, suggests that both strata may be in poorer condition than observed throughout the harbors historically. This is corroborated by the data for the secondary indicators (number of taxa and Shannon-Wiener diversity index). The data for these two indicators show that the proportion of samples above the threshold levels are less than the preset targets (for these indicators the measurement of interest is the proportion of samples above the threshold).

Thus, the three sets of indicator constituents for sediments show differing results. The mean ER-M quotient suggests conditions are about the same as historical data suggest in both strata although copper in marinas appears to continue to be a problem. The toxicity results suggest healthier conditions; and the benthic infauna results indicate somewhat poorer conditions than have been observed historically. The two strata have similar results in each of the sets of indicator data. These observations are only indicative of conditions sampled in 2006 and are not intended to make definitive statements about the health of the harbors. Such incongruous results are likely a function of having a high degree of variability in measurements that are represented by very few samples. While it was never intended for individual years within the Pilot Project to answer such questions, more data would aid in drawing confidently conclusions. A multi-year analysis, such as after the third year of data collection in the Pilot Project, will likely reveal much more robust conclusions.

Relationship Between Water and Sediment Chemistry

Results for common constituents measured in water and sediment are similar when compared to their respective threshold levels. The proportions of marina stations with copper concentrations below the threshold level are under the preset target for both water and sediment samples. In the freshwater influenced stratum, copper and nickel are above the preset target in both water and sediment samples.

Variability in Primary and Secondary Indicators

The marina and freshwater influenced strata were selected for the Pilot Program because they were expected to have more variability in the results than the open water strata. Table 4-2 shows the variability in the primary and secondary indicators. The coefficient of variation (CV = ratio of standard deviation to mean expressed as percent, a smaller CV means lower variability in the data) is used for comparison of variability over the different indicators. The water measurements in freshwater influenced stratum (mean CV = 82%) is more variable than the marina stratum (mean CV = 42%); possibly due to the stations' more open locations and the differing types of freshwater influences. In contrast, the sediment constituents are more variable in the marina stratum (mean CV = 87%) than in the freshwater influenced stratum (mean CV = 56%).

Table 4-2. Comparison of sample variability in primary and secondary indicators.

Constituent		Marina			Freshwater Influenced		
		Mean	St Dev	CV (%)	Mean	St Dev	CV (%)
Water	**Dissolved Metals**						
	Copper (Cu)	8.1	3.9	48	3.6	3.3	91
	Nickel (Ni)	0.5	0.1	23	0.8	0.4	49
	Zinc (Zn)	22.8	11.1	49	9.2	10.5	114
	Total Metals						
	Copper (Cu)	10.5	5.0	47	4.2	4.3	102
	Nickel (Ni)	0.4	0.1	36	0.6	0.2	32
	Zinc (Zn)	24.7	12.7	51	10.0	10.5	105
	Mean CV Water			**42**			**82**
Sediment	Cadmium (Cd)	0.2	0.1	63	0.3	0.2	53
	Chromium (Cr)	43.7	26.7	61	36.8	10.7	29
	Copper (Cu)	449	810	180	116	74.5	64
	Lead (Pb)	30.4	33.5	110	29.1	15.4	53
	Nickel (Ni)	11.9	5.2	44	12.7	3.0	24
	Zinc (Zn)	212	104	49	186	62.9	34
	Total Detectable PAHs	612	1203	197	770	850	110
Toxicity	*E. estuarius* mortality	3.1	4.2	136	9.3	9.6	103
Infauna	Number of Taxa	30.9	10.7	35	32.1	10.8	34
	Total Count	426	586	138	1344	1764	131
	Shannon-Wiener Diversity Index	2.4	0.4	17	2.2	0.6	26
	BRI Score	38.5	7.5	19	45.0	5.4	12
	Mean CV Sediment			**87**			**56**

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